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1 ANALYSIS AND MODELLING OF SEDIMENT TRANSFER IN MEDITERRANEAN RIVER BASINS

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3 **Spectral fingerprinting: Characterising suspended sediment sources by the use of VNIR-SWIR**  
4 **spectral information**

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6 **Arlena Brosinsky • Saskia Foerster • Karl Segl • José Andrés López-Tarazón • Gemma Piqué •**  
7 **Axel Bronstert**

8

9 Arlena Brosinsky (✉) • A. Bronstert

10 University of Potsdam, Institute of Earth and Environmental Science, Karl-Liebknecht-Str. 24-25,  
11 14476 Potsdam, Germany

12 e-mail: [arlena.brosinsky@gfz-potsdam.de](mailto:arlena.brosinsky@gfz-potsdam.de)

13

14 A. Brosinsky (✉) • S. Foerster • K. Segl

15 GFZ German Research Centre for Geosciences, Section 1.4 Remote Sensing, Telegrafenberg,  
16 14473 Potsdam, Germany

17

18 J. A. López-Tarazón

19 School of Natural Sciences and Psychology, Liverpool John Moores University, Byrom St, L3 3AF, UK

20

21 J. A. López-Tarazón • G. Piqué

22 Fluvial Dynamics Research Group, Department of Environment and Soil Sciences, University of  
23 Lleida, Alcalde Rovira Roure 191, 25198, Spain

24

25 G. Piqué

26 Catalan Institute for Water Research, Emili Grahit 101, 17003 Girona, Spain

27

28

29 (✉) **Corresponding author:**

30 Arlena Brosinsky

31 Tel +49 (0) 331 288 1195

32 Fax +49 (0) 331 288 1192

33 e-mail: [arlena.brosinsky@gfz-potsdam.de](mailto:arlena.brosinsky@gfz-potsdam.de)

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35

36 **Abstract**

37 Purpose: Knowledge of sediment sources is a prerequisite for sustainable management practices and  
38 may furthermore improve our understanding of water and sediment fluxes. Investigations have shown  
39 that a number of characteristic soil properties can be used as "fingerprints" to trace back the sources  
40 of river sediments. Spectral properties have recently been successfully used as such characteristics  
41 in fingerprinting studies. Despite being less labour-intensive than geochemical analyses, for example,  
42 spectroscopy allows measurements of small amounts of sediment material (>60 mg), thus enabling  
43 inexpensive analyses even of intra-event variability. The focus of this study is on the examination of  
44 spectral properties of fluvial sediment samples to detect changes in source contributions, both  
45 between and within individual flood events.

46 Materials and methods: Sediment samples from three different origins were collected in the Isábena  
47 catchment (445 km<sup>2</sup>) in the central Spanish Pyrenees: 1) soil samples from the main potential source  
48 areas; 2) stored fine sediment from the channel bed once each season in 2011; and 3) suspended  
49 sediment samples during four flood events in autumn 2011 and spring 2012 at the catchment outlet  
50 as well as at several subcatchment outlets. All samples were dried and measured for spectral  
51 properties in the laboratory using an ASD spectroradiometer. Colour parameters and physically based  
52 features (e.g. organic carbon, iron oxide and clay content) were calculated from the spectra. Principal  
53 component analyses (PCA) were applied to all three types of samples to determine natural clustering  
54 of samples, and a mixing model was applied to determine source contributions.

55 Results and discussion: We found that fine sediment stored in the river bed seems to be mainly  
56 influenced by grain size and seasonal variability, while sampling location – and thus the effect of  
57 individual tributaries or subcatchments – seem to be of minor importance. Suspended sediment  
58 sources were found to vary between, as well as within, flood events; although badlands were always  
59 the major source. Forests and grasslands contributed little (<10%) and other sources (not further  
60 determinable) contributed up to 40%. The analyses further suggested that sediment sources differ  
61 among the subcatchments and that subcatchments comprising relatively large proportions of  
62 badlands contributed most to the four flood events analyzed.

63 Conclusions: Spectral fingerprints provide a rapid and cost-efficient alternative to conventional  
64 fingerprint properties. However, a combination of spectral and conventional fingerprint properties  
65 could potentially permit discrimination of a larger number of source types.

66

67

68 **Keywords** Isábena river • Mixing models • Northeast Spain • Sediment fingerprinting • Spectroscopy  
69 • Suspended sediment

70

## 71 1 Introduction

72 In addition to negative impacts that suspended sediments can have on water quality, as described by  
73 numerous studies (e.g. Owens et al. 2005; Walling 2005; Davis and Fox 2009; Poulenard et al. 2009),  
74 large amounts of suspended sediments also affect water quantity, as is the case in the Isábena  
75 catchment (Spain) investigated in this study. Suspended sediment concentrations at the Isábena  
76 outlet can exceed  $350 \text{ g l}^{-1}$  during flood events (López-Tarazón et al. 2009). Such high concentrations  
77 result in severe siltation problems in the Barasona reservoir, located at the catchment outlet below the  
78 confluence with the River Ésera. The reservoir was built in 1932 and enlarged in the 1970s to a  
79 capacity of  $92 \times 10^6 \text{ m}^3$  (López-Tarazón 2011). Despite sluicing in the late 1990s, the initial capacity  
80 has been reduced considerably, adversely affecting the mid-term reliability of water supply (Mamede  
81 2008).

82 Knowledge of sediment provenance is a key factor in understanding sediment transport and delivery  
83 processes and thus a first step in the design of sustainable watershed management strategies  
84 (e.g. Walling 2005; Davis and Fox 2009; Navratil et al. 2012). Such strategies must target the primary  
85 sources in order to control sediment fluxes within the watershed (Mukundan et al. 2013). In the study  
86 catchment, field observations as well as modelling have indicated badlands to be the major sediment  
87 source (e.g. Fargas et al. 1997; Francke et al. 2008a; Alatorre et al. 2010; López-Tarazón et al.  
88 2012). Badlands are defined as “areas of unconsolidated sediments or poorly consolidated bedrock,  
89 with little or no vegetation [...] in an intensely dissected landscape” (Gallart et al. 2002). Lithology is a  
90 major factor for badland development and though they are commonly considered characteristic of  
91 dryland regions they also occur in more humid climates with high topographic gradients and intense  
92 rainstorms (Gallart et al. 2002). Despite badlands being considered as major sediment sources,  
93 significant changes in the colour of the suspended sediments have been observed between, and even  
94 within, runoff events, suggesting the influence of varying sources.

95 A direct approach to trace the origin of sediment is a method called fingerprinting. It is founded on the  
96 principal assumptions that: (1) potential sediment sources can be discriminated based on a set of  
97 characteristic properties (“fingerprints”); and (2) the comparison of these source characteristics with  
98 those of (suspended) sediment allow for determination of relative source contribution (Collins and  
99 Walling 2004). In the past 30 years, the source fingerprinting approach has been successfully applied

100 as a research tool in many ecoregions around the world (e.g. Walling 2005; Davis and Fox 2009).  
101 However, the adaptation of the technique as a management tool is hampered due to several reasons.  
102 Most importantly, the choice of successful fingerprint properties is highly site-specific and the lack of  
103 general guidelines for pre-selection of parameters capable of tracing back sources can make the  
104 approach very time consuming and costly (e.g. Collins and Walling 2002). Thus, recent studies have  
105 focused on the testing of robust and inexpensive methods for the derivation of such properties  
106 (e.g. Gibbs 2008; Poulenard et al. 2009; Martínez-Carreras et al. 2010b). Consequently, spectroscopy  
107 was found to offer considerable potential for time-efficient and cost-effective measurements  
108 (Poulenard et al. 2009, 2011; Martínez-Carreras 2010a, 2010b, 2010c; Evrard et al. 2012; Legout et  
109 al. 2013; Brosinsky et al., this issue). In addition to being less labour intensive regarding laboratory  
110 analyses, spectroscopy offers the advantage of small sample size requirements (Martínez-Carreras et  
111 al. 2010a).

112 The aim of the current study is the interpretation of visible (VIS) to shortwave infrared (SWIR)  
113 spectroscopic data (0.35 – 2.5  $\mu\text{m}$ ) to examine the relative contributions from different sources and  
114 how these contributions change both between and within individual storm events. Therefore, potential  
115 source areas (1), as well as suspended sediments (2), and fine sediment stored in the channel bed  
116 (3) were collected and spectrally measured. Colour parameters and spectral features with relation to  
117 organic carbon, iron oxide and clay content were calculated from spectra and subsequently tested to  
118 meet a number of assumptions. Parameters meeting the assumptions were used in principal  
119 component analyses (PCA) for all three sample types (1-3) to determine natural clustering, and a  
120 mixing model was applied to suspended sediment samples to determine source contributions.

121

## 122 2 Study area

123 The Isábena River drains a 445 km<sup>2</sup> basin in the southern Pyrenees (Ebro catchment, NE Spain, just  
124 before entering the Barasona reservoir together with the Ésera River (Fig. 1). The Isábena catchment  
125 comprises five main subbasins: the Cabecera subbasin in the North (146 km<sup>2</sup>, representing 33 % of  
126 the total catchment area), the Villacarli (42 km<sup>2</sup>, 9 %) and Carrasquero (25 km<sup>2</sup>, 6%) subbasins in the  
127 NW and Ceguera (28 km<sup>2</sup>, 6 %) and the Lascuarre subbasin (45 km<sup>2</sup>, 10 %) in the SE (Fig. 1). The  
128 remaining area drains directly into the Isábena River. Overall, the area is characterized by a rough

129 terrain (450 m a.s.l. in the southern lowlands to 2700 m a.s.l. in the headwaters), resulting in a  
130 pronounced climatic and land cover gradient. The climate is of Mediterranean mountainous type with  
131 mean annual temperatures ranging between 14 °C (south) to 9 °C (north) and a mean annual  
132 precipitation of 450 mm (south) to 1600 mm (north) (Verdú 2003). Precipitation is of high spatial and  
133 temporal variability with maxima generally occurring in spring and autumn (López-Tarazón et al.  
134 2010). Despite occasional gravel mining, the Isábena is an entirely unregulated river with a pluvial–  
135 nival runoff regime. Valero-Garcés et al. (1999) found that the major floods in the Ésera–Isábena  
136 basins are caused by late spring–early summer snow melt in combination with heavy rains, summer  
137 thunderstorms, and late autumn heavy rains. Frequent floods keep sediment transport rates at  
138 relatively high levels, with instantaneous suspended sediment concentrations occasionally attaining  
139 350 g l<sup>-1</sup> (López-Tarazón et al. 2009).

