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1	Environmental transcription of <i>mmoX</i> by methane oxidizing <i>Proteobacteria</i> in a Sub-								
2	Arctic palsa peatland								
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12	Keywords: soluble methane monooxygenase, methanotrophy, permafrost, carbon turnover,								
13	N <sub>2</sub> -fixation								
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15	Abstract								
16 17 18 19 20 21 22	Methane oxidizing bacteria (MOB) that possess the soluble form of methane monooxygenase (sMMO) are present in various environments, but unlike the prevalent particulate methane monooxygenase (pMMO) the <i>in situ</i> activity of sMMO has not been documented. Here we report on the environmental transcription of a gene ( <i>mmoX</i> ) for this enzyme which was attributed mainly to MOB lacking a pMMO. Our study indicates that the sMMO is an active enzyme in acidic peat ecosystems but its importance for the mitigation of methane releases remains unknown.								
23									
24	Methane oxidizing (methanotrophic) prokaryotes hold an important function in								
25	ecosystems around the globe because they mitigate the release of the greenhouse gas methane								
26	(CH <sub>4</sub> ). Most methanotrophs belong to the <i>Proteobacteria</i> and <i>Verrucomicrobia</i> (28).								
27	Additionally, candidatus Methylomirabilis oxyfera (14) and some methanogenic archaea (3)								
28	can oxidize CH <sub>4</sub> in the absence of oxygen using nitrate and sulfate as electron acceptors,								
29	respectively. The key enzyme, methane monooxygenase (MMO), which oxidizes methane to								
30	methanol at the expense of NAD(P)H, exists in two forms. A particulate, membrane bound								
31	enzyme (pMMO) occurs in nearly all methane oxidizing bacteria (MOB) with the exception								
32	of Methyloferula and Methylocella. These latter two genera feature only a soluble,								
33	cytoplasmic enzyme (sMMO). While the pMMO has a narrow substrate range, oxidizing only								
34	C1-C4 alkanes and alkenes, the sMMO can also utilize C5-C9 alkanes, alkenes, alicyclic and								

aromatic compounds, and the chlorinated compounds trichloroethylene (TCE) and chloroform 35 (6, 17, 33) exposing it to interest for bioremediation and biotechnology applications. The well 36 conserved genes *pmoA* and *mmoX* encode subunits of the pMMO and sMMO, respectively, 37 and are used as functional marker genes for both enzymes (27). To date, environmental 38 transcripts of mmoX have failed to be detected, even in acidic peatlands (5, 20) where pMMO 39 lacking methanotrophs thrive (8, 12, 13). This has led to the suggestion that the pMMO may 40 be involved in methane oxidation here (21), leaving the ecological relevance of the sMMO 41 unclear. 42

Various MOB also possess genes encoding a nitrogenase and can carry out N<sub>2</sub>-fixation
(1). The *nifH* gene is highly conserved, widely used as a functional marker gene for N<sub>2</sub>fixation, and suitable for phylogenetic analysis. Among the four phylogenetic clusters of *nifH*sequences, MOB related *nifH* sequences fall into the 'conventional' Mo-containing *nifH*cluster I (4). Despite the widespread ability among MOB species to fix nitrogen, little
attention has been devoted to N<sub>2</sub>-fixation by MOB in nitrogen-limited, remote ecosystems.

49 In this study we investigated the presence of functional transcripts for the genes *pmoA*, mmoX, and nifH among MOB along the succession of permafrost in a palsa peatland. Palsa 50 peatlands occur at the marginal zone of permafrost distribution and have been widespread 51 throughout Scandinavia until a very few decades ago (31). These usually pristine ecosystems 52 offer heterogeneous substrate conditions for microbial activity and display frozen peat 53 54 mounds (palsas), thermokarst ponds, hollows and hummock like structures in close proximity. At present, palsa formation and collapse are no longer in their natural balance since the 55 permafrost is thawing faster than new palsas can build up. In some areas of northern 56 Scandinavia, climate changes had already 10 years ago caused a decrease in the palsa area by 57 two thirds (26) and palsa regression still continues to be significant. 58

