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1 A synthesis of methane dynamics in thermokarst lake environments

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17 Abstract

18 Greenhouse gas emissions from physical permafrost thaw disturbance and subsidence, including 19 the formation and expansion of thermokarst (thaw) lakes, may double the magnitude of the 20 permafrost carbon feedback this century. These processes are not accounted for in current global 21 climate models. Thermokarst lakes, in particular, have been shown to be hotspots for emissions 22 of methane (CH₄), a potent greenhouse gas with 32 times more global warming potential than 23 carbon dioxide (CO_2) over a 100-year timescale. Here, we synthesize several studies examining 24 CH₄ dynamics in a representative first-generation thermokarst lake (Vault Lake, informal name) 25 to show that CH₄ production and oxidation potentials vary with depth in thawed sediments 26 beneath the lake. This variation leads to depth-dependent differences in both in situ dissolved 27 CO₂:CH₄ ratios and net CH₄ production responses to additional warming. Comparing CH₄ 28 production, oxidation, and flux values from studies at Vault Lake suggests up to 99% of 29 produced CH₄ is oxidized and/or periodically entrapped before entering the atmosphere. We 30 summarize these findings in the context of CH₄ literature from thermokarst lakes and identify 31 future research directions for incorporating thermokarst lake CH₄ dynamics into estimates of the 32 permafrost carbon feedback.

33 Keywords

34 methane, methane oxidation, permafrost carbon feedback, rapid thaw

35 1. Introduction

36 The Arctic is currently experiencing a geologically abrupt climate warming event where major 37 portions have warmed by more than 1 °C per decade during the past 40 years (Jansen et al., 38 2020). Temperatures in the Arctic are anticipated to increase a further 4-7 °C during the 21st 39 century (Post et al., 2019). Permafrost regions, which represent around a quarter of the Arctic, 40 are estimated to contain almost 1700 Pg of soil organic carbon (SOC) (Schuur et al., 2015), 41 making them a globally important carbon (C) stock. Permafrost warming and thaw, triggered by 42 climate warming, is expected to mobilize ~288 Pg of permafrost SOC by 2300 (Turetsky et al., 43 2020). A portion of this mobilized permafrost C will be microbially mineralized into the 44 greenhouse gasses (GHG) methane (CH₄), carbon dioxide (CO₂), and nitrous oxide (N₂O), which 45 will contribute a positive feedback to climate warming called the permafrost carbon feedback 46 (PCF) and potentially turn the Arctic to a net C source to the atmosphere (Schuur et al., 2015; 47 Voigt et al., 2020). Estimates of the PCF suggest that C released from thawing permafrost can 48 increase global climate warming by 0.13-0.27 °C by 2100 (Schuur et al., 2015). However, these 49 models are based on gradual, top-down permafrost thaw (e.g. active layer deepening). The 50 impact of comparatively rapid physical disturbance thaw processes (e.g. thermokarst and coastal 51 erosion) are not accounted for in current estimates of the PCF (Tanski et al., 2019; Turetsky et 52 al., 2019, 2020).

In the Arctic, lakes have been identified as hotspots where terrestrial C, including C from thawed permafrost, is mineralized and emitted as CH₄ and CO₂ (Kuhn et al., 2018; Walter et al., 2006; Wik et al., 2016). Lakes are a common feature in northern latitudes; by surface area, almost half of the world's lakes are located in Arctic and sub-Arctic regions (Smith et al., 2007; Grosse et al., 2013). Thermokarst (thaw) lake landscapes are widespread in permafrost-underlain

58	regions and cover 1.3 million km ² , representing 20-40% of the permafrost region and 15-75% of
59	arctic lowland surface area (Olefeldt et al 2016). These lakes usually form when ice-rich
60	permafrost thaws or massive ice melts (thermokarst) in areas with low relief, thick
61	unconsolidated sediments, and high ground ice contents (Jones et al., 2011); the formation of this
62	type of lake is expected to become more widespread with climate change (Grosse et al., 2013;
63	Jones et al., 2011). Once formed, thermokarst lake expansion can rapidly introduce new C into
64	the active C cycle. Talik (thaw bulb) development beneath the lake can thaw and mobilize older,
65	deeper C that otherwise would not thaw during this century, even under the most severe active
66	layer deepening scenarios (Langer et al., 2016; Walter Anthony et al., 2018).
67	Thermokarst lakes currently contribute almost a quarter (4.1 \pm 2.2 Tg CH ₄ yr ⁻¹ ; Wik et al.,
68	2016) of the total CH ₄ emitted annually from northern lakes. CH ₄ and CO ₂ emissions from
69	thermokarst lakes are expected to increase five-fold during the coming century, and models
70	suggest that accounting for GHG produced in and released from thermokarst lakes will more
71	than double the PCF this century (Schneider von Deimling et al., 2015; Walter Anthony et al.,
72	2018). Much of the C used to produce GHG originates from the surrounding terrestrial
73	environment, including reworked eroded and deposited sediments and permafrost thawing
74	beneath the lake. Carbon mass balance measurements of paleo thermokarst lake taliks by Walter
75	Anthony et al. (2014) suggest $28 \pm 12\%$ of thawed C beneath thermokarst lakes is converted to
76	GHG over the lakes' millennial-scale lifespans. Additional, modern C can be contributed by
77	primary productivity and vegetation within and surrounding the lakes (Dean et al., 2020; Elder et
78	al., 2018).
79	Understanding CH4 emissions from thermokarst lakes is particularly important given that

80 CH₄ has 96x more global warming potential (GWP) than CO₂ over a 20-year timescale and 32x

81 more GWP than CO₂ over a 100-year timescale (Alvarez et al., 2018; Etminan et al., 2016). One 82 approach towards constraining estimates of CH₄ emissions in these rapidly changing landscapes 83 is examining CH₄ emission, production, and oxidation dynamics in thermokarst lake 84 environments. Here, we synthesize research conducted on CH₄ dynamics at a thermokarst lake in 85 central Alaska to provide insights into whole-lake CH₄ cycling in a thermokarst lake system and 86 the role CH₄ emissions from thermokarst lakes may play in the PCF. To our best knowledge, this 87 is the most detailed biogeochemical and microbiological study of a sediment core capturing the 88 full depth profile of a thermokarst lake talik. We used sediment from a lake core collected from 89 the center of Vault Lake (informal name) and extending through the entire talik to examine: (1) 90 CH₄ production potentials (Heslop et al., 2015); (2) aerobic (Martinez-Cruz et al., 2015) and 91 anaerobic (Martinez-Cruz et al., 2017; Winkel et al., 2019) CH₄ oxidization; (3) temperature 92 sensitivities of CH₄ production (Heslop et al., 2019a); and (4) CH₄-producing and oxidizing 93 microbial communities (Heslop et al., 2019a; Martinez-Cruz et al., 2018; Winkel et al., 2019). 94 We also measured *in situ* sediment CH₄ concentrations (Winkel et al., 2019) and the whole-lake 95 CH₄ emission budget (Sepulveda-Jauregui et al., 2015) at the same lake for comparison with 96 results from our laboratory studies. Our examination provides unique and valuable insights on C 97 cycling dynamics in thawed sediments beneath thermokarst lakes. Further, we draw upon 98 research from lakes in other northern and global environments to fill knowledge gaps in CH₄ 99 dynamics not studied at Vault Lake, to place our findings in a global context, and to identify 100 future research directions for incorporating thermokarst lake CH₄ dynamics into estimates of the 101 PCF.

102 **2. Study site: Vault Lake**

103 *Lake location and characteristics*

104	Vault Lake (65.0293 °N, 147.6987 °W, elevation 216 m) is a 3,200 m ² thermokarst lake located
105	in a lowland boreal region of Interior Alaska approximately 40 km north of Fairbanks. The
106	region has a continental climate (mean annual air temperature -2.39 °C and 274.6 mm mean
107	annual precipitation; 1981-2010 data from the National Climate Data Center) and is
108	characterized by discontinuous permafrost reaching up to 120-m thick (Jorgenson et al., 2008).
109	At the Vault site, permafrost is comprised of 36-m thick ice- (21-208 wt% gravimetric ice
110	content with ice wedges up to 3 m width) and organic-rich (1-3 wt% SOC) aggraded syngenetic
111	loess-like silt (yedoma) underlain by schistose bedrock. These yedoma sediments were largely
112	deposited in the unglaciated region during the early and middle Wisconsin periods, and observed
113	ice wedge and sediment deformation suggests displacement (e.g. from landslides) following
114	permafrost formation (Schirrmeister et al., 2016). Similar ice- and organic-rich yedoma deposits
115	are found in unglaciated lowland regions across Alaska, Northeast Siberia, and Northwest
116	Canada (Olefeldt et al., 2016; Walter Anthony et al. 2018).