140 While valley bottoms are mainly used for agriculture, higher altitudes are dominated by shrubland  
141 (matorral), grassland, woodland (*Quercus* and *Pinus*) and bare soil and rock. Major land use changes  
142 that occurred over the past 50 years resulted in the abandonment of cultivated areas and subsequent  
143 revegetation, initiated by shrub species and followed by forest regrowth (e.g. Lasanta and Vicente-  
144 Serrano 2012).

145 The catchment is characterized by a heterogeneous lithology. Valero-Garcés et al. (1999) describe  
146 several WNW–ESE trending geologic units, i.e.: (1) the Axial Pyrenees composed of Paleozoic rocks  
147 (quartzites, limestone) with peaks above 3000 m a.s.l. in the north; (2) the Internal Ranges composed  
148 of Cretaceous and Paleocene sediments in the center; and (3) the Intermediate Depression, a  
149 relatively lowland area in the south of the catchment that is composed of Miocene continental  
150 sediment. The most important soil types developed in this area can be classified as shallow mineral  
151 soils (including regosols, leptosols and fluvisols) and soils with a considerable accumulation of  
152 organic matter, including kastanosems (Alatorre et al. 2010). There is high variability in soil colour that  
153 is most obvious in the agricultural fields, ranging from reddish brown to grayish and dark brown.

154 Valero-Garcés et al. (1999) describe several internal depressions formed upon highly erodible  
155 materials (marls, sandstones, and carbonates) that are located in the central part of the watershed.  
156 These areas with relatively high topographic gradients and moderate vegetation cover lead to the  
157 development of badlands. Despite representing <1 % of the total basin area (mainly in the Villacarli

158 and Carrasquero subbasins), badlands are considered the most important sediment source in the  
159 catchment (e.g. Fargas et al. 1997; Francke et al. 2008a; Alatorre and Beguería 2009; López-Tarazón  
160 et al. 2012) with erosion rates estimated to exceed  $550 \text{ t ha}^{-1} \text{ yr}^{-1}$  (Appel 2006). Although the  
161 topographic gradient is higher in the northern areas, lithology as well as higher vegetation cover  
162 (grassland, forest) reduce sediment production. In the southern areas, cultivated land predominates.  
163 While Valero-Garcés et al. (1999) suggest that the smaller topographic gradient seems to be reducing  
164 erosion, Alatorre et al. (2010) identify dryland crop areas as important contributors of suspended  
165 sediment yield.

166 The catchment has been subject to intensive hydrological studies over the past decade  
167 (e.g. Bronstert et al. this issue), resulting in a detailed understanding of hydrological and  
168 geomorphological processes and a favorable instrumentation situation, thus making it an ideal test  
169 site for innovative techniques.

170

### 171 **3 Material and methods**

#### 172 3.1 Sampling and data overview

173 Field campaigns were conducted to collect samples from three different origins: 1) 152 soil samples  
174 were collected from the main potential source areas within the Isábena catchment; 2) a total of  
175 48 samples of fine sediment stored in the main channel bed were collected at 14 cross sections once  
176 each season in 2011; and 3) suspended sediment samples from the river were collected during four  
177 flood events in autumn 2011 and spring 2012 at the catchment outlet (44 samples) as well as at  
178 several sub-catchment outlets (46 samples selected for analyses). The spatial location of sampling  
179 sites is shown in Fig.1, the temporal allocation of sediment samples (channel bed and suspended)  
180 with respect to discharge at the Capella gauge can be seen in Online Supplementary Material 1.

181

##### 182 *3.1.1 Source material sampling*

183 Source material samples were collected during two campaigns in October 2010 and June 2011.  
184 Sampling sites were selected based on previous analyses of land use distribution (Ministerio de  
185 Medio Ambiente y Medio Rural y Marino (MARM) 2008), knowledge on erosion susceptibility (Fargas  
186 et al. 1997) and accessibility. All major land uses that potentially contribute to erosion in the Isábena

187 catchment were sampled, namely agricultural land (covering 14 % of the catchment), forest (47 %),  
188 grassland (8 %) and shrubland (30 %) (Table 1). In addition, potential sources covering only small  
189 areas (< 1 %) but possibly contributing a high proportion of material were sampled, including  
190 badlands, unpaved roads and open slopes. Thereby, the distribution of sampling locations were  
191 considered to be spatially representative of all subcatchments. The number and details of samples  
192 collected are listed in Table 2.

193 Since soils in the study area are shallow with poorly developed diagnostic horizons (López-Tarazón  
194 2011), and lithology is not very distinct and overall rather homogeneous in large parts of the  
195 catchment, emphasis was placed on a land use-based sampling strategy. However, care was taken to  
196 ensure that the distribution of sampling locations were spatially representative over all lithological  
197 units (Ministerio de Agricultura, Alimentación y Medio Ambiente). Sampling sites were chosen in close  
198 vicinity (< 100 m) to stream or river reaches to make sure the material will be easily transported to the  
199 river. From each site, five grab samples of easily erodible material (top 1-3 cm) were collected from a  
200 5 m x 5 m area and well mixed. The location of sampling sites is shown in Fig. 1.

201

### 202 *3.1.2 River sediment sampling*

203 Suspended sediment samples were collected hourly during flood events by an ISCO automatic  
204 sampler (ISCO 3700, Teledyne, Lincoln, Nebraska, USA) at the Capella gauging station near the  
205 catchment outlet. The sampler was triggered by flow conditions (i.e., start sampling from a certain  
206 water level). Four of the events sampled were chosen to be analyzed and discussed in more detail in  
207 the present work, namely from (i) 24<sup>th</sup>/25<sup>th</sup> September 2011 (event A), (ii) 22<sup>nd</sup> March 2012 (event B),  
208 (iii) 3<sup>rd</sup>/4<sup>th</sup> June 2012 (event C) and (iv) 20<sup>th</sup> June 2012 (event D) (Table 3).

209 In addition, suspended sediment samples were collected near the outlets of the five main  
210 subcatchments described above, by means of ISCO samplers (ISCO 3700C, Teledyne). Again, the  
211 samplers were triggered by flow conditions and samples were collected in 15 minute to hourly  
212 intervals depending on the runoff behavior prevailing in the individual subcatchment (Table 3).

213 Furthermore, fine sediment stored within the bed of the Isábena main channel was sampled using the  
214 methodology developed by Lambert and Walling (1988). At 14 different cross-sections along the main  
215 channel, a metal cylinder of 50 cm in diameter and 60 cm height was carefully placed on the channel



216 bed and slowly rotated to create a seal with the underlying gravel. The sampling area thus created  
217 was manually disturbed using a rod, resulting in re-suspension of stored fine sediment (for details see  
218 López-Tarazón et al. 2011). This procedure was repeated four times, once per season (namely,  
219 25<sup>th</sup> February, 24<sup>th</sup> May, 26<sup>th</sup> July, and 1<sup>st</sup> December) in 2011; sampling was performed at exactly the  
220 same locations during all campaigns. The number and details of samples analyzed are listed in Table  
221 4, and sampling locations are shown in Fig 1.

222 Sediment concentration of all samples was determined by settling of known volumes of higher  
223 concentration samples (approx.  $> 2 \text{ g l}^{-1}$ ) and filtering of lower concentration samples using 1.2  $\mu\text{m}$   
224 FILTER-LAB glass microfiber filters. Loose material was dried at 60 °C for  $> 24$  hours or air dried for  
225 over one week and weighed; filters were weighed prior to material application, then dried at 60 °C for  
226 two to three hours or air dried for  $> 24$  hours and reweighed.

227

### 228 *3.1.3 Water discharge and rainfall measurements*

229 Water stage was recorded in 15 minute intervals at the Capella gauging station by the Ebro Water  
230 Authorities (CHE) and later transformed into discharge using the calibrated stage–discharge rating  
231 curve developed by López-Tarazón et al. (2010). Rainfall was measured by tipping-bucket gauges  
232 operated by the University of Potsdam / GFZ Potsdam. There are one or two rain gauges installed per  
233 subcatchment, resulting in a total number of eight rain gauges representing rainfall distribution over  
234 the catchment area. Sampling locations are shown in Fig. 1.

235

### 236 *3.2 Spectral measurements of source and sediment samples*

237 Spectral reflectance data were collected using an Analytical Spectral Device (ASD) FieldSpec3 High-  
238 Res portable spectroradiometer (Analytical Spectral Device, INC., Boulder, Colorado, USA), acquiring  
239 2151 bands in the 0.35 – 2.5  $\mu\text{m}$  range of the electromagnetic spectrum at a true sampling interval of  
240 1.4 nm in the VIS-NIR region (0.35 – 1.0  $\mu\text{m}$ ) and 2 nm in the SWIR region (1.0 – 2.5  $\mu\text{m}$ ). Relative  
241 reflectance was calculated automatically by using a white reference panel as standard (100 %).

242 Loose material was thoroughly mixed to provide homogeneous samples. Since suspended sediment  
243 samples were mainly  $< 63 \mu\text{m}$ , source material was sieved to 63  $\mu\text{m}$  to minimize spectral variations  
244 resulting from differences in particle size composition between source and suspended sediment

245 material (e.g. Walling 2005). Source and suspended sediment material was placed in shallow  
246 5 cm x 5 cm plastic containers and oven dried at 60 °C for 24 hours prior to spectral measurements.  
247 Spectral readings were taken in a dark room facility (for details see Brosinsky et al. (this issue)). Four  
248 readings per sample were taken, with the sample rotated 90° after each reading to reduce illumination  
249 effects.

250 We detected spectral differences between loose material and material on filters that can partly be  
251 attributed to a loss in fine material (filter pore size 1.2 µm) and partly to alignment of sediment  
252 particles resulting in changes in reflectance behaviour. Thus, loose material and material retained on  
253 filters can both be used but the measurements should not be compared directly. While all soil and  
254 suspended sediment analyses were based on loose samples, there was not always enough material  
255 from resuspension samples. Thus, all resuspension samples were applied to glassfiber filters as  
256 described above for spectral measurements. In addition, some of the resuspension samples were  
257 much coarser than suspended samples collected by ISCOs; the resuspension samples were not  
258 sieved prior to spectral measurements.