Our study site was located in northern Norway (UMT: 69.694 N; 29.383 E) at the 59 transition from the Arctic to the Sub-Arctic using the 10 degrees July isotherm as the boarder. 60 Annual average temperatures and precipitation from 1965 to2011 were -0.6 °C and 435 mm, 61 respectively (Norwegian Meteorological Institute, Stations Veines/Neiden and Kirkenes 62 Airport). In this period, a positive trend of both annual mean temperature and precipitation 63 was observed (Fig S1). Three different successional palsa stages were selected as sites for 64 sampling and analysis. They covered a currently degrading palsa (DP), a thermokarst pond 65 (TP) adjacent to a DP and a hollow, which represents an old successional stage of a 66 previously collapsed palsa (CP). At the palsas (elevated sites), the vegetation was dominated 67 68 by Ledum palustre, Empetrum sp., Pleurozium sp. and Rubus chamaemorus while the mire

sites were dominated by Eriophorum vaginatum, Andromeda polifolia, Carex rotundata, 69 *Carex cannescens, Carex lapponica, Sphagnum riparium* (TP) and *Sphagnum lindbergii* 70 (CP). We sampled duplicates of pore water for the analysis of vertical concentration profiles 71 72 of methane, ammonium, and nitrate and triplicates of soil cores/blocks per site for the determination of carbon and nitrogen content and for molecular analysis. In addition, plot 73 scale methane emissions were conducted in triplicates per site. The procedures for sampling 74 and analyzing pore water and for methane emission measurements are described in detail 75 elsewhere (24). Ammonium concentrations were determined by an analytical laboratory 76 77 (TosLab, Tromsø, Norway). For molecular analysis, the top layer of fresh plant material was removed and the blocks were sectioned into an upper 10 cm layer and a lower ~15-25 cm 78 79 layer. Subsamples of those sections were pooled, distributed to sterile 50 ml tubes and stored in a liquid nitrogen saturated dry shipper on site. The environmental data of the sampling sites 80 81 are presented in Table 1. Briefly, the pH varied between 4.2 and 4.6 and methane emissions and soil methane concentrations were significant in the thermokarst pond, lower in the 82 83 collapsed palsa site and negligible from the palsa itself. Nitrogen was grossly limited in particular in the top 10 cm layer of the wet sites dominated by Sphagnum. Here, the C/N 84 85 ratios varied between 65 and 97 (compared to 42 in the palsa), and pore water ammonium and nitrate concentrations were below the detection limit of 0.56  $\mu$ M and 2.4  $\mu$ M, respectively. 86 Below a depth of 10 cm, the C/N ratios varied between 30 and 62. Unlike nitrate, ammonium 87 could be detected here but did not exceed  $2.6 \,\mu$ M. 88

The samples preserved for RNA analysis were ground in liquid N<sub>2</sub> to a fine powder. 89 Subsequent extraction of total nucleic acids was carried out in duplicates. Approximately 0.3 90 g of sample were mixed with 0.5 ml of extraction buffer (5% CTAB/120 mM  $K_3PO_4$ , pH8) 91 and subjected to bead beating for 45 s. After phenol-chloroform extraction, nucleic acids were 92 precipitated by incubation with linear acrylamide and 2 volumes of 30% PEG-8000 for 120 93 94 min at room temperature, collected subsequently by centrifugation for 60 min at 4 °C and resuspended in DEPC treated water. To retrieve only the RNA for downstream cDNA 95 96 generation and analysis, the solutions were treated with RNase inhibitor while the DNA was digested with the TURBO DNA-free kit (Ambion). Synthesis of cDNA was carried out using 97 100-500ng RNA as template, random hexamers (Invitrogen) at a final concentration of 500 98 nM, and SuperScript<sup>TM</sup> III Reverse Transcriptase (Invitrogen). The obtained cDNA was first 99 used as template for universal bacterial 16S rRNA gene amplification using RNA as NTC to 100 check for successful cDNA synthesis and complete DNA digestion. Then, the different 101 102 functional genes were targeted. Amplification of pmoA, mmoX, nifH, pxmA and 16S rRNA

gene fragments, cloning, sequencing and bioinformatics was performed as described in thesupplementary material.