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Figure 1. Location map of Vault Lake (a) and historical and recent georeferenced aerial images and LiDAR digital elevation model of the Vault Lake study area: (b) false color near-infrared aerial photo from 22 August 1986 from the Alaska High-Altitude Aerial Photography (AHAP) program; (c) true color aerial photo from 18 August 2007 from the USGS Digital Orthophoto Quadrangle (DOQQ) program for Alaska; and (d) LiDAR-based digital elevation model of the same study area from May 2011 produced by the Alaska Division of Geological and Geophysical Surveys (DGGS).

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Vault Lake is representative of a young thermokarst lake in a yedoma permafrost setting.

127 We note that not all thermokarst lakes are located in yedoma-type permafrost and, compared to

- 128 thermokarst lakes formed in non-yedoma environments, thermokarst lakes in yedoma permafrost
- 129 are typically deeper due to higher ground subsidence from melting massive ice (Grosse et al.,
- 130 2013). Vault Lake has an average water depth of 3.7 m (maximum depth 4.6 m; Heslop et al.,
- 131 2015) and reached a maximum lake ice thickness of 60-80 cm in late winter during 2012-2014
- 132 (Bondurant, *unpublished data*). Temperature and dissolved oxygen profiles measured in the
- 133 water column during February-June 2013 indicate Vault Lake stratifies at 1.0-1.5 m depth (Fig.





151 Figure 2. Water column temperature (a) and dissolved oxygen (b) profiles measured at Vault 152 Lake during February-June 2013. The data suggest Vault Lake stratifies at 1.0-1.5 m depth and 153 show the hypolimnion is anoxic.

154

155 Sediment core

156 In March 2013, we collected a 590-cm long sediment core from the center of Vault Lake. Core 157 drilling, processing, and storage methods are described in detail in Heslop et al. (2015). The 158 Vault Lake core captured the near-complete deposited lake sediment and thawed talik sediment 159 sequence beneath the lake, in addition to the top 40 cm of permafrost beneath the talik. Sediment 160 characteristics (Heslop et al., 2015) and organic matter composition (Heslop et al., 2017) were 161 used to classify the core into five facies: (1) Organic-rich mud (0-152 cm); (2) Lacustrine silt 162 (152-330 cm); (3) Taberite (330-508 cm); (4) Recently thawed taberite (508-550 cm); and (5) 163 Transitional permafrost (550-590 cm). The Organic-rich mud (mean 3.83 wt% SOC; 1.86 wt% 164 IC) and Lacustrine silt (1.04 wt% SOC; 0.56 wt% IC) facies represent terrestrial sediments 165 deposited following the lake formation that were exposed to the lake water column during 166 erosion and re-deposition (Heslop et al., 2015). Taberite (0.84 wt% SOC; 0.38 wt% IC) represents yedoma sediment which thawed in situ and remained underneath the lake (Walter 167 168 Anthony et al., 2014; Schirrmeister et al., 2011); the bottom 43 cm of the taberite, representing 169 the most recently thawed sediments, was designated as Recently-thawed taberite (1.36 wt% 170 SOC; 0.70 wt% IC). The Transitional permafrost (1.52 wt% SOC; 0.88 wt% IC) facies consisted 171 of permafrost beneath the talik thaw front that was frozen at the time of core collection; we refer 172 to this as "transitional permafrost" because it is close to the thaw transition and therefore 173 contains a large amount of unfrozen water in the inter-pore space (Williams and Smith, 1989). 174 Lake age

175 We estimate Vault Lake formed 100-400 years before present (yr BP). Using AMS radiocarbon 176 ages of macrofossils picked from the Vault Lake sediment core, we had originally estimated the 177 lake age to be around 400 yr BP (Heslop et al., 2015). However, estimating the date of lake 178 formation using expansion rates between 1951 and 2007 shorelines (Fig. 3) yields a lake age 179 estimate of 100 yr BP. Using the depth of the deposited Organic-rich mud (0-152 cm) facies and 180 assuming a sediment deposition rate of 0.2 to 0.8 cm yr⁻¹ yields lake age estimates of 190-760 yr 181 BP. The variance in lake age estimates obtained using different methods highlights the 182 challenges of accurately dating thermokarst lake initiation. Using AMS radiocarbon ages of 183 macrofossils has the potential to overestimate lake age if pre-aged materials enter the sediment 184 column. Similarly, lake expansion and sediment disposition rates are unlikely to be constant over 185 the lakes' history, which can lead to over- or underestimating lake age. These challenges to 186 accurately estimating thermokarst lake age point towards the difficulties of modeling and 187 projecting CH₄ emissions from thermokarst lakes in the coming centuries, given that younger 188 thermokarst lakes typically have higher CH₄ emission rates than older thermokarst lakes 189 (Turetsky et al., 2020; Walter Anthony et al., 2018).





194 *Talik temperatures*

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195 Temperatures measured in two vertical profiles within the talik over a seven-year period (2013-196 2020; Fig. 4) indicate sediments in the lake center (4.0 m water depth) range from -0.40 to 4.51 197 $^{\circ}$ C (annual mean \pm SD 1.11 \pm 1.42 $^{\circ}$ C) and sediments near the lake margin (1.4 m water depth) 198 range from -0.40 to 14.45 °C (annual mean \pm SD 2.50 \pm 3.37 °C). Temperatures at all depths in 199 both talik profiles were significantly correlated with atmospheric temperatures. Temperatures 200 from the shallower sediment depths (0.5 and 1.0 m) in the profile near the lake margin were 201 positively correlated with atmospheric temperatures (Pearson coefficient r = 0.59 and 0.44, 202 respectively; p < 0.001), while temperatures in the same sediment depths in the lake center were 203 negatively correlated with atmospheric temperatures (Pearson coefficient r = -0.34 and -0.52; p 204 < 0.001). This is most likely a result of: (1) the deeper water in the lake center causing less

efficient heat transfer from the atmosphere to the sediments, resulting in a lagged temperature response compared to the sediments in the lake margin profile, and (2) water beneath the hypolimnion in the lake center remaining colder during the thaw season than water in the shallower littoral zones.



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Figure 4. Mean daily atmospheric temperatures measured at Fairbanks, Alaska (a; data from ACIS) and hourly temperatures measured at the Vault Lake talik in the lake center (b; 4.0 m water depth) and near the lake margin (c; 1.4 m water depth). Temperature data were recorded continuously, except for gaps and sensor errors noted in gray on the figure.

- 214
- 215 **3. Methane emission**

216 High latitude lakes emit CH₄ by four main seasonal pathways: diffusion, direct ebullition, ice

- 217 bubble storage flux, and storage flux turnover events (Greene et al. 2014). Atmospheric diffusion
- 218 occurs during ice-free conditions when the lake water becomes oversaturated in CH₄, leading to

219 positive diffusive flux to the atmosphere. Diffusive CH₄ flux is the lowest contribution of CH₄ to the atmosphere, contributing approximately 5% of total annual emissions from interior Alaska 220 221 yedoma lakes (Sepulveda-Jauregui et al., 2015). Ebullition is the dominant pathway of CH₄ 222 emission in most thermokarst lakes (Walter et al., 2006). Direct ebullition occurs when high 223 concentrations of CH_4 , which is less soluble than CO_2 , accumulate in sediment pore space. When 224 concentrations exceed the solubility limits, free gas bubbles form. Bubbles form and migrate 225 through secondary pore channels in sediments to release gas to the atmosphere (Liu et al., 2016; 226 Scandella et al., 2011), with the highest ebullition rates occurring when hydrostatic pressure 227 drops (Varadharajan, 2009). In winter surface lake ice impedes bubble release to the atmosphere, 228 entrapping many CH₄ ebullition bubbles in and under surface lake ice. Up to 80% of their CH₄ 229 content may be lost to the water column when CH4 diffuses out of bubbles before the ice 230 thickens around them. When the ice melts in spring, the remaining entrapped CH₄ is released to 231 the atmosphere through a process called ice bubble storage flux (Greene et al., 2014). Combined, 232 ebullition and ice bubble storage flux typically account for around 86% of annual CH₄ emissions 233 from thermokarst lakes formed in yedoma-type permafrost (Sepulveda-Jauregui et al., 2015). 234 Elevated concentrations of dissolved CH₄ in the water column beneath ice are also subject to 235 diffusive flux as the ice melts and during overflow events (Greene et al., 2014). These emissions 236 associated with springtime storage flux turnover events, which derives largely from dissolution 237 of ebullition bubbles beneath ice in winter (Greene et al., 2014), contributes 1.2% of total annual 238 CH₄ emissions (Sepulveda-Jauregui et al., 2015).