259

### 260 3.3 Preprocessing and parameter calculation

261 Spectral readings per sample were averaged and smoothed using a Savitzky-Golay filter (Savitzky  
262 and Golay 1964). Then, red, green and blue (RGB) colour parameters were calculated from spectra  
263 by averaging values of spectral reflectance ranges corresponding to the blue, green and red Landsat  
264 bands (0.45 – 0.52 µm, 0.52 – 0.6 µm, and 0.63 – 0.69 µm, respectively) and multiplication with 255  
265 to get 8-bit colour encoding (Viscarra Rossel et al. 2006). These RGB values were transformed to  
266 other colour space models using ColoSol software developed by Viscarra Rossel et al. (2006). All  
267 transformation algorithms are described in detail by Viscarra Rossel et al. (2006) and details on colour  
268 models are explained by Wyszecki and Stiles (1982).

269 Following the description by Bayer et al. (2012), features found in the previous literature to be  
270 diagnostic of physically based information on soil organic carbon, clay, iron and carbonate content  
271 were calculated. The selected spectral parameters can be divided into spectral indices and three  
272 feature types, namely curve features, hull features, and absorption features. Details on the calculation

273 of parameters can be found in Chabrilat et al. (2011), Bayer et al. (2012) and Brosinsky et al. (this  
274 issue).

275 In total, a set of 98 colour and physically based soil reflectance parameters was calculated (see  
276 Brosinsky et al (this issue)). Since colour coefficients may be easily converted and all spectral  
277 features are potentially used in spectroscopy and soil science, they were all considered in subsequent  
278 analyses, although some of these parameters may be highly correlated (Viscarra Rossel et al. 2006;  
279 Martínez-Carreras et al. 2010c). In the following, the term “spectral parameter(s)” is used as a  
280 synonym for spectral fingerprint property, describing colour parameters and/or physically based  
281 reflectance features calculated as outlined above.

282  
283  
284 3.4 Test of assumptions

285 A number of fundamental assumptions of the fingerprinting procedure were tested in an attempt to  
286 limit uncertainty of sediment provenance assessment to a minimum. In recent studies, the potential  
287 non-conservativeness of tracer properties has been identified as a major concern, with key issues  
288 being particle size selective transport and tracer transformation (e.g. Koiter et al. 2013). For example,  
289 Smith and Blake (2014) found the relationships between these processes to be highly complex and  
290 discourage the use of correction factors. Thus, the problem of size selective transport was addressed  
291 by sieving all soil and suspended sediment material to  $< 63 \mu\text{m}$  (e.g. Martínez-Carreras 2010a;  
292 Mukundan et al. 2013; Smith and Blake 2014). Grain size analyses of a selection of sieved samples  
293 pointed to no enrichment or depletion effects. Secondly, although tracer transformation cannot be  
294 entirely excluded, it was addressed by limiting the analyses to spectral parameters whose values  
295 calculated from suspended sediment lie wholly in the range of those calculated from potential source  
296 samples (92 out of 98) (Smith and Blake 2014). The high number of parameters meeting this  
297 prerequisite indicates that any alteration effects may have been relatively small (Walden et al. 1997).  
298 Though spectroscopic measurements are sensitive to alterations during transport, such as reduction  
299 of iron or decomposition of organic matter, Legout et al. (2013) found changes in VIS spectra and  
300 colour parameters to remain  $< 10\%$  when comparing original samples to samples immersed in a river  
301 for a maximum period of 63 days. Linear additivity of spectral properties was explicitly tested by  
302 comparing properties calculated from artificial mixture spectra to properties calculated from mixture

303 spectra produced by a linear mixing algorithm (Brosinsky et al. this issue). Only 48 out of 92  
304 parameters met this assumption and were used in subsequent procedures. Following Walling (2005),  
305 all remaining parameter values were scaled between 0 and 1 to ensure equal consideration of  
306 individual properties. A non-parametric Kruskal-Wallis H-test used to assess the existence of any  
307 significant interclass contrasts (Collins and Walling 2002) revealed that all 48 parameters were able to  
308 detect contrasts between the seven source types at the 95% confidence level.

309

### 310 3.5 Statistical analyses to assess natural clustering of samples

311 Principal component analyses (PCA) were performed on the 48 parameter source and sediment  
312 datasets to determine natural clustering of samples and to evaluate overall variability and potential  
313 overlap between classes (Poulenard et al. 2009). The PCA was applied on source and suspended  
314 sediment samples (Capella) together, providing an indication of how successful subsequent  
315 quantitative mixing modelling is likely to be (Walden et al. 1997). In addition, PCA was applied on  
316 suspended sediment from river samples only (Capella and subcatchments) in order to get an  
317 impression of sample clustering between and within individual runoff events.

318 Furthermore, a PCA was performed on the resuspended material data only, since, unlike all other  
319 samples, the resuspended material was retained on glassfiber filters. Therefore these data could not  
320 be directly compared to the suspended sediment or source samples which were not filtered. Thus,  
321 this analysis will provide only qualitative results on general changes or resemblances within the stored  
322 sediment and will not allow us to draw conclusions on source contributions or similarities with the  
323 suspended load collected during individual flood events. All PCA analyses were performed using The  
324 Unscrambler® X 10.2 software (CAMO Software AS., Oslo, Norway).

325

### 326 3.6 Mixing model analyses

327 Since previous PCA and discriminant function analyses (DFA) results calculated from source samples  
328 suggested confusion between forest and grassland samples, as well as between shrubland and  
329 arable land, road and open slopes (Brosinsky et al., this issue), the seven source types were  
330 aggregated into three source types for input to the mixing model; namely: badland; forest/grassland;

331 and others. Only 45 parameters passed the Kruskal-Wallis H-test for the three groups and were used  
332 for subsequent unmixing analyses.

333 Relative contributions of potential sources were estimated by comparing the fingerprint properties of  
334 the sediment samples with those of the potential sources using a multivariate mixing model; a detailed  
335 description of the model can be found in Brosinsky et al. (this issue). Errors between measured and  
336 estimated values were approximated using the non-negative least squares algorithm introduced by  
337 Lawson and Hanson (1974), where the best approximation is defined as the one minimizing the sum  
338 of squared differences between the measured data values and their corresponding modelled values.  
339 The model was restricted by the constraints that the source type contributions must all be non-  
340 negative and sum to 100%. Uncertainty associated with modelling results due to source heterogeneity  
341 was assessed by producing Gaussian distribution functions from the mean and standard deviation of  
342 each tracer property per source type (Martínez-Carreras et al. 2010a). The mixing model was run  
343 10,000 times, choosing source information randomly from the Gaussian distribution functions, thus  
344 allowing source describing properties to vary in each solution. This replicate random sampling  
345 provides confidence estimates for the modelled contribution results by permitting the calculation of  
346 percentiles. The model was implemented using in-house software (ANSI-C).

347

## 348 **4 Results**

### 349 4.1 Principal component analyses – natural clustering of samples

#### 350 *4.1.1 Resuspension samples*

351 Figure 2 shows PCA score plots of all samples of resuspended channel bed sediment (first three  
352 components). Generally, the heterogeneity of the samples reflected in the first three components is  
353 rather high. The first three components together explain 88 % of the variance (44 %, 35 % and 9 %,  
354 respectively), and seven components explain 98 % of the total variance.

355 A scatterplot of the first two components (Fig. 2a), in general, seems mainly influenced by grain size.  
356 Though grain size distribution was not analyzed in particular, visual inspection revealed that some  
357 filters contained very fine material while other filters contained coarser material and/or sand grains.  
358 The majority of finer samples cluster in the upper right corner of the scatter plot while the coarser  
359 samples are mainly distributed to the lower left. Samples from February (I) and July (III) seem

360 generally finer than samples from May (II) and December (IV) but this is inconsistent and not true for  
361 all sections sampled (Fig. 2b).

362 Fig. 2c again depicts the first two components, where the samples are identified by the 14 sections  
363 they were collected from. Both samples S1 and two of the three S2 samples cluster well away from  
364 the other samples and rather close together. Then, a rough zonation trend can be observed for the  
365 upstream sections with samples from S3 situated to the right of the plot and samples from S4, S5,  
366 and S6 distributed further to the left. Samples from S7 and S8 are again distributed further to the right.  
367 With the exception of S13, samples S9 to S14 are very heterogeneous. The observed “zonation  
368 trend” is not consistent with either season or grain size variation.

369 Figure 2d shows the second and third PC, that mainly seem to reflect seasonal variation. Thus,  
370 samples collected in May (II) cluster between the samples from February (I) and December (IV) while  
371 samples collected in July (III) are separate. As in the first two component plots, samples from S1 and  
372 S2 are somewhat separate. Apart from that, the sampling section does not seem to have a detectable  
373 influence on any of the first seven components.

374

#### 375 *4.1.2 Suspended sediment samples*

376 Figure 3 shows several PCA score plots of source and suspended sediment samples (first two  
377 components). Overlap between samples from different land use classes is evident from Fig. 3a,  
378 whereas badland and forest/meadow samples form somewhat separate clusters and all other land  
379 uses seem not to differentiate. The Capella suspended sediment samples plot within the catchment  
380 source materials, between badland and road, agricultural and shrubland samples. Compared to  
381 source material samples, sediment material is very homogeneous. However, a distinction between  
382 individual events is clearly evident.

383 No grouping is visible with respect to the source samples' subcatchment of origin (see Fig. 3b), with  
384 the exception of the northernmost subcatchment (Cabecera), which is underlain by different  
385 substrates than the more southern subcatchments. However, no grouping by lithology is evident from  
386 the PCA scatter plots (results not shown).

387 Although Fig. 3b does not reveal clustering of source samples by subcatchment, Fig. 3c shows a  
388 clear distinction of all suspended sediment samples by subcatchment (Capella and subcatchment

389 samples). A distinct clustering of most samples into the four subcatchments of their origin is evident.  
390 Due to a lack of sampling material, the northernmost subcatchment (Cabecera) is missing from this  
391 analysis. Samples collected near the basin outlet (Capella) plot completely in the center, with a shift  
392 towards the Villacarli and Carrasquero subcatchments.