We were able to show that *pmoA* gene products are present in all sites, while 105 environmental transcripts of mmoX and nifH were found only in the two wet sites (TP and 106 107 CP). This is, to our knowledge, the first study that reports the detection of environmental transcripts of *mmoX*. Considering the presence of transcripts as an indication for activity, this 108 strongly points at an environmental relevance of the soluble methane monooxygenase in 109 acidic peat ecosystems. The majority of mmoX transcripts were assigned to the group of 110 111 Beijerinckiaceae, presumably to relatives of Methylocella (Fig. 2B) and thus to species that 112 lack a pMMO. *Methylocystis* was the dominant group based on total number of both *pmoA* and mmoX sequences retrieved from DNA, but did not transcribe mmoX. In contrast, the 113 identification of *Methylocystis* related *pmoA* transcripts that belong to a cluster of *pmoA* 114 115 sequences originating solely from peat ecosystems (2) (Fig. 1B) indicates that Methylocystis utilize the pMMO rather than the sMMO. This was expected for MOB hosting both forms of 116 117 MMOs. In addition to mmoX transcripts from the Beijerinckiaceae, transcripts of this gene detected in the collapsed palsa were assigned to an OTU distantly related to Methylomonas 118 119 and thus belong to the group of type I MOB. There are two possible scenarios that could 120 explain this exciting finding. The first is that in contrast to current beliefs, type I MOB that lacks a pMMO does exist. The alternative explanation would be that methanotrophs 121 preferentially transcribe *mmoX* under certain conditions, although they host both the soluble 122 and the particulate enzyme. Thus, our findings pose interesting questions regarding the 123 competition between 'sMMO dependent' methanotrophs and MOB possessing a pMMO with 124 regard to enzyme kinetics, *in-situ* substrate preferences and the general importance of species 125 lacking a pMMO for mitigating methane emissions. Recalling its broad substrate range and 126 the diverse pool of potential compounds in northern peatlands (39), pMMO lacking MOB 127 could even utilize alternative substrates rather than methane. Similarly to Methylocystis, 128 Methylobacter and Methylobacter related sequences (type Ia) were detected in all sites and 129 130 one OTU was represented also by *pmoA* transcripts. Furthermore, active species were identified among Methylomonas and Methylocapsa in the thermokarst pond and in the 131 collapsed palsa, respectively. In the palsa site (DP), cDNA synthesis and detection of mmoX 132 was not successful. Also, the amplification of pmoA and 16S rRNA was problematic with 133 only a nested approach yielding pmoA products. This is indicative for a low abundance of 134 MOB in the palsa site where methane concentrations and emissions were negligible and it is 135 136 consistent with qPCR data (unpublished data).

Targeting PmoA, altogether 12 OTUs were assigned, with the highest species richness 137 in the latest successional stage of palsa degradation (CP) and the lowest in the palsa itself (Fig. 138 1A). The largest number of OTUs based on MmoX was also detected in the late successional 139 stage of CP (Fig. 2A). A summary on the number of DNA and cDNA sequences of each site 140 used for phylogenetic and diversity analysis and the respective amount of OTUs is given in 141 Table S2. Targeting the 16S rRNA gene using MOB specific primers, only 8 OTUs were 142 revealed, indicating primer based failure of detecting some MOB. Nevertheless, all dominant 143 groups that were identified based on functional genes were also found by targeting the 16S 144 145 rRNA gene (Figs. S2 and S3) and rarefaction analysis in general revealed a good coverage of species richness (Fig. 3). 6 OTUs were assigned based on the MmoX of which at least 3 most 146 147 likely lacks a pMMO (pmoA gene), increasing the total number of detected MOB to 15. In order to define the MmoX OTU cutoff on the species level, we used a distance of 4% which 148 149 was based on the correlation between MmoX and 16S rRNA gene sequence distances of selected species (Fig. 3B). Plotting of pairwise distances also proved that *mmoX* is an 150 151 appropriate phylogenetic marker within MOB. In comparison with *mmoX* homologues (Fig. 3C), mmoX seems to have evolved within MOB species and is presumably an essential 152 enzyme. Overall, the diversity of palsa MOB is moderate and ranges between MOB species 153 numbers of rice paddies (25) and Arctic soils (16, 23, 38). However, only a very few species 154 were observed to be active which most likely is a result of the low pH (15, 22, 30). 155 In general, the composition of the MOB community of this palsa peatland and the 156 dominance of Methylocystis related sequences is representative for what has been reported for 157 acidic Sphagnum dominated peat (5, 7-10, 12, 13, 19, 20, 32). Most cultivars from acidic 158 Sphagnum peat are known to be capable of N<sub>2</sub>-fixation which was reported for Methylocapsa 159 acidiphila (12), Methylocella tundrae (8) and Methylocella palustris (13). Also species of the 160 genotypes Methylomonas, Methylocystis (1) and Methylobacter (e.g.(37)) are known to carry 161 out N<sub>2</sub>-fixation. This set of MOB reflects the palsa community. Thus it supports the 162 assumption that nitrogen availability influences soil bacterial communities (29), in particular 163

since in our study MOB related *nifH* sequences made up  $\ge 10\%$  of all sequences both on the

DNA and on the cDNA level (Fig. 4). The presence of MOB related *nifH* transcripts suggests
 a direct compensation for nitrogen deficiency through N<sub>2</sub>-fixating MOB. So far,

167 environmental transcripts of *nifH* related to MOB were neither reported in *Sphagnum* peat nor

in any other pristine and oligotrophic habitat. Our findings point at an important ecosystem

169 function carried out by MOB both in the carbon and the nitrogen cycle of acidic peatlands.