At Vault Lake, total annual CH₄ emissions were estimated to be 40.9 g CH₄ m⁻² yr⁻¹; 77% of the annual CH₄ flux is from ebullition, 11% is from ice bubble storage flux, and 12% is from diffusive flux (Sepulveda-Jauregui et al., 2015). Diffusive flux and ice bubble storage flux values

242 at Vault Lake are near the median for the yedoma lakes, while ebullitive flux is slightly higher 243 than the median but still within the interquartile range for yedoma lakes (Fig. 5). Compared to 244 other reported CH₄ emissions values from northern lakes, net Vault Lake CH₄ emission values 245 are near mean values for yedoma-type lakes (50.2 g CH₄ m⁻² yr⁻¹; n = 13 lakes) and more than non-yedoma type lakes (mean 6.37 g CH₄ m⁻² yr⁻¹; n = 36 lakes). This is consistent with prior 246 247 research suggesting CH₄ emissions from thermokarst lakes in regions underlain by yedoma-type 248 permafrost emit approximately 6x more CH₄ than lakes underlain by other permafrost types 249 (Sepulveda-Jauregui et al., 2015). It is important to note that, as a relatively young thermokarst 250 lake (100 - 400 yr BP), Vault Lake functions as a net C source to the atmosphere. As thermokarst 251 lakes mature and eventually drain (typically over millennial time scales), their net C emissions 252 become lower and, in some cases, negative due to C being sequestered in increased vegetative 253 biomass (Walter Anthony et al., 2014). Emissions from newly formed thermokarst lakes (< 60 254 years old) are typically even higher than from older thermokarst lakes (Walter Anthony et al., 255 2018; Turetsky et al., 2020), such as Vault Lake. This variance in C emissions amongst 256 thermokarst lakes highlights the importance of examining lakes in different environments to 257 constrain their potential circum-arctic GHG emissions.





259 Figure 5. Magnitudes of CH_4 emission pathways from northern yedoma (n = 13) and nonyedoma (n = 36) lakes. IBS refers to ice bubble storage flux. Significant (Two-sample t test; $p < 10^{-10}$ 260 0.05) differences between yedoma and non-yedoma lakes are denoted. Emissions at Vault Lake, 261 measured by Sepulveda-Jauregui et al. (2015), are highlighted in red. CH₄ flux values for 262 263 diffusive flux and IBS at Vault Lake are near the median for the vedoma lakes, while ebullitive 264 flux is slightly higher than the median, but still within the interquartile range. Spring storage flux 265 following ice-out was not determined at Vault Lake. Data are reported by Huttunen et al. (2003), 266 Repo et al. (2007), Sepulveda-Jauregui et al. (2015), Walter Anthony et al. (2010), and Zimov et al. (1997). 267

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269

Within individual thermokarst lakes, CH₄ emissions vary spatially and temporally.

270 However, most studies of northern lakes only report diffusive CH₄ emissions based on sampling

- 271 < 1% of the lake surface area at 1-2 points in time (Engram et al., 2020). Further, the sporadic
- 272 nature and spatial variability of ebullitive flux make it difficult to accurately estimate its
- 273 magnitude. The bubble trap method, which is frequently used to measure ebullitive flux, is
- strongly biased as: (i) bubble traps are intentionally fixed over ebullition seeps; (ii) it is almost
- 275 impossible to cover an stratified analysis for ebullition events due to variation in bottom
- topography, water depth, and sediment properties; (iii) the different number of traps installed in

277 the ecosystems generates large uncertain in the literature reported; and (iv) the season and time 278 measuring bubbles by traps varies between studies and sites (Wik et al., 2016). This variation 279 makes it difficult to accurately quantify and compare CH₄ emissions from circum-arctic lakes. In 280 addition to the lack of information on spatial and temporal variability, measurements of CH4 281 emissions from lakes are scarce compared to the number of lakes. A 2016 synthesis by Wik et al. 282 collected all known field measurements from northern lakes; the lakes studied (n = 733)283 represented only 0.02% of the 3.8 million lakes north of 50° N (Engram et al., 2020). This 284 highlights the needs for additional data collection to constrain the spatial and temporal variability 285 in CH₄ emissions from both thermokarst and non-thermokarst northern lake systems. Remote 286 sensing methods have been suggested as one approach for constraining spatial and temporal 287 variability in circum-arctic lake CH₄ emissions (Engram et al., 2020; Matthews et al., 2020). 288 While few field studies measure year-round temporal variability in CH₄ emissions, results 289 from a process-based climate-sensitive lake biogeochemical model indicate that seasonal 290 variation in CH₄ emissions can be explained by energy input and substrate availability (Tan et 291 al., 2015). High temporal resolution CH₄ emissions have not been quantified at Vault Lake, but 292 Sepulveda-Jauregui et al. (2015) determined annual summer CH₄ flux (ebullition and diffusive flux; 31.4 g CH₄ m⁻² yr⁻¹) was approximately 3 times greater than annual winter CH₄ flux 293 (ebullition and ice bubble storage; 9.4 g CH₄ m⁻² yr⁻¹). Spatially, both field observations (Engram 294 295 et al., 2020; Walter Anthony et al., 2016) and modeling results (Tan et al., 2015) show CH4 296 emissions along the thermokarst margins are typically higher than emissions from non-297 thermokarst margins and centers in the same lake. Mapping of ebullition hotspots at Vault Lake 298 during October 2014 suggest ebullition seep locations are generally more abundant in areas 299 where the lake has expanded since 1951 (Fig. 3).

300

0 4. Methane production

301 Methane emissions from northern lakes originate from different sources, including fossil 302 geological (microbial and thermogenic) sources and more recent microbial sources (Walter 303 Anthony et al., 2012). Recently, in temperate lakes, CH₄ production in the water column has also 304 been found to be associated with phytoplankton under oxic conditions (Günthel et al., 2020); 305 however, it is currently unknown if oxic CH₄ production is also present in northern lakes. δ^{13} C-306 CH₄ values collected from ebullition bubbles at Vault Lake (- $67.7 \pm 3.4\%$; Winkel et al., 2019), 307 in combination with Vault Lake having a closed talik without geologic CH₄ seeps, indicate CH₄ 308 production at Vault Lake is microbial. Microbial production of CH₄ during anaerobic respiration, 309 termed methanogenesis, occurs when Archaea called methanogens use simple substrate (CO_2 , 310 H₂, acetate, methyl compounds) to produce CH₄. Methanogenesis using CO₂ and H₂ as substrate, 311 which is carried out by hydrogenotrophic methanogens (autotrophy), occurs when CO₂ is 312 reduced to CH₄ according to the reaction: 313 $CO_2 + 4H_2 \rightarrow CH_4 + 2H_2O$ (Equation 1) 314 Methanogenesis using acetate as a substrate (heterotrophy), carried out by acetoclastic 315 methanogens, occurs when acetate is reduced to CH₄ according to the reaction: 316 $C_2H_3O_2^- + H^+ \rightarrow CH_4 + CO_2$ (Equation 2) 317 The primary pathway for CH₄ production (hydrogenotrophic vs. acetoclastic) varies both 318 among yedoma thermokarst lakes and within the same lake depending on physiochemical 319 parameters and substrate availability (Walter et al. 2008). At Vault Lake, a combination of 320 archaeal sequencing measured from sediments and δ^{13} C-CH₄ values measured from ebullition 321 bubble gas indicates both hydrogenotrophic and acetoclastic methanogenesis by 322 Methanosarcina, Methanosaeta and Methanoregula (Winkel et al., 2019). A less common

pathway of methanogenesis that has been shown to be activated in thawing permafrost uses
methyl compounds such as methylamines as a substrate (Canfield et al. 2005; Coolen and Orsi,
2015; Tveit et al., 2015). While we did not observe evidence of this CH₄ producing pathway at
Vault Lake, incubations of thermokarst lake sediments from Utqiagvik, Alaska (de Jong et al.,
2018) and Interior Alaska (Liebner, *unpublished data*) observed stimulation of methylotrophic
methanogens following substrate addition.