393 Figure 3d shows the first and second PC of a PCA performed on Capella sediment samples only.  
394 Despite being generally much more homogeneous than source samples, the distinction between  
395 individual events evident from Fig. 3a is even more pronounced when looking at the sediment  
396 samples separately. While events B and C plot closely together, most samples from event D are  
397 clearly different. Intra-event variability is mainly visible from PC-2. Event A contains most samples and  
398 is most heterogeneous, with large variability along both first PCs, with some samples resembling  
399 events B, C and D and other samples being obviously of different composition. Therefore, it is the  
400 early (samples A1-A4) and later (samples A15-A22) stages of the event that more closely resemble  
401 the other events sampled at Capella while, the middle (samples A7-A14) and end (samples A23 and  
402 A24) stages seem most different.

403

#### 404 4.2 Application of mixing model to suspended sediment samples

405 Source tracing results of four events sampled at Capella from September 2011 to June 2012 including  
406 information on runoff and suspended sediment concentration (SSC) are shown in Fig. 4 to Fig. 7.  
407 Details on precipitation (sum, average, intensity, duration) and a rough characterization of distribution  
408 over the subcatchments can be found in Table 5. Figure 4 provides evidence of the high intra-event  
409 variability from 24<sup>th</sup> to 25<sup>th</sup> September 2011. Runoff shows a pronounced peak with steep rising and  
410 falling limbs and a much smaller second peak about 12 hours after the first peak. The peak runoff was  
411  $33 \text{ m}^3 \text{ s}^{-1}$  and maximum SSC was  $35 \text{ g l}^{-1}$  at the peak and decreased quickly. Precipitation during the  
412 event was of short duration (< 3 hours) and high intensity (up to  $40 \text{ mm h}^{-1}$ ) while the average amount  
413 of rainfall was moderate (40 mm). Overall, the samples comprise high proportions of badland material  
414 (mean 58 to 80 %). During the fallings limbs, other sources became more dominant (on average up to  
415 mean 42 %) while the contribution of forest/grassland was generally low (mean 0 to 11 %).

416 For the three events sampled in 2012, sample availability as well as intra-event variability, were lower  
417 (Fig. 5-7). The event from 22<sup>nd</sup> March 2012 is characterized by a much broader runoff peak with a

418 steep rising and a shallow falling limb (Fig. 5). Peak runoff was  $28 \text{ m}^3 \text{ s}^{-1}$  and maximum SSC was  
419  $19 \text{ g l}^{-1}$  at the peak and decreased quickly. Precipitation during the event was of longer duration  
420 ( $> 15$  hours) and of lower intensity ( $< 20 \text{ mm h}^{-1}$ ) than the September event while the average amount  
421 of rainfall was higher (70 mm). The contribution of badland material to suspended sediment was  
422 estimated to exceed 80 % for all but the first sample and forest/grassland contribution was low ( $< 4$  %)  
423 for all samples.

424 The event from 3<sup>rd</sup> to 4<sup>th</sup> June, 2012 is again characterized by a steep runoff peak (Fig. 6) yielding a  
425 maximum of only  $19 \text{ m}^3 \text{ s}^{-1}$ . However, SSC was high, with a maximum of  $68 \text{ g l}^{-1}$  during the peak.  
426 Precipitation during the event was again of short duration ( $< 3$  hours) and partly of very high intensity  
427 ( $10\text{-}40 \text{ mm h}^{-1}$ , in the Villacarli subcatchment up to  $105 \text{ mm h}^{-1}$ ) while the average amount of rainfall  
428 was low (25 mm). Mean estimated contribution of badland material to suspended sediment varied  
429 between 78 and 85 %, and forest/grassland contribution was again low ( $< 5$  %) for all samples.

430 As for the event sampled in March 2012, the event from 20<sup>th</sup> June 2012 was characterized by a broad  
431 runoff peak with shallow rising and falling limbs (Fig. 7). Runoff and SSC were low yielding maxima of  
432  $15 \text{ m}^3 \text{ s}^{-1}$  and  $5.5 \text{ g l}^{-1}$ , respectively. Precipitation during the event was characterized by several short  
433 low intensity – low amount events followed by a longer ( $> 12$  hours) event of low intensity  
434 ( $< 20 \text{ mm h}^{-1}$ ), while the average amount of rainfall was moderate (40 mm). The mean estimated  
435 contribution of badland material to suspended sediment varies between 68 and 81 %, and the  
436 contribution increased with the rising and decreased with the falling limb. The forest/grassland  
437 contribution was estimated to below ( $< 5$  %) for all but the last sample, where this source was  
438 estimated to account for 20 %. Other sources were estimated to contribute 12 to 26 %.

439

## 440 **5 Discussion**

441 5.1 Principal components analyses – natural clustering of samples

### 442 5.1.1 *Resuspension samples*

443 The PCA performed on resuspended sediment samples indicates that spectral reflectance of the  
444 samples was mainly influenced by variations in grain size as well as seasonal variations. Since  
445 seasonal variations are not completely identical with grain size variations there must be other/further  
446 influencing factors such as source contribution variations or variations in storage and transition



447 behaviour of the River Isábena (Piqué et al. this issue). Samples collected at sections S1 and S2,  
448 upstream of the Cabecera outlet, were found to differ from most other samples. Apart from that, a lot  
449 of heterogeneity was found in the samples within, as well as between, sampling sections, indicating  
450 that if there is a pattern at all it is superimposed on other factors. This is consistent with results of  
451 López-Tarazón et al. (2011) and Piqué et al. (this issue) who analyzed the amount of stored sediment  
452 in the lower Isábena reaches (S11 – S14) for the periods 2007-2008 and 2011-2012, respectively.  
453 They found considerable variations in sediment storage along the main channel, identifying an annual  
454 cycle of sediment production and transfer downstream. However, it is difficult to capture this pattern  
455 since sediment accumulation is not linear in time and space, but is largely influenced by spatial and  
456 temporal (i.e. seasonal) heterogeneity in the catchment's hydrology and sediment production in the  
457 badlands.

458

#### 459 *5.1.2 Suspended sediment samples*

460 Results of the source and suspended sediment PCAs can be interpreted as indicators of the feasibility  
461 of the spectral fingerprinting approach (Walden et al. 1997). The fact that all suspended sediment  
462 samples plot within the margins of potential source samples may be seen as an indicator that all  
463 major sources have been sampled and that changes in suspended sediments which may have  
464 occurred during transport and storage have been relatively small (Walden et al. 1997).

465 A distinct pattern of source samples by land use was found, however, no clustering was evident by  
466 lithological units, indicating that (in this catchment) land use practices may supercede soil type or  
467 lithological differences. The plotting of suspended sediment samples close to badland and unpaved  
468 road, agricultural and shrubland source samples, and further away from forest and grassland source  
469 samples, confirm the mixing model results, which suggest: a high contribution of badland materials in  
470 all samples; a medium to high contribution of other materials; and a low contribution from  
471 forest/grassland materials. Compared to source material samples, the heterogeneity of sediment  
472 material is very limited. However, a distinction between individual events is evident, confirming  
473 changes in source contributions between the events as suggested by the mixing model. Samples  
474 estimated to contain a higher proportion of badland materials (events B and C, some samples of  
475 events A and D) plot closer to badland source samples, while samples containing higher shares of

476 other materials (event A) plot closer to unpaved road, open slope, agricultural and shrubland source  
477 samples.

478 There is no clustering of source samples by subcatchment but there is a clear distinction in the  
479 suspended sediment samples from the subcatchments by origin, indicating major differences of  
480 source contributions in individual subcatchments. Unfortunately, the number of source samples per  
481 subcatchment is limited, constraining a subcatchment-based fingerprinting approach. Samples  
482 collected near the basin outlet (Capella) plot in the center of the subcatchment sediment samples and  
483 thus seem to represent mixtures of the subcatchment's materials, with higher contributions from the  
484 Villacarli and Carrasquero subcatchments. This is consistent with the results of other studies  
485 undertaken in the area, which identified Villacarli as a major contributor for the same (Francke et al.  
486 this issue) and other study periods (Francke et al. 2008a, 2008b; López-Tarazón et al. 2012).  
487 However, López-Tarazón et al. (2012) identified Lascuarre as the second most important sediment  
488 source, indicating that there may be major changes in spatial source contribution between different  
489 years/study periods. In addition, analyses of resuspension samples indicate that tributaries not  
490 attributed to any subcatchment (i.e. S6, S7, S9, S11-S14) may cause substantial heterogeneity and  
491 thus contribute to sediment collected at the catchment outlet.

492

## 493 5.2 Mixing model

494 The mixing model results are consistent with the PCA results in terms of source ascription, suggesting  
495 that contribution estimates are reliable (Walden et al. 1997). Previous experiments on artificial  
496 laboratory mixtures where source contributions were known revealed high levels of uncertainty  
497 (Brosinsky et al. this issue), thus suggesting reliance on the general ascription of source contributions  
498 but not on exact values. However, when comparing the suspended sediment mixing model results to  
499 artificial laboratory unmixing results it is evident that unmixing results obtained in the present study fall  
500 within the ranges that seemed reliable in the previous, controlled study (i.e. Brosinsky et al. this issue)  
501 using artificial mixtures (> 60 % badland, < 40 % others, < 20 % forest/grassland).