170 The contribution of MOB to N<sub>2</sub>-fixation in acidic peatlands could be substantial considering

- the pronounced number of MOB related *nifH* transcripts detected and should attract more
- 172 attention.
- 173
- 174

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- 182

## 183 **References**

- Auman AJ, Speake CC, Lidstrom ME. 2001. *nifH* sequences and nitrogen fixation in type I and type II methanotrophs. Appl. Environ. Microbiol. 67:4009-4016.
   Bodrassy L. Strolis Payson N. Myrrell JC. Padaiawski S. Weilberter A. Sociitach A. 2002.
- Bodrossy L, Stralis-Pavese N, Murrell JC, Radajewski S, Weilharter A, Sessitsch A. 2003.
   Development and validation of a diagnostic microbial microarray for methanotrophs. Environ Microbiol 5:566-582.
- Boetius A, Ravenschlag K, Schubert CJ, Rickert D, Widdel F, Gieseke A, Amann R, Jorgensen
   BB, Witte U, Pfannkuche O. 2000. A marine microbial consortium apparently mediating
   anaerobic oxidation of methane. Nature 407:623-626.
- Burgmann H, Widmer F, Sigler WV, Zeyer J. 2003. mRNA Extraction and Reverse
   Transcription-PCR Protocol for Detection of *nifH* Gene Expression by Azotobacter vinelandii
   in Soil. Appl. Environ. Microbiol. 69:1928-1935.
- 1955.Chen Y, Dumont MG, McNamara NP, Chamberlain PM, Bodrossy L, Stralis-Pavese N,196Murrell JC. 2008. Diversity of the active methanotrophic community in acidic peatlands as197assessed by mRNA and SIP-PLFA analyses. Environmental Microbiology 10:446-459.
- Colby J, Stirling ID, Dalton H. 1977. The Soluble Methane Mono-oxygenase of Methylococcus
   capsulatus (Bath) ITS ABILITY TO OXYGENATE n-ALKANES, n-ALKENES, ETHERS, AND
   ALICYCLIC, AROMATIC AND HETEROCYCLIC COMPOUNDS. Biochem. J. 165:395-402.
- 201 7. Dedysh SN. 2002. Methanotrophic Bacteria of Acidic Sphagnum Peat Bogs. Microbiology
   202 71:638-650.
- B. Dedysh SN, Berestovskaya YY, Vasylieva LV, Belova SE, Khmelenina VN, Suzina NE,
   Trotsenko YA, Liesack W, Zavarzin GA. 2004. *Methylocella tundrae* sp. nov., a novel
   methanotrophic bacterium from acidic tundra peatlands. Int. J. Syst. Evol. Microbiol. 54:151 156.
- Dedysh SN, Derakshani M, Liesack W. 2001. Detection and enumeration of methanotrophs
   in acidic *Sphagnum* peat by 16S rRNA fluorescence in situ hybridization, including the use of
   newly developed oligonucleotide probes for *Methylocella palustris*. Appl. Environ. Microbiol.
   67:4850-4857.
- Dedysh SN, Dunfield PF, Derakshani M, Stubner S, Heyer J, Liesack W. 2002. Differential
   detection of type II methanotrophic bacteria in acidic peatlands using newly developed 16S
   rRNA-targeted fluorescent oligonucleotide probes. FEMS Microbiol. Ecol. 1466:1-10.