329 Through an anaerobic incubation study of the Vault Lake sediment core, we found that 330 two-thirds of total CH₄ measured during a 175-day incubation at 3 °C originated in the surface 331 organic-rich mud facies (0-152 cm depth), despite this facies only representing a quarter of the 332 core's total 590 cm thickness (Heslop et al., 2015). The organic-rich mud facies comprise of 333 reworked allochthonous and autochthonous materials deposited since lake formation, as 334 described by Farquharson et al. (2016) for other yedoma lakes. Archaeal communities were 335 dominated (84% of detected sequences) by the acetoclastic methanogens Methanosaetaceae, 336 with only a few hydrogenotrophic *Methanoregulaceae* and *Methanosarcinacae* (Heslop et al., 337 2019; Winkel et al., 2019). Given that all species within *Methanosaetaceae* only use acetate as 338 substrate, their high abundance in the organic-rich mud facies suggests high availability of 339 acetate. Methanosaetaceae and Methanosarcinacae are frequently detected methanogens in 340 thermokarst lakes (Kallistova et al., 2020).

In an examination of bulk SOM composition using pyrolysis GC-MS, Heslop et al. (2017) found that the organic-rich mud facies had both greater substrate availability (higher C and N concentrations) and greater proportions of compounds associated with allochthonous terrestrial materials (alkanes, alkenes, lignin products, and phenols and phenolic precursors) compared to the underlying sediments. The abundance of these compounds was positively

346	correlated with CH ₄ production potentials measured in 3 °C incubations. Based on ¹⁴ C ages of
347	picked macrofossils from the top 214 cm of the lake core $(170 - 429 \text{ years BP}; \text{Heslop et al.},$
348	2015), the organic carbon (OC) in these deposited facies is far younger than in the Pleistocene-
349	aged permafrost studied in the nearby Vault Creek Permafrost Tunnel (>20,000 years BP;
350	Schirrmeister et al., 2016), which also underlies Vault Lake. Both ¹⁴ C ages of <i>in situ</i> dissolved C-
351	CO ₂ and C-CH ₄ (Elder et al., 2018) and laboratory culture experiments (Douglas et al., 2020)
352	suggest CH4 is produced faster from younger C than older C in northern lakes. However, we note
353	Elder et al. (2018) focused on the ¹⁴ C ages of dissolved C-CO ₂ and C-CH ₄ , which escapes lakes
354	by diffusion. In most arctic lakes, diffusion is a more minor component of total lake CH4
355	emissions (~5% of total annual emissions; Sepulveda-Jauregui et al., 2015).
356	The radiocarbon age of CH4 in naturally-emitted ebullition bubbles, which are the
357	dominant form of CH ₄ emissions in northern lakes (Bastviken et al., 2011; Sepulveda-Jauregui et
358	al., 2015; Walter et al., 2006; Wik et al., 2016), is always 14 C-depleted and older than CH ₄
359	dissolved in the water column (Walter Anthony et al., 2018). At Vault Lake, this was supported
360	by the highest dissolved pore water CH4 concentrations being at the talik thaw boundary, in the
361	recently-thawed taberite facies (Winkel et al., 2019). The sediments directly above the thaw front
362	also had high CH ₄ production potentials in incubation studies of the Vault Lake core (Heslop et
363	al., 2015), and above-median CH ₄ ebullition rates observed at Vault Lake compared to other
364	yedoma lakes (Fig. 4) suggest high CH ₄ production rates and subsequent gas build up at the thaw
365	front. The $^{14}\text{C-CH}_4$ age (up to 28,500 ±140 yrs BP) and δD (-383 to -413‰) of ebullition bubbles
366	emitted from Vault Lake reflect Pleistocene-yedoma carbon and hydrogen (from Pleistocene ice
367	wedges) as the atomic components of the CH4 molecules, as opposed to more modern
368	(Holocene) sources which have younger/heavier CH ₄ isotopic values (Brosius et al., 2012). This

369 agrees with ebullition bubble mapping (Lindgren et al. 2016, 2019; Walter Anthony et al. 2016; 370 Walter Anthony et al., *submitted*) and modeling studies (Kessler et al., 2012) of thermokarst 371 lakes that suggest recently-thawed former permafrost at and directly above the thaw boundary is 372 a source of high CH₄ production. The high CH₄ production at the thaw boundary may be due to 373 high bioavailable substrate potential. The idea that permafrost dissolved OC (DOC) is more 374 biodegradable than younger DOC is supported by a metanalysis showing higher observed DOC 375 biodegradability with increasing permafrost landscape extent (Vonk et al., 2015). In terrestrial 376 settings, DOC from recently thawed former permafrost has been shown to have higher 377 degradation potentials than DOC from the overlying seasonally-thawed active layer (Heslop et 378 al., 2019b; Liu et al., 2019; Selvam et al. 2017). Yedoma permafrost sediments have additionally 379 been shown to have high acetate concentrations (Drake et al., 2015; Ewing et al., 2015), which 380 can be directly utilized as substrate for acetoclastic methanogenesis following thaw. While 381 acetate concentrations were not directly measured in the Vault Lake core, higher relative 382 abundance of the acetoclastic methanogens *Methanosaeta* near the thaw front suggest higher 383 acetate substrate potential. Like the organic-rich mud facies, archaeal communities in the 384 recently-thawed taberite facies were dominated (89% of detected sequences) by the acetoclastic 385 and hydrogenotrophic methanogens Methanosaetaceae, Methanosarcinacae, and 386 Methanoregulaceae (Heslop et al., 2019a; Winkel et al., 2019). 387 Former permafrost sediments that have thawed *in situ* beneath Vault Lake over century 388 timescales (152-508 cm core depth) had the lowest CH₄ production potentials indicated by both 389 incubation studies (Heslop et al., 2015) and dissolved CH₄ concentrations (Winkel et al., 2019). 390 The lacustrine silt and taberite facies, which represented 60% of the total Vault Lake core length, 391 only accounted for 21% of the whole-column CH₄ production potential in the 3 °C incubation

392 study (Heslop et al., 2015). Methanogens were only detected in the rare biosphere, which we 393 defined as taxa below 0.1% relative sequence abundance of all detected archaea in our 16S RNA 394 analyses. Detected methanogens included Methanocellales, Methanoregulaceae, 395 Methanosarcinacae, and Methanosaetaceae (Heslop et al., 2019a; Winkel et al., 2019). Archaeal 396 communities in this region were dominated by CH_4 -consuming methanotrophs (see Section 5). 397 Using temperature sensitivity incubations, Heslop et al. (2019a) showed that these 398 deeper, thawed *in situ* sediments had higher temperature sensitivities for CH₄ production 399 compared to the overlying reworked and deposited material in the organic-rich mud facies. CH4 400 production potentials did not increase with warming at incubation temperatures consistent with 401 temperatures measured in the talik (0 °C and 3 °C) but increased with higher incubation 402 temperatures (10 °C to 25 °C; Heslop et al., 2019a). This suggests CH₄ production in lacustrine 403 silt and taberite facies is substrate-limited, and the input of additional energy with higher 404 temperatures may increase CH₄ production potentials. In the lacustrine silt and taberite facies, we 405 hypothesize the most labile "fast carbon pool" (days to years turnover time; < 5% of permafrost 406 OC; Schädel et al., 2014) was previously exhausted. Per kinetic theory, the remaining, more 407 complex OC (turnover times of decades to centuries) require higher activation energies for 408 microbial decomposition due to the higher number of enzymatic steps needed for biological 409 degradation (Davidson and Janssens, 2006, Conant et al., 2011). Hence, higher temperatures are 410 needed to decompose the more recalcitrant sediment organic C. In contrast, the most deeply 411 thawed sediments located at the talik boundary (i.e. the permafrost thaw front) showed the 412 highest temperature sensitivity at a much lower temperature regime. The large increase in CH₄ 413 production potentials when temperatures increased from 0 to 3 °C points to the higher

bioavailability of the recently thawed, more labile permafrost organic C at the thaw front. Wefurther discuss this in the context of changing climate in Section 7.