502 Overall, although badlands were found to be the major contributing sources to all samples analyzed,  
503 there are differences between, as well as within, events. Rainfall is distributed irregularly over the  
504 catchment and runoff response was found to be highly variable (e.g. López-Tarazón et al. 2010),

505 resulting in differences in the occurrence of floods at the subcatchment level and in the production of  
506 secondary SSC peaks (López-Tarazón et al. 2012). Differences in precipitation (short duration/high  
507 intensity around 24<sup>th</sup>/25<sup>th</sup> September and 3<sup>rd</sup>/4<sup>th</sup> June vs. long duration/low intensity around  
508 22<sup>nd</sup> March and 20<sup>th</sup> June) caused differences in runoff behavior. Whereas the runoff peak was very  
509 pronounced with steep limbs in September and early June, it was a lot broader in March and late  
510 June. The SSC was found to be generally lower in the two broad peak events, whereas it was found  
511 to be extremely low in the 20<sup>th</sup> June event. This is likely due to material depletion after the event from  
512 3<sup>rd</sup>/4<sup>th</sup> June, where SSC exceeded 60 g l<sup>-1</sup>. Material availability may also be the explanation for the  
513 differences between events B and D which, due to similar rainfall and runoff characteristics, were  
514 expected to be of more similar composition. However, mixing model as well as PCA analyses suggest  
515 a greater similarity between events B and C, whereas samples from event D resemble first peak/early  
516 falling limb and the later samples of event A. No concurrent explanation was found on the changes  
517 within the very long event A. Samples from the late stages of the first falling limb of event A are  
518 unique in their low badland/high other sources contribution as compared to the other events sampled,  
519 which again may be due to the long duration of the event and thus sediment exhaustion. Samples  
520 taken during the second, smaller peak largely resemble events B and C. Together with the  
521 occurrence of the second runoff and SSC peak, this indicates that a local rainburst may have led to  
522 the delayed contribution of a source, tributary or subcatchment.

523 This leads to the limitations of the fingerprinting technique, which is capable only of providing  
524 information on the ultimate sediment source and not the proximal one; sediment stored from previous  
525 events will be identical to sediment arriving at the sampling location at the same time from the  
526 ultimate source (Parsons 2012). Furthermore, mixing models do not consider the travel time of  
527 sediment from source to sampling point. Thus, differences in travel time arising, for example, from  
528 differences in particle size or source distance in larger catchments, can invalidate the analysis of  
529 spatial origin (Parsons 2012) and impede interpretation of changes in source type contribution.

530 However, the overall suggestion of badland sources being the main contributors meets the general  
531 expectation for this catchment (e.g. Valéro-Garcés et al. 1999; Alatorre et al. 2010). There are no  
532 analyses specifying the proportion of badland material in suspended sediment. Nevertheless,  
533 previous studies found badland erosion rates to be very high with specific sediment yields exceeding

534 6200 t km<sup>-2</sup> for a three month study period (Francke et al. 2008a). In addition, Francke et al. (2008a,  
535 2008b) calculated suspended sediment yield of the Villacarli subcatchment – the subcatchment with  
536 highest portion in badlands (6 % of the area) – to account for about 45 % of the yield measured at the  
537 catchment outlet from September to December 2006. López-Tarazón et al. (2012) estimated  
538 Villacarlis' suspended sediment contributions to vary between 61 % and 27 % for the study periods  
539 2007-2008 and 2008-2009, respectively. Forest and grassland sources contributed little to the  
540 samples analyzed in the current study, which was also expected since the soil is predominantly well  
541 protected by high vegetation cover. Unfortunately, the aggregated “other” sources could not be  
542 analyzed in more detail. Mukundan et al. (2012) discuss a catchment size of < 250 km<sup>2</sup> as the  
543 maximum scale at which fingerprinting is likely to be meaningful. Collins et al. (1998) state < 500 km<sup>2</sup>  
544 appropriate for source type fingerprinting, while increasing source heterogeneity in larger catchments  
545 could make source type fingerprinting less successful. A combination of spectral with classic  
546 fingerprinting properties such as geochemistry, mineral magnetism or radionuclides, as proposed by  
547 Martínez-Carreras et al. (2010a), could potentially provide a deeper understanding of the contribution  
548 of sources classified as “others” in this study.

549 Though differences between source contributions to different events were detected, no obvious  
550 seasonal variation was found. Instead, sediment availability seems to play a major role in the Isábena  
551 catchment. This is consistent with findings of López-Tarazón et al. (2010), who found that while there  
552 was no correlation between rainfall intensity and SSC, sediment availability in badlands and sediment  
553 storage in the channels influence the river's sedimentary response. López-Tarazón et al. (2012) found  
554 sediment delivery ratios of 90 %, indicating that large parts of the sediment mobilized in the  
555 catchment is easily transported to the outlet. However, sediment storage values in the Isábena main  
556 channel were estimated by López-Tarazón et al. (2011, 2012) to range from an average of 5 % of the  
557 annual total load to up to 55 % during certain periods. Thus, they conclude that the fine-grained  
558 sediment stored in the channel can represent an important factor in controlling the suspended  
559 sediment dynamics.

560

## 561 **6 Conclusions**

562 The focus of this study was on the examination of spectral parameters for changes in sediment  
563 storage over the seasons and in suspended sediment, both within and between flood events. Results  
564 suggest that variability in stored fine sediment is most likely due to grain size and seasonal variation.  
565 Apart from the two uppermost sampling sections, no clear trends by sampling location were observed.  
566 Overall, the influence of inter-storm and/or seasonal variation on storage and transition behaviour of  
567 the Isábena seems to be much greater than the influence of sampling location and thus, for example,  
568 the influence of individual tributaries or subcatchments.

569 However, regarding suspended sediment, considerable variability was detected between  
570 subcatchments as well as in source type contribution, both between and within individual runoff  
571 events. Badlands, with a total aerial fraction cover of <1%, were found to be the major contributing  
572 sources with values of 60 – 80 %. Other sources, covering 45% of the study catchment, contributed  
573 up to ~40 % and forest/grasslands usually contributed < 10 % despite covering 54% of the study area.  
574 This is consistent with expectations based on field observations and previous studies. Unfortunately, it  
575 was not feasible to trace the aggregated “other sources” in more detail by this spectral approach. The  
576 PCA further suggest that the Villacarli and Carrasquero subcatchments contribute most material to  
577 the flood events sampled, and that suspended sediment sources are very different in the  
578 subwatersheds.

579 The results of this study point to badlands as the major sources of suspended sediment, thus  
580 management actions should focus on controlling sediment production from these areas. While  
581 García-Ruiz et al. (2013) think that more studies are needed to understand the evolution and  
582 functioning of badland ecosystems and that little can be done to prevent sediment export, Lee et al.  
583 (2013) demonstrate successful application of erosion control measures on previously bare badland  
584 structures developed on steep mudstone slopes in Taiwan. On studying Italian badlands that have  
585 become overgrown in recent years, Bierbaß et al. (2014) found that vegetation plays a key role in  
586 altering soil properties, resulting in more stable slope conditions. However, considering that reservoir  
587 siltation rather than loss of surface soil is the main problem in the study catchment, trapping material  
588 in or near the outlet of badlands, at the place of production, might be a more economically feasible  
589 option. Morgan (2005) describes measures such as siltation fences and artificial sedimentation ponds  
590 – capturing suspended sediment and allowing clearer runoff to flow – as suitable (temporary)

591 measures. These might be adapted for use in badland areas. Nevertheless, for sustainable  
592 management, other sources of suspended sediment – which were found to contribute up to 40% –  
593 should not be neglected.

594 Overall, spectral fingerprints were found to provide a rapid, cost-efficient and non-destructive  
595 alternative to classic fingerprint properties. In the future, it is planned to compare the results of this  
596 spectral fingerprinting approach with “classic” fingerprinting based on geochemistry/radionuclides  
597 using the same samples. Thereby, a composite of spectral and classic properties may enable  
598 discrimination of other sources which could not be determined in this study. In addition, it would be of  
599 interest to study in more detail the sediment sources within the various subcatchments.

600

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611

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767 **Tables**

768

769 **Table 1:** Main physical characteristics of the study catchments, Isábena basin, Spain

Catchment	Area (km <sup>2</sup> )	Min. altitude (m)	Max. altitude (m)	Main lithology	Land use			
					cultivated	grassland	shrubland	forest
Cabecera	145	827	2745	limestones, conglomerates, quartzite	< 1	20	33	45
Villacarli	41	836	2366	marls, limestones	5	12	28	44
Carrasquero	25	728	1539	marls, limestones, sandstones	13	2	36	49
Ceguera	29	606	1314	sandstones	13	< 1	32	54
Lascuarre	44	552	1139	sandstones	30	< 1	15	55
Con-Isábena	160	450	1739	sandstones	24	< 1	31	44

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771

772 **Table 2:** Number of samples collected from potential suspended sediment sources per subcatchment

Catchment	Area [km <sup>2</sup> ]	Number of samples							total
		cultivated	grassland	shrubland	forest	badland	open slope	road	
Cabecera	145	1	10	7	6	-	-	3	<b>27</b>
Villacarli	41	3	1	5	4	6	1	1	<b>21</b>
Carrasquero	25	3	1	4	2	4	1		<b>15</b>
Ceguera	29	4		3	6	1	4	2	<b>20</b>
Lascuarre	44	8	1	10	5	-	8	3	<b>35</b>
Con-Isábena	160	8	3	7	7	3	5	1	<b>34</b>
<b>Total</b>		<b>27</b>	<b>16</b>	<b>36</b>	<b>30</b>	<b>14</b>	<b>19</b>	<b>10</b>	<b>152</b>

773

774 **Table 3:** Number of suspended sediment samples collected during different events at the outlet of the  
 775 main channel and at four subcatchment outlets, along with information on suspended sediment  
 776 concentration (SSC). Vaues are number of samples measured with the spectrometer, with the total  
 777 number of samples collected during the event in parentheses

Date	Catchment	SSC [g l <sup>-1</sup> ]			
		number of samples	min	mean	max
24 Sept 2011	Villacarli	5 (16)	11.08	151.08	332.62
24 Oct 2011	Villacarli	5 (15)	4.55	63.09	215.33
27 Oct 2011	Villacarli	5 (16)	2.13	32.54	99.94
03 Nov 2011	Villacarli	3 (12)	3.68	22.57	101.60
24 Sept 2011	Carrasquero	3 (03)	40.20	71.00	116.57
24 Oct 2011	Carrasquero	5 (16)	1.06	6.19	22.05
27 Oct 2011	Carrasquero	3 (11)	2.03	5.72	10.81
03 Nov 2011	Carrasquero	4 (10)	3.02	8.23	40.22
24 Sept 2011	Ceguera	3 (16)	9.87	25.86	54.25
03 Nov 2011	Ceguera	3 (24)	3.09	12.48	30.79
24 Sept 2011	Lascuarre	3 (14)	3.55	7.00	11.32
03 Nov 2011	Lascuarre	4 (23)	0.98	11.82	42.25
24-25 Sep 2011	Capella	24 (24)	3.28	10.36	35.72
22 Mar 2012	Capella	5 (05)	4.71	11.73	19.88
03-04 Jun 2012	Capella	8 (08)	2.54	34.23	65.25
20 Jun 2012	Capella	7 (07)	1.78	3.60	5.47