214 11. Dedysh SN, Khmelenina VN, Suzina NE, Trotsenko YA, Semrau JD, Liesack W, Tiedje JM. 215 2002. Methylocapsa acidiphila gen. nov., sp. nov., a novel methane-oxidizing and dinitrogen-216 fixing acidophilic bacterium from Sphagnum bog. Int. J. Syst. Evol. Microbiol. 52:251-261. 217 12. Dedysh SN, Khmelenina VN, Suzina NE, Trotsenko YA, Semrau JD, Liesack W, Tiedje JM. 218 2002. Methylocapsa acidiphila gen. nov., sp. nov., a novel methane-oxidizing and dinitrogen-219 fixing acidophilic bacterium from Sphagnum bog. Int. J. Syst. Evol. Microbiol. 52:251-261. 220 13. Dedysh SN, Liesack W, Khmelenina VN, Suzina NE, Trotsenko YA, Semrau JD, Bares AM, 221 Panikov NS, Tiedje JM. 2000. Methylocella palustris gen. nov., sp. nov., a new methane-222 oxidizing acidophilic bacterium from peat bogs, representing a novel subtype of serine-223 pathway methanotrophs. Int. J. Syst. Evol. Microbiol. 50:955-969. 224 14. Ettwig KF, Butler MK, Le Paslier D, Pelletier E, Mangenot S, Kuypers MMM, Schreiber F, 225 Dutilh BE, Zedelius J, de Beer D, Gloerich J, Wessels HJCT, van Alen T, Luesken F, Wu ML, 226 van de Pas-Schoonen KT, Op den Camp HJM, Janssen-Megens EM, Francoijs K-J, 227 Stunnenberg H, Weissenbach J, Jetten MSM, Strous M. 2010. Nitrite-driven anaerobic 228 methane oxidation by oxygenic bacteria. Nature **464**:543-548. 229 15. Fierer N, Jackson RB. 2006. The diversity and biogeography of soil bacterial communities. 230 Proc. Natl. Acad. Sci. U. S. A. 103:626-631. Graef C, Hestnes AG, Svenning MM, Frenzel P. 2011. The active methanotrophic community 231 16. 232 in a wetland from the High Arctic. Environmental Microbiology Reports **3**:466-472. 233 17. Green J, Dalton H. 1986. Steady-state kinetic analysis of soluble methane mono-oxygenase 234 from Methylococcus capsulatus (Bath). Biochem. J. 236:155-162. 235 18. Grosse G, Harden J, Turetsky M, McGuire AD, Camill P, Tarnocai C, Frolking S, Schuur EAG, 236 Jorgenson T, Marchenko S, Romanovsky V, Wickland KP, French N, Waldrop M, Bourgeau-237 Chavez L, Striegl RG. 2011. Vulnerability of high-latitude soil organic carbon in North America 238 to disturbance. J. Geophys. Res. 116:G00K06. 239 19. Jaatinen K, Tuittila ES, Laine J, Yrjälä K, Fritze H. 2005. Methane-Oxidizing Bacteria in a 240 Finnish Raised Mire Complex: Effects of Site Fertility and Drainage. Microb. Ecol. 50:429. 241 20. Kip N, Ouyang W, van Winden J, Raghoebarsing A, van Niftrik L, Pol A, Pan Y, Bodrossy L, 242 van Donselaar EG, Reichart GJ, Jetten MSM, Sinninghe Damste JS, Op den Camp HJM. 2011. 243 Detection, Isolation, and Characterization of Acidophilic Methanotrophs from Sphagnum 244 Mosses. Appl. Environ. Microbiol. 77:5643-5654. 245 21. Kip N, van Winden JF, Pan Y, Bodrossy L, Reichart GJ, Smolders AJP, Jetten MSM, Damste 246 JSS, Op den Camp HJM. 2010. Global prevalence of methane oxidation by symbiotic bacteria 247 in peat-moss ecosystems. Nat Geosci 3:617-621. 248 22. Lauber CL, Hamady M, Knight R, Fierer N. 2009. Pyrosequencing-Based Assessment of Soil 249 pH as a Predictor of Soil Bacterial Community Structure at the Continental Scale. Appl. 250 Environ. Microbiol. 75:5111-5120. 23. Liebner S, Rublack K, Stuehrmann T, Wagner D. 2009. Diversity of aerobic methanotrophic 251 252 bacteria in a permafrost active layer soil of the Lena Delta, Siberia. Microb. Ecol. 57:25-35. 253 24. Liebner S, Schwarzenbach S, Zeyer J. 2012. Methane emissions from an alpine fen in central 254 Switzerland. Biogeochemistry 109:287-299. 255 25. Lüke C, Krause S, Cavigiolo S, Greppi D, Lupotto E, Frenzel P. 2010. Biogeography of wetland 256 rice methanotrophs. Environmental Microbiology **12**:862-872. 257 26. Luoto M, Seppälä M. 2003. Thermokarst ponds as indicators of the former distribution of 258 palsas in Finnish Lapland. Permafrost and Periglacial Processes 14:19-27. 259 27. McDonald IR, Bodrossy L, Chen Y, Murrell JC. 2008. Molecular ecology techniques for the 260 study of aerobic methanotrophs. Appl. Environ. Microbiol. 74:1305-1315. 28. 261 Pol A, Heijmans K, Harhangi HR, Tedesco D, Jetten MS, Op den Camp HJ. 2007. 262 Methanotrophy below pH 1 by a new Verrucomicrobia species. Nature **450**:874-878. 263 29. Ramirez KS, Lauber CL, Knight R, Bradford MA, Fierer N. 2010. Consistent effects of nitrogen 264 fertilization on soil bacterial communities in contrasting systems. Ecology 91:3463-3470.