416 **5. Methane oxidation**

417 Methane production estimated from anaerobic laboratory incubations of the Vault Lake sediment

418 core (2,819 g CH₄ m⁻² at 3 °C ; Heslop et al., 2015) greatly exceeds the magnitude of CH₄

419 emitted annually by field flux measurements at the same lake (40.9 g CH₄ m⁻²; Sepulveda-

420 Jauregui et al., 2015). Using these values to calculate a first-order estimate suggests up to 99% of

421 produced CH₄ at Vault Lake either accumulates and is held in sediments (presumably released

422 during rare events not captured in field measurements) or is oxidized before being emitted to the

423 atmosphere. A third possibility is that incubation conditions in closed vessels led to inflated CH₄

424 production potentials (i.e. the bottle effect; Ionescu et al., 2015) compared to actual CH₄

425 production, which is unknown; however, the 3 °C temperature of our incubation was slightly

426 lower than the mean annual temperature of surface lake sediments $(3.47 \pm 3.07 \text{ °C} \text{ at the } 0.5 \text{ and}$

427 1.0 m depths), where most of the CH₄ was produced. This suggests that the actual rates in these

428 surface lake sediments could be higher than our estimate. We acknowledge that our CH₄

429 production estimate does not take into account spatial and temporal heterogeneity affecting *in*

430 *situ* CH₄ production and emissions at Vault Lake. Modelled CH₄ production at other Alaskan

431 thermokarst lakes also indicate significantly more CH₄ is produced in sediments than is emitted

432 into the atmosphere, with one study finding CH₄ production in a modeled thermokarst lake talik

433 was up to 10 times higher than observed emissions in the field (Kessler et al., 2012).

Both globally and in the Arctic, CH₄ oxidation plays an important mitigating role in
preventing a significant portion of CH₄ produced in lakes and wetlands from entering the
atmosphere (Oh et al., 2020; Osudar et al., 2016; Reeburgh 2006). Calculations along the length

437	of the Vault Lake core using stable isotopes suggest 41-83% of CH ₄ produced in the core is
438	anaerobically oxidized in situ (Winkel et al., 2019). Under anaerobic conditions, CH ₄ oxidation
439	often occurs as the result of syntrophic relationships between methanotrophic Archaea (ANME)
440	and bacteria that reduce inorganic electron acceptors (Knittel and Boetius, 2009). However,
441	recent research also showed that ANME can perform the process without a bacterial partner (Cai
442	et al., 2018), the existence of anaerobic bacteria that internally produce oxygen to oxidize CH ₄
443	(Ettwig et al., 2010), and that aerobic methanotrophs that can perform CH ₄ oxidation under
444	anaerobic conditions (Oswald et al., 2016). In the Vault Lake hypolimnion and sediments, which
445	are under anoxic conditions year-round (Fig. 2b), anaerobic oxidation of methane (AOM) can
446	occur using sulfate, nitrogen oxidizes, organic matter, chlorite, or metals (e.g. iron and
447	manganese) as terminal electron acceptors. AOM values reported in the Vault Lake core by
448	Winkel et al. (2019) ranged from 0 to 2.87 nmol CH ₄ cm ⁻³ d ⁻¹ . For comparison, reported AOM
449	rates in global freshwater lake systems (n = 66) were 0.01 to 100 nmol CH ₄ cm ⁻³ d ⁻¹ (Martinez-
450	Cruz et al., 2018). The lower values reported by Winkel et al. (2019) compared to global
451	freshwater AOM rates are likely due to a combination of: (1) lower organic matter quality in the
452	taberite facies, which had been thawed beneath Vault Lake for centuries, compared to in near-
453	surface deposited sediments, which are where sediment AOM is typically quantified; and (2) the
454	incubations conducted by Winkel et al. 2019 being done at 4 °C, which is a lower temperature
455	than most of the incubations reported in literature.
456	To our knowledge, the study by Winkel et al. (2019) is the first to identify and quantify
457	AOM in cold thawing and former permafrost sediments beneath a thermokarst lake. Tracer

458 incubation studies of AOM potentials found low AOM potential rates in the transitional

459 permafrost (~0.9 pmol cm⁻³ d⁻¹), but high (~660 μ M) concentrations of nitrite (Winkel et al.,

460 2019). Studies from another lake core collected in Interior Alaska suggest yedoma sediments 461 also have high concentrations of Fe, which may serve as an alternate electron acceptor for AOM 462 following thaw (Winkel, *unpublished data*). Evidence from the Vault Lake core suggests former 463 permafrost sediments that have been thawed in situ for centuries (311 to 532 cm depth; taberite 464 and recently-thawed taberite facies) are a region of high AOM in thermokarst lake environments. 465 Compared to the transitional permafrost, AOM rates doubled in the overlying recently-thawed 466 taberite facies. This region of the Vault Lake core had the highest AOM potential rates (up to 2.88 pmol cm⁻³ d⁻¹; mean 1.7 ± 0.7 pmol cm⁻³ d⁻¹), high nitrite concentrations, and the most 467 468 fractionated δ^{13} C-CH₄ in the pore waters (Winkel et al., 2019). Archaea in this region were 469 almost exclusively the nitrate-driven AOM (Haroon et al 2013) or iron-driven AOM (Ettwig et al 470 2015, Cai et al 2018) archaea Methanoperedenaceae (Winkel et al., 2019). Sulfate 471 concentrations in the taberite (up to 35 ± 2 mM) were also found to be sufficient for potentially 472 supporting AOM (Winkel et al., 2019). 473 A second region of high AOM was observed in the deposited near-surface sediments of 474 Vault Lake. Winkel et al. (2019) found the top 100 cm of the Vault Lake core to have high AOM potential rates (up to 2.3 pmol cm⁻³ d⁻¹ at 3 °C). Mean AOM potential rates separately measured 475

476 by Martinez-Cruz et al. (2017) in sediments collected from the top 25 cm were 12.27 ± 3.96

477 nmol cm⁻³ d⁻¹ (mean \pm SD) at 4 °C. Compared to net CH₄ production potential rates measured

478 from the same 25 cm core at the same incubation temperature, this represents approximately one-

479 third of CH₄ produced in surface sediments at Vault Lake being anaerobically oxidized

480 (Martinez-Cruz et al., 2017). AOM in this region used different electron acceptors and had

481 different methanotrophs compared to the high AOM in the underlying taberite. This AOM was

482 mainly attributed to type I aerobic methanotrophs belonging to the genus *Methylobacter*; only

483 iron and manganese were found to be present in sufficient quantities to support the observed484 AOM (Martinez-Cruz et al., 2017).

485 Synthesizing published rates of CH_4 oxidation from Alaskan yedoma (n = 38) 486 observations) and non-vedoma (n = 53 observations) lakes revealed that lakes located in vedoma 487 regions did not have significantly different oxidation rates (aerobic or anaerobic) in their 488 sediment than non-yedoma northern lakes (Fig. 6). While we did not examine CH₄ oxidation in 489 the water column at Vault Lake, research from nearby Alaskan thermokarst lakes and other arctic 490 lakes suggests further CH₄ oxidation occurs in the water column and under ice under both 491 aerobic and anaerobic conditions. A study by Martinez-Cruz et al. (2015) in yedoma-type lakes near Vault Lake found mean potential aerobic CH₄ oxidation rates of 0.71 mg L⁻¹ d⁻¹ in the 492 summer and 0.18 mg L⁻¹ d⁻¹ in the winter. Our literature synthesis suggests median potential CH₄ 493 494 oxidation rates in the water columns of yedoma lakes are higher than in non-yedoma lakes (Fig. 495 6). A recent study by Thalasso et al. (2020) found that, in a stratified lake in north-central 496 Siberia, all CH₄ produced in the sediments and deep hypolimnion was removed below the 497 oxycline, suggestive of high AOM. This case study suggests, in northern lakes that stratify, it is 498 possible for most or all CH₄ produced in sediments to be anaerobically oxidized in the 499 hypolimnion, resulting in little to no CH₄ exchange with the epilimnion and/or atmosphere. In 500 northern Alaska, it has been suggested that the threshold depth for lakes to fully stratify is 501 approximately 4 m (Arp et al., 2015). At Goldstream Lake, a thermokarst lake located in interior 502 Alaska, approximately 50% of CH₄ in ebullition bubbles trapped under winter ice cover oxidize 503 prior to ice out (Elder et al., 2019). Greene et al. (2014) estimated the under-ice CH₄ oxidation at the same lake to be 56%. ¹³C-CH₄ analyses of ice cores from thermokarst lake and lagoon 504 505 environments underlain by continuous permafrost in Siberia revealed that CH₄ at the ice-water

506 interface was oxidized to concentrations near atmospheric equilibrium, while the overlying ice

507 was supersaturated in CH₄ compared to atmospheric concentrations (Spangenberg et al., *in*

508 review).