778

779 **Table 4:** Number of channel bed resuspension samples collected on four dates, once each season, in  
 780 2011. Sampling site location can be seen in Fig. 1. \* indicates that the amount of material was not  
 781 enough for measurement or that filters were damaged

25.02.2011	24.05.2011	26.07.2011	01.12.2011
(I)	(II)	(III)	(IV)
*	*	S01	S01
*	S02	S02	S02
S03	S03	S03	S03
*	*	S04	S04
S05	S05	S05	S05
*	S06	*	S06
S07	S07	S07	S07
S08	S08	S08	S08
S09	S09	S09	S09
S10	S10	S10	S10
S11	S11	S11	S11
S12	S12	S12	S12
S13	S13	S13	S13
*	S14	S14	S14

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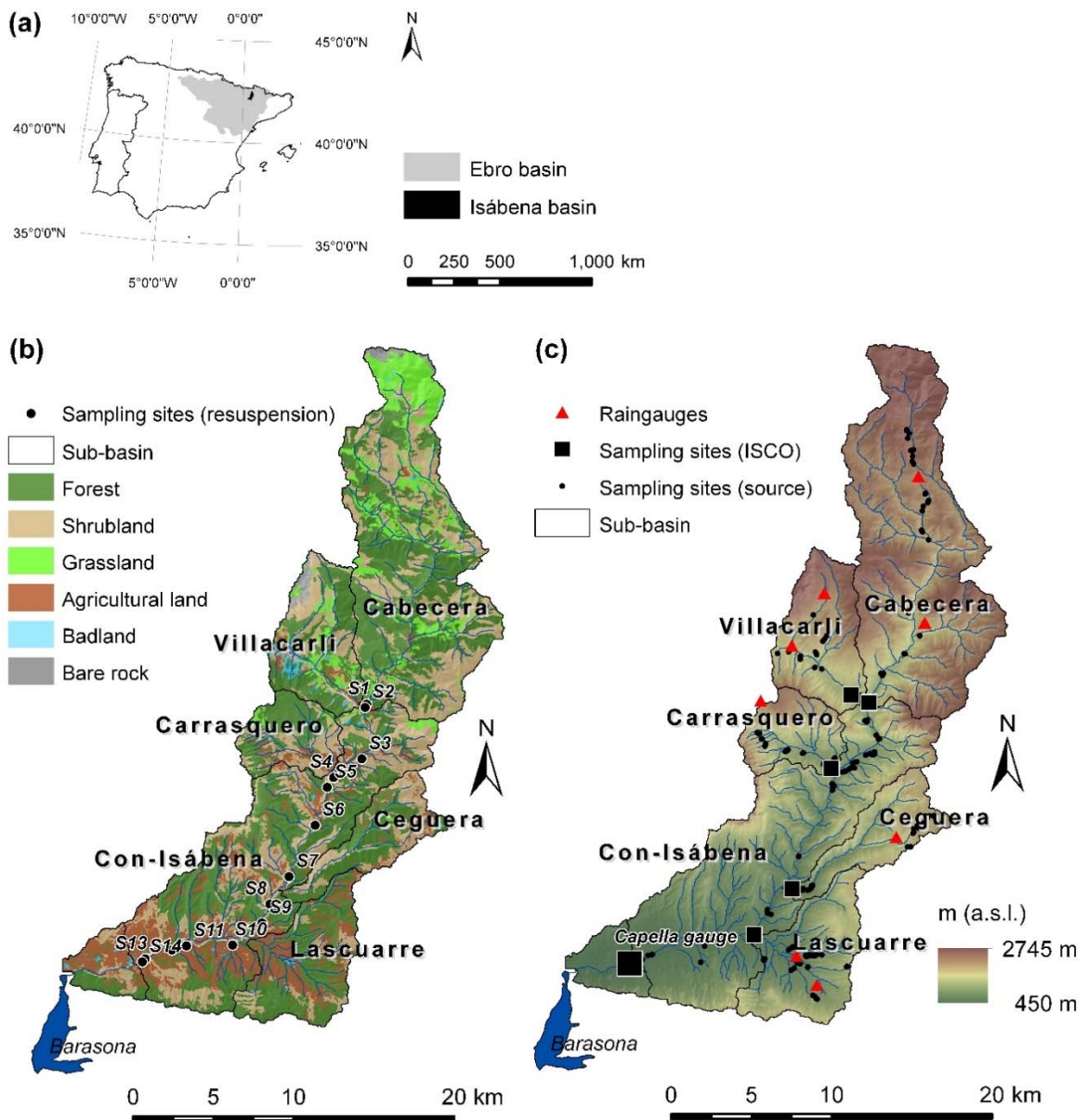
783 **Table 5:** Characterization of precipitation events that caused the runoff sampled for suspended  
784 sediment at Capella. Av gives the average amount of precipitation interpolated over the study area  
785 from information measured at the subcatchment gauges. Events per subcatchment are summaries of  
786 events that occurred in the period given in column date, whereas an event was defined as continuous  
787 rainfall with interruptions < 1 hr. The numbers in brackets represent the number of rain gauges per  
788 subcatchment. Maximum intensity was calculated for 15 minute intervals

Date	Catchment	Precipitation (mm)		Duration (h)			Max. intensity (mm h <sup>-1</sup> )
		mean	sum	min	mean	max	
18-24 Sep 2011	Cabecera (2)	7.28	118.30	1.40	2.24	3.37	20.27
Av 39.86 mm	Villacarli (2)	5.93	63.98	0.23	1.74	2.82	31.79
	Carrasquero (1)	10.55	57.96	2.68	2.75	2.82	68.04
	Ceguera (1)	4.86	28.14	0.35	1.13	3.15	9.22
	Lascuarre (2)	5.46	31.36	0.43	1.78	2.88	40.37
18-22 Mar 2012	Cabecera (2)	2.45	162.45	0.20	8.48	19.85	12.00
Av 67.99 mm	Villacarli (2)	1.93	142.42	0.22	4.77	19.13	20.95
	Carrasquero (1)	2.57	72.66	0.68	5.55	19.38	16.38
	Ceguera (1)	2.65	91.01	1.43	6.41	19.25	20.09
	Lascuarre (2)	2.26	110.03	0.43	3.18	10.08	16.46
28 May 2012 -	Cabecera (2)	5.34	58.10	0.27	1.17	2.10	28.80
03 Jun 2012 Av 25.46 mm	Villacarli (2)	10.24	170.32	0.73	1.99	6.02	102.51
	Carrasquero (1)	9.37	44.46	1.25	1.70	2.20	37.62
	Ceguera (1)	4.20	16.94	0.25	0.92	1.63	8.27
	Lascuarre (2)	8.61	38.40	0.63	1.23	1.70	42.55
18 - 23 Jun 2012	Cabecera (2)	1.90	40.48	0.62	4.25	12.97	18.48
Av 40.22 mm	Villacarli (2)	2.62	99.48	0.40	4.50	13.73	15.68
	Carrasquero (1)	2.53	92.04	0.50	4.35	12.42	15.96
	Ceguera (1)	2.29	44.13	1.00	4.19	12.32	13.97
	Lascuarre (2)	1.62	45.92	0.47	2.98	12.30	9.21

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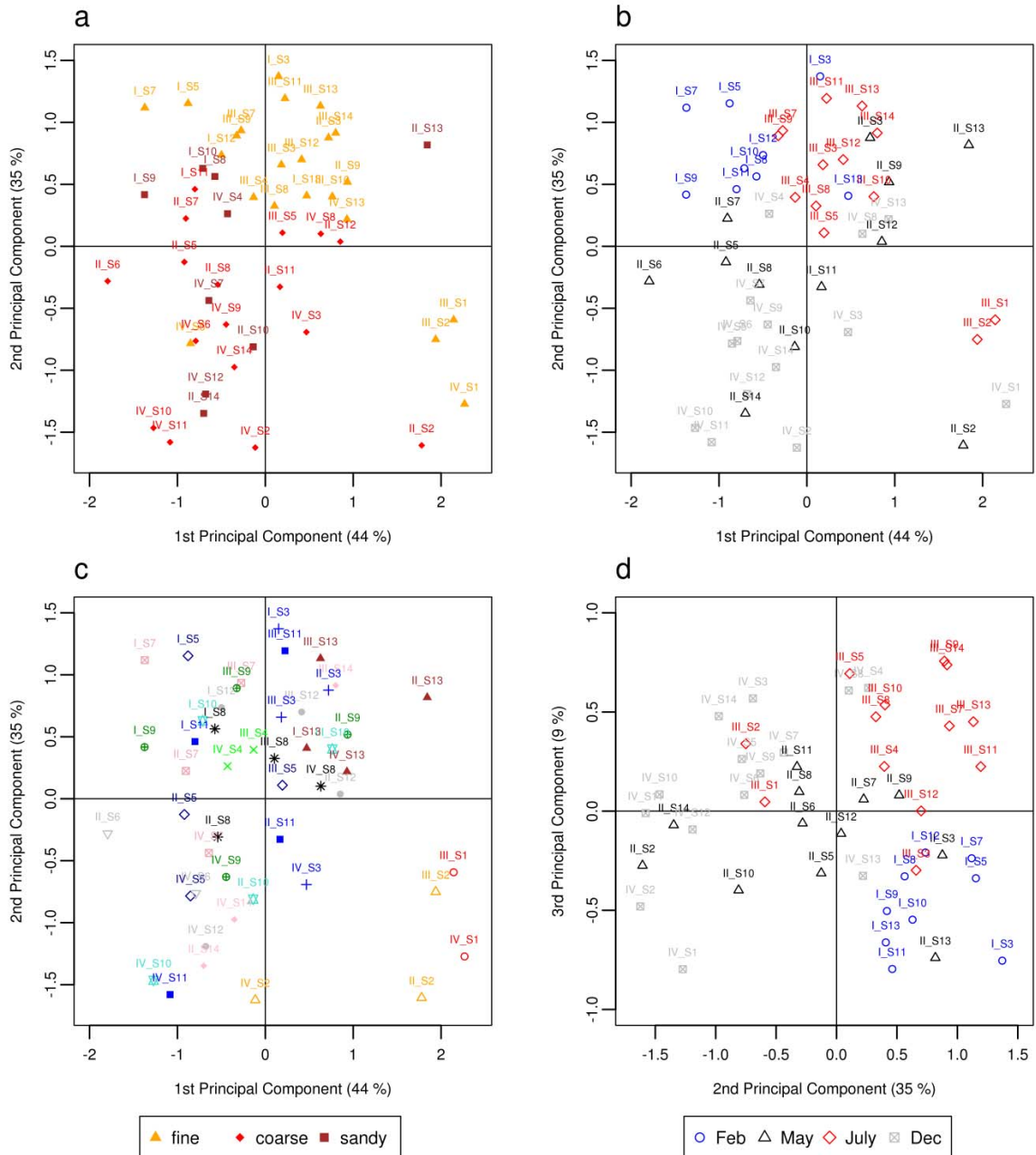