265 30. Rousk J, Baath E, Brookes PC, Lauber CL, Lozupone C, Caporaso JG, Knight R, Fierer N. 2010. 266 Soil bacterial and fungal communities across a pH gradient in an arable soil. ISME J 4:1340-267 1351. 268 31. Seppala M. 1986. The origin of palsas, vol. 68. Swedish Society for Anthropology and 269 Geography, Stockholm, Suede. 270 32. Siljanen HMP, Saari A, Krause S, Lensu A, Abell GCJ, Bodrossy L, Bodelier PLE, Martikainen 271 PJ. 2011. Hydrology is reflected in the functioning and community composition of 272 methanotrophs in the littoral wetland of a boreal lake. FEMS Microbiol. Ecol. 75:430-445. 273 33. Smith TJ, Dalton H. 2004. Chapter 6 Biocatalysis by methane monooxygenase and its 274 implications for the petroleum industry, p. 177-192. In Rafael V-D, Rodolfo Q-R (ed.), Studies 275 in Surface Science and Catalysis, vol. Volume 151. Elsevier. 276 34. Sullivan JP, Dickinson D, Chase HA. 1998. Methanotrophs, Methylosinus trichosporium 277 OB3b, sMMO, and their application to bioremediation. Critical Rev. Microbiol. 24:335-373. 278 35. Walter KM, Smith LC, Chapin FS. 2007. Methane bubbling from northern lakes: present and 279 future contributions to the global methane budget. Philosophical Transactions of the Royal 280 Society a-Mathematical Physical and Engineering Sciences 365:1657-1676. 281 36. Walter KM, Zimov SA, Chanton JP, Verbyla D, Chapin FS. 2006. Methane bubbling from 282 Siberian thaw lakes as a positive feedback to climate warming. Nature 443:71-75. 283 37. Wartiainen I, Hestnes AG, McDonald IR, Svenning MM. 2006. Methylobacter tundripaludum 284 sp. nov., a methane-oxidizing bacterium from Arctic wetland soil on the Svalbard islands, 285 Norway (78° N). Int. J. Syst. Evol. Microbiol. 56:109-113. 286 38. Wartiainen I, Hestnes AG, Svenning MM. 2003. Methanotrophic diversity in high arctic 287 wetlands on the islands of Svalbard (Norway)--denaturing gradient gel electrophoresis 288 analysis of soil DNA and enrichment cultures. Can. J. Microbiol. 49:602-612. 39. Williams CJ, Yavitt JB, Wieder RK, Cleavitt NL. 1998. Cupric oxide oxidation products of 289 290 northern peat and peat-forming plants Can. J. Bot. 76:51-62. 291

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Figure 1: A) Venn diagram comparing OTUs at the PmoA level. Numbers in brackets refer to cDNA. B) Neighbor joining tree of partial *pmoA* sequences based on deduced amino acid residues retrieved from the three sampling sites, degrading palsa, thermokarst pond, and collapsed palsa (in bold), compared with public database sequences. Numbers in brackets next to the OTU assignment refer to the number of DNA and cDNA sequences retrieved, respectively. Closed circles mark nodes that were verified by a maximum likelihood tree.

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Figure 2: A) Venn diagram comparing OTUs at the MmoX level. Numbers in brackets refer
to cDNA. B) Neighbor joining tree of partial *mmoX* sequences based on deduced amino acid
residues retrieved from the three sampling sites, degrading palsa, thermokarst pond, and
collapsed palsa (in bold), compared with public database sequences. Numbers in brackets next to
the OTU assignment refer to the number of DNA and cDNA sequences retrieved, respectively.
Closed circles mark nodes that were verified by a maximum likelihood tree.





Figure 3: A) Rarefaction analysis of deduced PmoA, MmoX, and *16S rRNA* gene sequences.
B) Correlation of MmoX versus *16S rRNA* gene sequence distances of 32 methanotrophic species
and C) of 43 species including MmoX homologues.



Figure 4: Neighbor joining tree of partial *nifH* sequences based on deduced amino acid residues retrieved from the studied palsa peatland (in bold) compared with public database sequences. Numbers in brackets next to the cluster labeling refer to the numbers of DNA and cDNA sequences retrieved, respectively, and indicate affiliation with existing and new clusters. The detailed view shows clusters 1.11 and 1.7 which primarily consists of methanotroph retrieved *nifH* sequences. Closed circles mark nodes that were verified by a maximum likelihood tree.

Environmental data of the three sampling sites, degrading palsa (DP), Table 1: thermokarst pond (TP), and collapsed palsa (CP), obtained in July 2010.