509

Figure 6. Rates of CH₄ oxidation from Alaskan yedoma (n = 38 observations) and non-yedoma (n = 53 observations) lakes. Significant (two-sample t-test; p < 0.05) differences between yedoma and non-yedoma lakes are denoted. Sediment AOM at Vault Lake, measured by Martinez-Cruz et al. (2017) and Winkel et al. (2019), are highlighted in red. Data are reported by de Jong et al. (2018), He et al. (2012), Martinez-Cruz et al. (2015, 2017), Miller et al. (2019), and Winkel et al. (2019).

516

517 6. Carbon dioxide to methane production ratios

- 518 One issue of great importance in determining the potential impact of C emissions from
- 519 thermokarst lakes on the global C cycle is determining proportions of C released as CO₂ versus
- 520 CH₄. Northern lakes emit more CO₂ than CH₄; on average 26x more CO₂ was emitted than CH₄
- 521 in a transect of 40 lakes in Alaska (Fig. 7; Sepulveda-Jauregui et al., 2015). Several processes in
- 522 both the lake sediments and overlying water column affect the magnitudes of CO₂ versus CH₄

523	emissions. Analysis of data from $5,118$ boreal lakes found that CO_2 emissions were
524	predominately (60% of studied lakes) sustained by loading from the catchment (e.g. inorganic C
525	loading from surface and ground waters), as opposed to internal CO ₂ production (Weyhenmeyer
526	et al., 2015). In the aerobic portion of the water and sediment column, internal processes that
527	produce CO_2 include: biological respiration in the water column (3.4–91% of internal CO_2
528	production in boreal lakes; Weyhenmeyer et al., 2015) and sediments (4.0-60% internal CO_2
529	production); inorganic C fixation; and, in shallow non-yedoma lakes (Tan et al., 2017), light
530	degradation (3.0-48% internal CO ₂ production). In anaerobic portions of the water column and in
531	lake sediments, CO ₂ is produced both as a by-product of methanogenesis and fermentation, and
532	as an end-product of CH ₄ oxidation; CH ₄ is produced as an end-product of methanogenesis. In
533	both the sediments and water column, under both aerobic and anaerobic conditions, a significant
534	amount of CH ₄ is removed and CO ₂ is produced during CH ₄ oxidation.

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535

536 Figure 7. In situ ratios of CO₂:CH₄ emissions measured in Alaskan lakes vs. CO₂:CH₄ porewater 537 concentrations in Vault Lake sediments. Annual CO2 and CH4 emissions were reported by Sepulveda-Jauregui et al. (2015). Emission CO₂:CH₄ ratios were not significantly different in 538 539 yedoma versus non-yedoma lakes (two-sample t-test; p = 0.51). Pore water dissolved gas concentrations in anaerobic lake sediment, measured in the Vault Lake core, were reported by 540 541 Winkel et al. (2019). 542 543 It had been previously thought that, under stable anaerobic CH₄-producing conditions, the 544 lowest possible ratio of CO₂:CH₄ production is 1:1 (Conrad, 1999). However, some recent 545 studies from Arctic environments have reported CO₂:CH₄ ratios of less than 1, indicating more 546 CH₄ production than CO₂ production (Heslop et al., 2019b; Knoblauch et al., 2018; Lokshina et 547 al., 2019; Walter Anthony et al., 2014). At a eutrophic fen site in Siberia, the higher CH₄ 548 production than CO₂ production was attributed to proportions of CH₄ produced through 549 acetoclastic (Equation 2) versus hydrogenotrophic (Equation 1) methanogenesis; adding an 550 inhibitor to change the major pathway from acetoclastic to hydrogenotrophic methanogenesis 551 increased CO₂:CH₄ ratios from 0.3-0.4 to 0.7-0.9 (Lokshina et al., 2019). 552 The notion that the CO₂:CH₄ production ratio can be less than 1:1 during anaerobic C 553 mineralization is also supported by field measurements and modeling. Due to the diffusion of 554 gases through meters of dense, silt-dominated sediments in yedoma lake taliks being 555 prohibitively slow, yedoma talik sediments favor ebullition through secondary pore channels as a gas escape mechanism (Tan et al., 2015). ¹⁴C-CH₄ measured in ebullition bubbles is older than 556 557 14 C-CO₂ from the same bubbles (Walter Anthony et al. 2018), suggesting that the CH₄ in 558 ebullition bubbles comes from older, deeper sediment and the CO₂ comes from younger 559 sediments closer to the sediment-water interface. The concept that CH₄ production dominates in 560 deep taliks and that the dissolved gases giving rise to diffusive fluxes originate from surface lake sediments is further supported by ¹⁴C-CH₄ ages of dissolved CH₄ being much younger than ¹⁴C-561 562 CH₄ in ebullition bubbles (Elder et al., 2019). Ebullition bubbles released from deep with taliks

563	are typically $> 85\%$ CH ₄ and $< 2\%$ CO ₂ (Walter et al. 2006; Walter Anthony et al., <i>submitted</i>),
564	which may indicate that the CO ₂ :CH ₄ production rates are skewed towards greater CH ₄ than CO ₂
565	production. In order for CH_4 to be produced in excess of CO_2 , an inorganic source of H_2 is
566	required. Telling et al. (2015) showed that pulverization of silicate rocks beneath glaciers
567	produces abiogenic H ₂ ; future research should explore the possibility that pulverization of rocks
568	generating wind-blown silt could have also led to abiogenic H ₂ in syngenetically aggraded
569	yedoma permafrost sediments, enabling higher CH4 production in deep yedoma talik sediments.
570	CO ₂ :CH ₄ ratios of pore water dissolved gas measured in the Vault Lake core ranged from
571	1.2 to 10.7 (mean 4.6; Fig. 8; Winkel et al., 2019). The lowest CO ₂ :CH ₄ ratio (1.2; 557 cm depth)
572	was found at the thaw boundary, a region which also had higher relative abundance of the
573	acetoclastic methanogens Methanosaeta (Winkel et al., 2019). These results would be consistent
574	with the suggestion that acetoclastic methanogenesis leads to lower CO ₂ :CH ₄ production ratios.
575	However, the α_C values (an indicator of CH ₄ production pathway; $\alpha_C > 1.06$ indicates CO ₂
576	reduction; $\alpha_C < 1.04$ indicates acetate fermentation; Walter et al., 2008) in both the dissolved gas
577	and ebullition bubbles indicate that CH ₄ in the deepest part of the talik is produced primarily by
578	CO ₂ reduction (Fig. 8d). The α_C of dissolved gas in the pore water ranged from 1.044 to 1.087
579	(mean \pm SD 1.066 \pm 0.009; n = 38) and in bubbles released from the base of the talik in
580	boreholes was 1.066 \pm 0.005 (mean \pm SE; n = 5); α_C in hotspot ebullition seeps, which are
581	thought to be the deepest-seated naturally occurring ebullition seeps in taliks was 1.054 ± 0.006
582	(n = 5); and, α_C in ebullition seeps originating from shallower sediments was 1.048 \pm 0.005 (n =
583	8).