790 **Figures**



791

792 **Fig. 1:** Study area (Isábena basin) including location in Spain, land use, topography, main river  
 793 network, subcatchments, and sampling locations (source sampling, ISCOs for suspended sediment,  
 794 channel bed resuspension sampling, rain gauges)



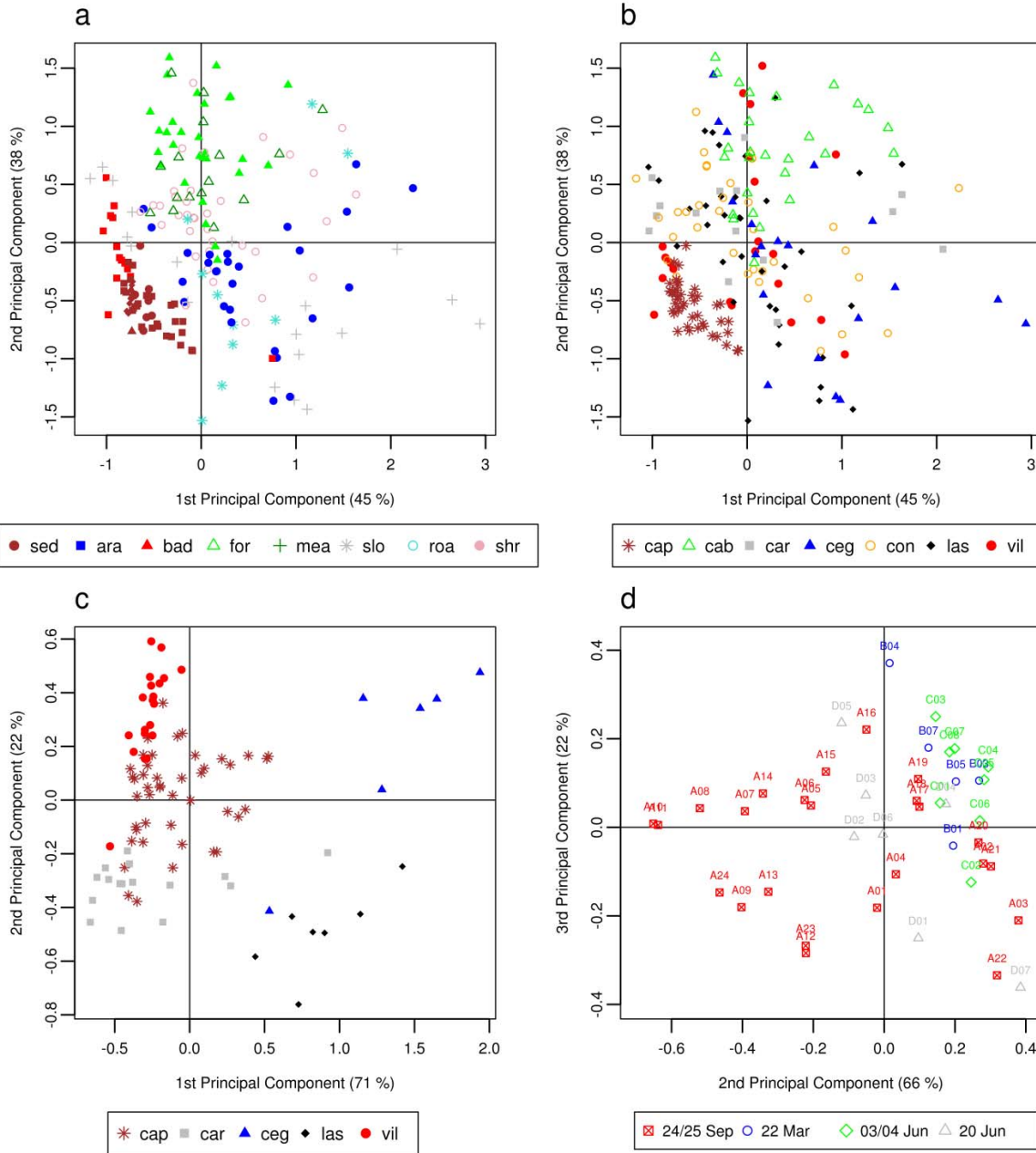
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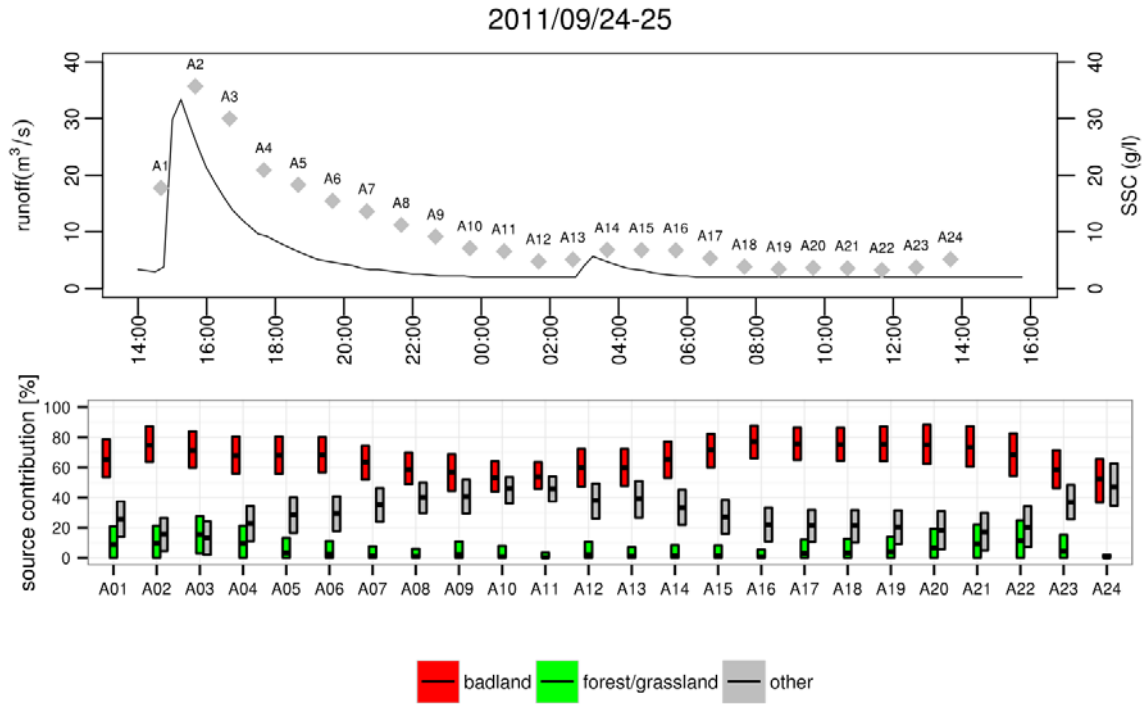
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**Fig. 2:** Two-dimensional scatter plot of scores for the first three principal components (PC) from the PCA for channel bed resuspension samples: a) first two PCs by grain size; b) first two PCs by season; c) first two PCs by sampling location; and d) second and third PC by season



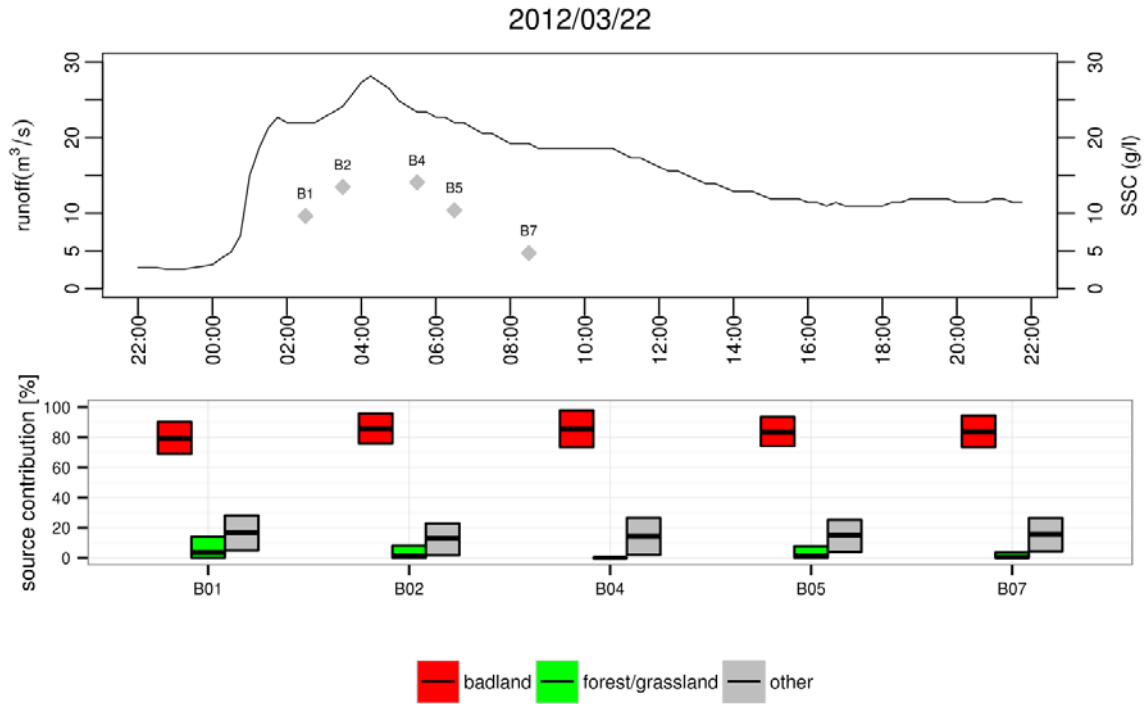
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800 **Fig. 3:** Two-dimensional scatter plot of scores for the first and second principal component (PC) from  
 801 the PCA for: a) source data by land use with suspended sediment sampled at the basin outlet  
 802 (Capella); b) source data by subcatchment with suspended sediment sampled at the basin outlet  
 803 (Capella); c) suspended sediment samples from the basin outlet (cap) and four subcatchments; and  
 804 d) suspended sediment samples from the basin outlet by flood event. Acronyms describe Capella  
 805 suspended sediment (sed) and source sample land use classes (ara = agricultural land, bad =  
 806 badland, for = forest, mea = grassland, slo = open slope, roa = unpaved road, shr = shrubland) and  
 807 subcatchment/location of ISCO sampler (cap = Capella, Cab = Cabecera, car = Carrasquero, ceg =  
 808 Ceguera, las = Lascuarre, vil = Villacarli)