	рН	CH <sub>4</sub> (µM)	$CH_4$ emission (mg m <sup>-2</sup> d <sup>-1</sup> )
DP	<b>4.2</b> $(\pm 0.11, n=5)^1$	<b>0.06</b> ( $\pm 0.08$ , n=5) (0-10cm) <sup>2</sup>	b.d.
		<b>b.d.</b> $(20-25 \text{ cm})^2$	
ТР	<b>4.2</b> (±0.3, n=4)	<b>9.15</b> ( $\pm$ 7.76, n=4) (0-10cm) <sup>3</sup>	<b>623</b> (±419, n=3)
		<b>403.51</b> (±58.01, n=4) (20-35cm) <sup>3</sup>	
СР	<b>4.6</b> (±0.6, n=4)	<b>0.04</b> ( $\pm 0.04$ , n=4) (0-10cm) <sup>3</sup>	<b>31</b> (±2, n=3)
		<b>96.15</b> (±106.83, n=4) (20-35cm) <sup>3</sup>	

<sup>1)</sup> measured in July 2011 <sup>2)</sup> values refer to pore gas concentrations <sup>3)</sup> values refer to pore water concentrations b.d.=below detection

# 1 Supplementary material

Gene	Primer	Primer sequence (5'-3')	Annealing	Reference			
	combination		Temp				
			(°C)				
pmoA	A189f	GGNGACTGGGACTTCTGG	62-52	(4)			
	A682	GAASGCNGAGAAGAASGC	(TD)				
	Mb661r	CCGGMGCAACGTCYTTACC		(2)			
pxmA	pxmA 230f	GGCARTGGTGGCCNTTGGT	59	(10)			
	pxmA 732r-1	TGGCGAACCATTTACCGATGTAC					
	pxmA 732r-2	TSGCAAACCACTTGCCGATRTRC					
mmoX	mmoX-206f	ATCGCBAARGAATAYGCSCG	60	(5)			
	mmoX-886r	ACCCANGGCTCGACYTTGAA					
nifH	Pol-f	TGCGAYCCSAARGCBGAC TC	57	(8)			
	Pol-r	ATSGCCATCATYTCRCCGGA					
16S rRNA	MethT1dF	CCTTCGGGMGCYGACGAGT	56.5	(11)			
gene, type I	MethT1bR	GATTCYMTGSATGTCAAGG					
MOB specific							
16S rRNA	typeII-f	GGGAMGATAATGACGGTACCWG	60	(1)			
gene, type II	typeII-r	GATCAARAGCTGGTAAGGTTC					
MOB specific							

2 Table S1: Targeted genes, primer details and PCR conditions of this study

3

- 5 Table S2: Total and site specific numbers of DNA and cDNA sequences used for
- 6 phylogenetic tree construction and diversity analysis and number of assigned OTUs of the
- 7 different genes targeted.

Gene	#sequences DNA			#sequences cDNA			#OTUs DNA				#OTUs cDNA					
	DP	TP	СР	total	DP	ТР	СР	total	DP	ТР	СР	total	DP	ТР	СР	total
ртоА	27	139	179	345	n.d.	19	26	45	4	7	9	12	-	2	4	5
(PmoA)																
ттоХ	n.d.	83	50	133	n.d.	19	70	89	-	2	5	6	-	2	2	4
(MmoX)																
nifH	-	34	33	67	n.d.	29	34	63	-	-	-	-	-	-	-	-
(NifH)																
16S rRNA	-	47	27	74	-	-	-	-	-	6	4	8	-	-	-	-

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11 Figure S1: Climate data of Kirkenes in the period 1965-2011.



- 21
- 22 Figure S3: Neighbor joining tree of type II MOB specific *16S rRNA* gene sequences
- retrieved from the three sampling sites degrading palsa, thermokarst pond, and collapsed palsa
- 24 (in bold) compared with public database sequences.

### 25 *PCR on pmoA, mmoX, nifH, pxmA and 16S, cloning, and sequencing*

26 PCR reactions (50 µl) contained 25 ml PremixF (Epicentre Biotechnologies), gene specific

27 primers (Sigma-Aldrich) at a final concentration of 500 nM, 2.5 U Taq polymerase

28 (Invitrogen), and 5-50 ng of template. The final elongation step was extended to 30 min if

29 subsequent cloning was intended. Primer details, targeted genes and PCR programs are

30 summarized in Table S1. The PCR products were excised, purified (QIAquick gel extraction

kit, QIAGEN), poly-A ligated into a pCR®4-TOPO® vector (Invitrogen) and transformed via

32 heat-shock into chemically competent TOP10 as described by the manufacturer. Overnight at

33 37 °C grown colonies were screened for correct inserts using the gene specific primers and

34 vector primers, respectively. Sanger sequencing was performed by MACROGEN.