Figure 8. Magnitudes of AOM potentials (a; Winkel et al., 2019) and CH₄ production potentials
(b; Heslop et al., 2015) measured in incubations of sediments from the Vault Lake core.
Measurements of CO₂ and CH₄ concentrations in sediment pore waters from the Vault Lake core



590 The highest CO₂:CH₄ ratios in porewater dissolved gases (8.6-10.7) were found in the

surface organic rich mud (18-38 cm depth) and in the taberite (296-306 cm depth) facies. This is

592 consistent with regions in the core found to have higher rates of AOM, which would decrease

- 593 CH₄ concentrations and increase CO₂ concentrations. δ^{13} C-CO₂ becomes increasingly depleted
- as CH₄ is oxidized to CO₂. Hence, enriched values of δ^{13} C-CO₂ indicate very little CH₄
- 595 oxidation, while depleted values indicate more CH₄ oxidation. With the exception of one

borehole bubble gas sample, in which δ^{13} C-CO₂ was -20.7 ‰ and the corresponding δ^{13} C-CH₄ was -66.2 ‰, borehole bubbles at Vault Lake had enriched ‰ values (δ^{13} C-CO₂ -5.1 to -15.4 ‰; δ^{13} C-CH₄ -74.2 to -75.5 ‰; Winkel et al., 2019). This indicates CH₄ produced at the base of the talik is subject to less AOM, which is supported by the low CO₂:CH₄ porewater ratios at this depth. In contrast, bubbles occurring from natural ebullition seeps draining free-phase gas from shallower sediments had more depleted δ^{13} C-CO₂ (-14.0 to -26.2 ‰) and more enriched δ^{13} C-CH₄ (-63.0 to -75.5 ‰), suggesting these bubbles were formed from pore space gas that was

603 more influenced by AOM.

604 7. Potential changes with climate warming

605 Thermokarst lake formation alters local thermal dynamics, leading to lateral and vertical thaw 606 and erosion, the formation of taliks (Brewer, 1958), and mobilization C previously sequestered in 607 permafrost (Fig. 9). A process-based climate-sensitive lake biogeochemical model by Tan et al. 608 (2015) indicated that seasonal variation in CH₄ emissions can be explained by energy input. In 609 mesocosm warming experiments conducted in a shallow pond in the Netherlands, warming by 4 610 °C increased ebullitive CH₄ flux by 51%, but did not significantly affect diffusive CH₄ flux 611 (Aben et al., 2017). The combination of increased energy input due to warmer temperatures and 612 increased substrate potential due to additional permafrost thaw suggest additional climate 613 warming will increase CH₄ fluxes and production in thermokarst lakes such as Vault Lake.



614

Figure 9. Current and future CH4 dynamics in a thermokarst lake environment. Climate warming 615 is expected to lead to increased lateral and vertical thaw and erosion, which may lead to lake 616 expansion or lake drainage (see section 8 below). In addition, warmer temperatures are expected 617 618 to lead to increased CH₄ production and oxidation and increased ebullitive and diffusive flux. 619 Fewer days under ice cover are anticipated to reduce ice storage bubble flux and under-ice CH₄ 620 oxidation. The net effects on annual CH₄ flux remain uncertain, although literature suggests 621 these combined factors will lead to increases in CH₄ flux from thermokarst lakes with additional 622 warming. See Section 8 for additional discussion of uncertainties. 623



629 methanogens and CH₄ production rates (Knoblauch et al., 2018, Wei et al., 2018). Acetoclastic 630 methanogens, which were in higher relative abundance near the thaw front, have lower 631 temperature optima than hydrogenotrophic methanogens (Allen et al., 2014) and therefore may 632 be more suited to metabolic processes at these colder temperatures. While incubated sediments 633 from near the thaw boundary had high temperature sensitivities at lower incubation temperatures, 634 the sediments did not produce significantly more CH₄ at higher incubation temperatures (10 °C 635 and 25 °C; Heslop et al., 2019a). This suggests current CH₄-producing microbial communities 636 near the thaw boundary are not substrate-limited and optimized for colder temperatures. 637 While increased energy input from permafrost thaw increases CH₄ production at lower 638 temperatures, at temperatures consistent with thermokarst lake talks (4 °C) CH₄ production is 639 only at 17% of maximum activity (Metje and Frenzel, 2007). Laboratory incubation studies 640 suggest the optimal temperatures for methanogenesis are between 26-28 °C (Metje and Frenzel, 641 2007). This implies additional warming will cause a second increase in CH₄ production. In 642 sediments of the Vault Lake core estimated to have been thawed centuries (100 to 400 years), 643 CH₄ production did not significantly increase with warming at lower incubation temperatures 644 $(0 \,^{\circ}C \text{ to } 3 \,^{\circ}C)$ but increased when sediments were warmed to temperatures above those observed 645 in situ (10 °C and 25 °C; Heslop et al., 2019a). This implies that the former permafrost thawed 646 for centuries, which currently has low CH₄ production potentials (Heslop et al., 2015; Kessler et 647 al., 2012) due to low substrate bioavailability (Heslop et al., 2017) and/or the establishment of 648 AOM microbial communities (Winkel et al., 2019), has the potential to support additional CH_4 649 production following additional warming. This increased CH₄ production potential may be due 650 to increased substrate bioavailability due to higher ambient energy in warmer temperatures, a 651 shift in microbial communities from predominately methanotrophs to predominately

methanogens, or a combination thereof. In a landscape context, such additional warming of these
sediments may occur with warmer atmospheric temperatures and/or further thermokarst or
thermo-erosion exposing and mobilizing these sediments in the landscape.

655 While warming will increase CH₄ production rates in lake sediments, we expect the 656 increased production may be partially offset by increased CH₄ consumption rates (Oh et al., 657 2020). However, in a nearby yedoma thermokarst lake in Interior Alaska, higher Q₁₀ values for 658 methanogenesis ($Q_{10} = 8.5 \pm 0.1$) versus aerobic CH₄ oxidation ($Q_{10} = 2.7 \pm 0.3$) in the water 659 column suggest increases in CH₄ production will outpace increases in CH₄ consumption 660 (Sepulveda-Jauregui et al., 2018). Further, it has been suggested that the shortening of ice cover periods in thermokarst lakes due to warming and increasing snow thickness (Arp et al. 2018) 661 662 leads to more CH₄ being emitted into the atmosphere via direct ebullition instead of being 663 oxidized while trapped under ice (Elder et al., 2019; Greene et al., 2014). In Finnish boreal lakes, 664 more ice-free days have also been linked to greater diffusive CH₄ flux (Guo et al., 2020). 665 Additional research is necessary to determine how these different controls will affect net CH₄ 666 emissions with warming in thermokarst and northern lakes.

667 8. Future research directions

668 *Thermokarst lake expansion versus drainage.*

Ongoing efforts to constrain the diverse impacts of climate change and permafrost thaw on thermokarst lake expansion or drainage are critical for determining future CH₄ release potentials, as the landscape wetting (lake expansion) versus drying (lake drainage) will influence whether permafrost C is thawed under aerobic or anaerobic conditions (Lawrence et al., 2015). While in the Interior Alaska region near Vault Lake, net thermokarst lake area has increased by ~43% since 1949 (Walter Anthony et al., *submitted*), an examination of continental-scale Landsat 675 imagery transects in North America and Eurasia found northern net lake area decreased 1.45% 676 between 1999 and 2014 (Nitze et al., 2018). Despite net lake area decreases, thermokarst lake 677 expansion by lateral shore erosion in Arctic-Boreal lowlands with ice-rich permafrost continues 678 to cause significant gross lake area growth (transformation of land area to lake area; (Nitze et al., 679 2018). This in turn leads to growth of taliks and thawed sediment volumes under lakes (Walter 680 Anthony et al., 2018). Mobilized permafrost C in taliks would be processed in anaerobic 681 environments, meaning it is released as both CH_4 and CO_2 . On the other hand, lake drainage 682 changes local thermal dynamics such that: (1) taliks formed beneath a lake can refreeze and (2) 683 near-surface C is more likely to be mineralized under aerobic conditions as CO₂. 684 *Physical entrapment of CH*₄. 685 Our first-order calculations suggest up to 99% of CH₄ produced at Vault Lake is not emitted into 686 the atmosphere. Coupled with our estimates that 32-83% of produced CH4 is oxidized in situ 687 (Martinez-Cruz et al. 2017; Winkel et al., 2019), this suggests that 16-67% of produced CH₄ 688 becomes entrapped within the Vault Lake system. While we acknowledge these values are rough 689 first order estimates, they suggest periodic physical entrapment plays a substantial role in 690 thermokarst lake CH₄ budgets. The location and fate of this entrapped CH₄ is not well 691 constrained; however, rapid and sustained ebullition (for at least one day) observed in boreholes 692 distributed across Vault Lake drilled in 2013 to map the talik dimensions suggests a large 693 volume of gas was entrapped in sediments. Voids observed in freshly split sediment cores 694 collected from thermokarst lakes, including Vault Lake (Heslop, unpublished data Walter 695 Anthony, *unpublished data*), are suggestive of gas pockets storing CH₄ within the sediment. 696 While to our knowledge these gas bubbles have not been quantified in thermokarst lake 697 sediments, laboratory incubations examining gas bubble formation in wetland sediments suggest