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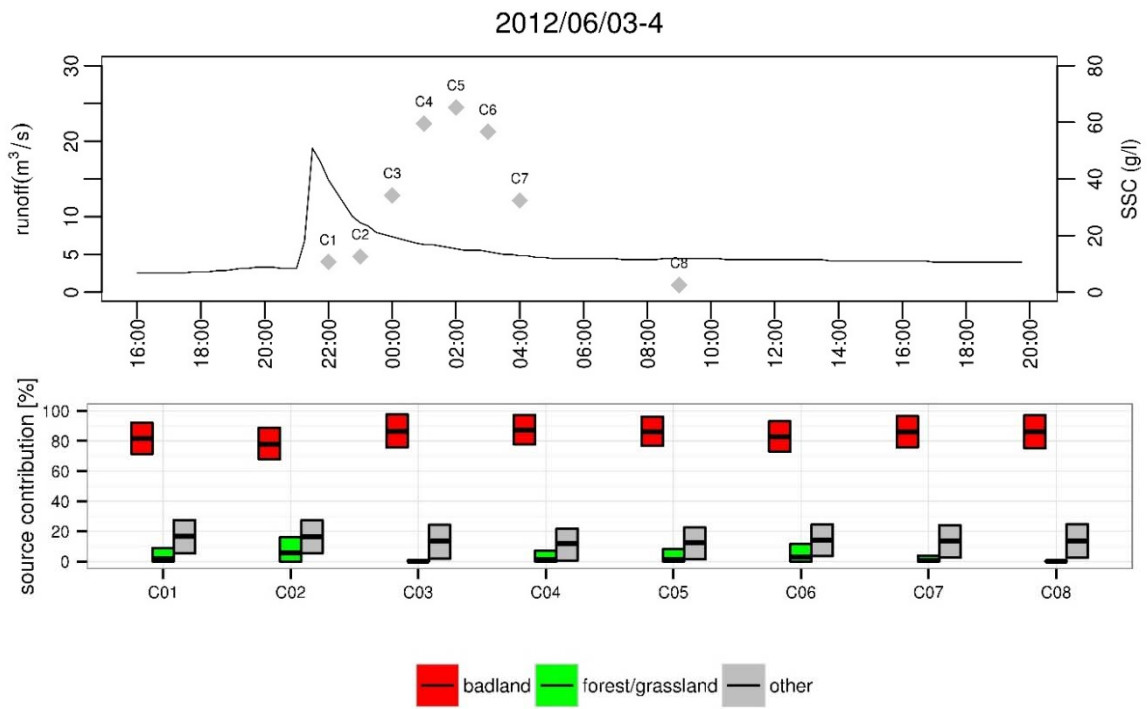
810 **Fig. 4:** Discharge, suspended sediment concentration (SSC) and estimated suspended sediment  
 811 source contribution for the event that occurred on 24th/25th September 2011. Black bars indicate  
 812 mean estimates, confidence intervals are at the 90 % level



813

814 **Fig. 5:** Discharge, suspended sediment concentration (SSC) and estimated suspended sediment  
 815 source contribution for the event that occurred on 22th March 2012. Black bars indicate mean  
 816 estimates, confidence intervals are at the 90 % level

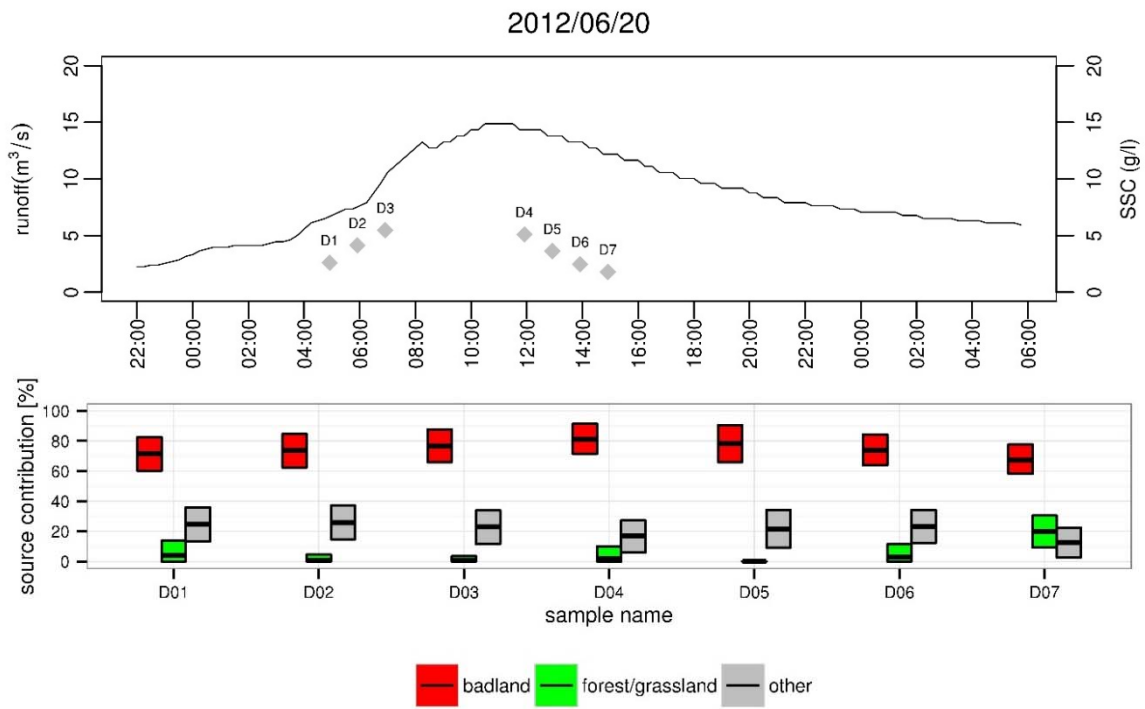
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819 **Fig. 6:** Discharge, suspended sediment concentration (SSC) and estimated suspended sediment  
820 source contribution for the event that occurred on 3rd/4th June 2012. Black bars indicate mean  
821 estimates, confidence intervals are at the 90 % level

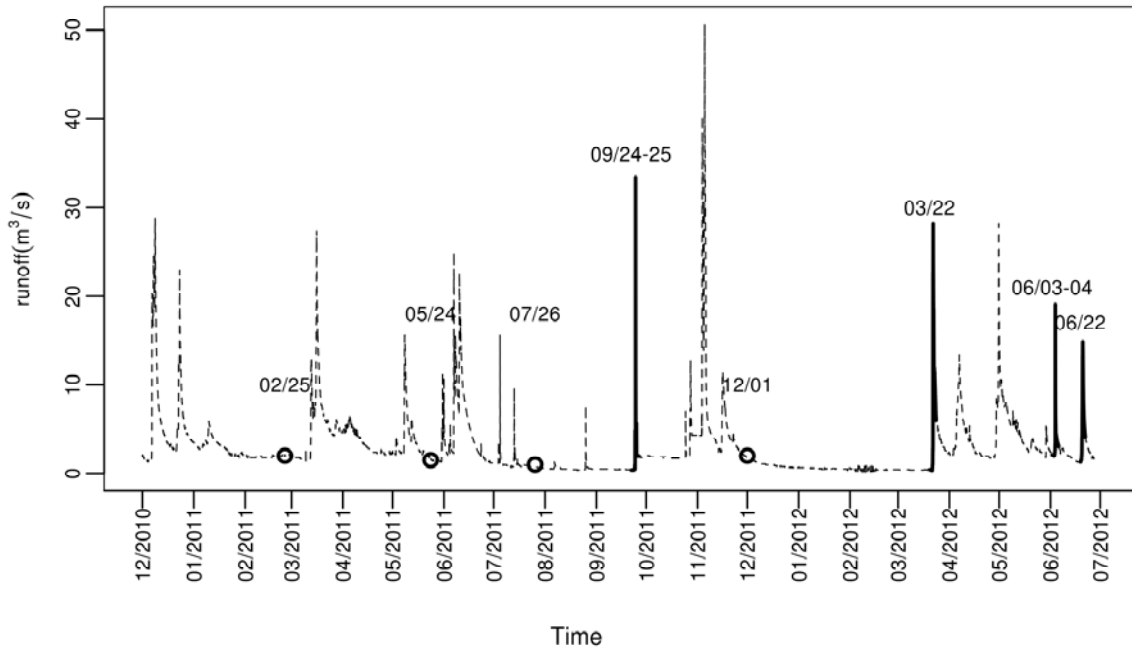
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824 **Fig. 7:** Discharge, suspended sediment concentration (SSC) and estimated suspended sediment  
825 source contribution for the event that occurred on 20th June 2012. Black bars indicate mean  
826 estimates, confidence intervals are at the 90 % level

827 **Online Resources**



828

829 **Online Resource 1:** Discharge at Capella weir over the study period including sampling dates for

830 resuspension samples (circle) and suspended sediments during storms (thick lines)

831

832