35

### 36 Bioinformatics

37 The sequences were processed with the Sequencing Analysis 5.2 software package applying a quality and trimming filter. The post-processed sequences were compared with the GenBank 38 39 database using the blastn, blastx, and vecscreen algorithms. Using the CLC Sequence Viewer 6.5.1 package, PmoA, MmoX, and NifH sequences were deduced and imported into databases 40 maintained with the phylogenetic software package ARB (6). Thereby, the *nifH* database was 41 obtained from the Zehr lab (http://www.es.ucsc.edu/~wwwzehr/research/database/). 42 Phylogenetic trees of the functional genes were constructed selecting the neighbor joining 43 method implemented in the ARB software using a cutoff and a 30% maximum frequency 44 filter for the amino acid sequences (PmoA: 140 aa, MmoX: 360 aa, NifH: 107 aa). 45 Phylogenetic trees of the 16S rRNA were constructed selecting the maximum likelihood 46 method also implemented in the ARB software using a preset termini and a bacterial 47 positional variability filter. The 16S rRNA gene sequences were prior to that aligned with the 48 SINA alignment service (http://www.arb-silva.de/aligner/). Rarefaction analysis and 49 assignment of operational taxonomic units (OTUs) was performed with DOTUR-1.53 fed 50 with distance matrices calculated with ARB using the above mentioned cutoff filters without 51 correction. OTUs were defined using a 7% cutoff for PmoA (3, 7), a 4% cutoff for MmoX 52 (this study, refer to Fig. 4) and a 3% cutoff for the 16S rRNA gene (9). Sequences of pmoA, 53 *pxmA*, 16S rRNA, nifH, and mmeX have been submitted in the GenBank database under 54 accession numbers xxxx-xxxx. 55

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- 57

- 58 1. Chen Y, Dumont MG, Cebron A, Murrell JC. 2007. Identification of active methanotrophs in a 59 landfill cover soil through detection of expression of 16S rRNA and functional genes. 60 Environmental Microbiology 9:2855-2869. 61 2. **Costello AM, Lidstrom ME.** 1999. Molecular characterization of functional and phylogenetic 62 genes from natural populations of methanotrophs in lake sediments. Appl. Env. Microbiol. 63 **65:**5066-5074. 64 3. Degelmann DM, Borken W, Drake HL, Kolb S. 2010. Different Atmospheric Methane-65 Oxidizing Communities in European Beech and Norway Spruce Soils. Appl. Environ. Microbiol. 66 76:3228-3235. 67 4. Holmes AJ, Costello A, Lidstrom ME, Murrell JC. 1995. Evidence that particulate methane 68 monooxygenase and ammonia monooxygenase may be evolutionarily related. FEMS 69 Microbiol. Lett. 132:203-208. 70 5. Hutchens E, Radajewski S, Dumont MG, McDonald IR, Murrell JC. 2004. Analysis of 71 methanotrophic bacteria in Movile Cave by stable isotope probing. Environ Microbiol 6:111-72 120. 73 6. Ludwig W, Strunk O, Westram R, Richter L, Meier H, Yadhukumar, Buchner A, Lai T, Steppi 74 S, Jobb G, Forster W, Brettske I, Gerber S, Ginhart AW, Gross O, Grumann S, Hermann S, 75 Jost R, Konig A, Liss T, Lussmann R, May M, Nonhoff B, Reichel B, Strehlow R, Stamatakis A, 76 Stuckmann N, Vilbig A, Lenke M, Ludwig T, Bode A, Schleifer K-H. 2004. ARB: a software 77 environment for sequence data. Nucl Acids Res **32**:1363-1371. 78 7. Lüke C, Krause S, Cavigiolo S, Greppi D, Lupotto E, Frenzel P. 2010. Biogeography of wetland 79 rice methanotrophs. Environmental Microbiology **12**:862-872. 80 8. Poly F, Monrozier LJ, Bally R. 2001. Improvement in the RFLP procedure for studying the 81 diversity of nifH genes in communities of nitrogen fixers in soil. Res. Microbiol. **152**:95-103. 9. 82 Rosselló-Mora R, Amann R. 2001. The species concept for prokaryotes. FEMS Microbiol. Rev. 83 **25:**39-67. 84 10. Tavormina PL, Orphan VJ, Kalyuzhnaya MG, Jetten MSM, Klotz MG. 2011. A novel family of 85 functional operons encoding methane/ammonia monooxygenase-related proteins in 86 gammaproteobacterial methanotrophs. Environmental Microbiology Reports 3:91-100. 87 11. Wise MG, McArthur JV, Shimkets LJ. 1999. Methanotroph diversity in landfill soil: isolation 88 of novel type I and type II methanotrophs whose presence was suggested by culture-89 independent 16S ribosomal DNA analysis. Appl. Env. Microbiol. 65:4887-4897.
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