698 gas-filled voids, which are a precursor to the formation of ebullition bubble conduits, store more 699 gas in clay and silt sediments (maximum 18.8% volumetric gas content) than sandy and silty 700 sediments (maximum 13.2% and 12.0% volumetric gas content, respectively; Liu et al., 2016). In 701 these laboratory incubations, hydrostatic head drop was linearly correlated with the magnitude of 702 ebullition episodes releasing stored gas (Liu et al., 2016).

703 Depending on the phase and location of the entrapped gas, it may or may not be subject 704 to CH₄ oxidation. Only the dissolved phase of CH₄ is subject to oxidation, so if gases are stored 705 as bubbles in sediments, then the bubble phase protects CH₄ from *in situ* oxidation. For instance, 706 the δ^{13} C-CH₄ values of bubbles released from the borehole drilling at Vault Lake (mean -73%), n 707 = 5 samples) suggests that this trapped gas was not enriched by oxidation. Additional, high-708 resolution analyses in multiple thermokarst lakes to determine if and where entrapped CH₄ is 709 subject to additional oxidation and what physical factors determine if and how it can be released 710 will help better constrain the role of CH₄ emissions from thermokarst lakes in the PCF.

711 Incorporating thermokarst C dynamics into models

712 First-order models indicate up to 20% of the permafrost region is anticipated to 713 experience physical disturbance thaw processes by 2300 (Turetsky et al., 2020), with GHG 714 emissions from thermokarst lakes having the potential to more than double the PCF this century 715 (Schneider von Deimling et al. 2015, Walter Anthony et al. 2018). Yet, large-scale global models 716 predicting the magnitude of the PCF currently only consider gradual thaw processes (e.g. active 717 layer deepening) and do not incorporate potential C emissions from rapid thaw processes, 718 including thermokarst lakes (Turetsky et al., 2020). Improved understanding CH₄ dynamics 719 within thermokarst lake environments will aid in incorporating their GHG emissions into future 720 estimates of the PCF. While there have been studies modelling the formation of taliks beneath

721 water bodies and subsequent potential C emissions under different climate scenarios (Kessler et 722 al., 2012; Langer et al., 2016; Schneider von Deimling et al., 2015) and a process-based climate-723 sensitive lake biogeochemical model predicting CH₄ emissions (Tan et al., 2015), additional 724 research is needed to incorporate these models into larger scale permafrost thaw and C dynamic 725 models. Our work at Vault Lake suggests that the sediment layers of the talik experience 726 different rates of methanogenesis and AOM due to varying substrate potentials and microbial 727 communities, supporting the suggestion by Tan et al. (2015) that variability of CH₄ emissions 728 can be primarily explained by energy input and substrate availability. In turn, this led to the 729 facies experiencing different responses to additional warming. Integrating this depth-dependent 730 variability in thermokarst lake GHG production dynamics into thermokarst models, including 731 spatially dynamic estimates of CO₂:CH₄ production ratios, CH₄ oxidation, and gas release versus 732 entrapment, is a first step in refining our estimates of how rapid thaw processes will contribute to 733 the PCF. Given the variation we see between sediment layers within the same core at one 734 thermokarst lake, it is necessary to examine additional thermokarst lake cores from diverse 735 Arctic environments to determine how permafrost thaw history, sediment background, and OM 736 and microbial characteristics influence CH₄ dynamics. Examining spatial variability within the 737 same lake, including variation in facies thicknesses, talik depth, microbial communities, and 738 substrate potential, will further refine estimates of potential GHG emissions from thermokarst 739 lakes. More drilling studies capturing entire talik profiles are necessary to determine if patterns 740 observed at Vault Lake hold true for other thermokarst lakes, including thermokarst lakes in non-741 yedoma regions.

In additional to additional field sampling to constrain variability, we suggest one
approach to accounting for this spatial and temporal heterogeneity in models is to adopt a

microscale approach, where variabilities in substrate availability, the microbiome, and energy are incorporated into geochemical rate models (e.g. Neumann et al., 2016) to predict CH₄ production and consumption rates in each facies. The predicted GHG production and consumption rates can then be integrated into whole-lake models that take into account thaw beneath lakes (e.g. Kessler et al., 2012; Langer et al., 2016; Schneider von Deimling et al., 2015), and eventually into larger models predicting the impact of thermokarst lakes on the PCF (e.g. Turetsky et al., 2020; Walter Anthony et al., 2018).

751 9. Summary and conclusions

752 At Vault Lake in Interior Alaska, a typical recently-formed thermokarst lake in a lowland 753 permafrost setting with yedoma deposits, we observed depth-dependent variations in CH₄ 754 production and oxidation. Variations in CH₄ production and oxidation potentials corresponded 755 with changes in CO₂:CH₄ ratios and net CH₄ production response to temperature increases. The 756 top 150 cm of the Vault Lake core had the highest CH₄ production potentials in 3 °C incubations, 757 but oxidation potentials suggest 41-83% of the produced CH₄ is anaerobically oxidized in situ. 758 This region also had higher CO_2 : CH₄ ratios compared to the remainder of the core. The 759 sediments near the thaw boundary, which prior modelling and ebullition bubble mapping studies 760 of thermokarst lakes suggested are an important source of CH₄ production, had the highest *in situ* 761 CH₄ concentrations and high CH₄ production potentials in incubations. CH₄ production near the 762 thaw boundary was sensitive to warming at temperatures consistent with in situ temperatures in 763 the talik and did not respond to additional warming. Further, the lowest CO₂:CH₄ ratio was 764 observed near the thaw boundary, potentially due to higher proportions of CH₄ being produced 765 through acetoclastic versus hydrogenotrophic methanogenesis. Former permafrost sediments, 766 thawed centuries ago, had high CO₂:CH₄ ratios that were potentially influenced by high AOM

rates. These sediments also experienced significant increases in net CH₄ production when
warmed to temperatures above those observed *in situ*.

769	While not studied at Vault Lake, additional processes within the water column such as
770	CH4 oxidation and heterotrophic C respiration affect CO2:CH4 ratios and how much CH4 is
771	emitted into the atmosphere. The form of CH4 in the thermokarst lake system (bubbles versus
772	dissolved) is a key control on whether CH4 is subject to <i>in situ</i> oxidation. Therefore, more
773	research is necessary to better constrain how processes such as CH ₄ storage within the system,
774	including storage in sediments and under ice cover, affect lake CH4 dynamics and emissions.
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791 References

- Aben, R.C.H., Barros, N., van Donk, E. et al., 2017. Cross continental increase in methane
 ebullition under climate change. Nature Communications, 8, 1682.
- Allen, D.T., Torres, V.M., Thomas, J., et al., 2013. Measurements of methane emissions at
 natural gas production sites in the United States. PNAS, 110 (44), 17768-17773.
- Alvarez, R.A., Zavala-Araiza, D., Lyon, D.R., et al., 2018. Assessment of methane emissions
 from the U.S. oil and gas supply chain. Science, 361(6398), 186-188.
- Arp, C.D., Jones, B.M., Liljedahl, A.K., et al., 2015. Depth, ice thickness, and ice-out timing
 cause divergent hydrologic responses among Arctic lakes. Water Resources Research,
 51(12), 9379-9401.
- 801 Arp, C.D., Jones, B.M., Engram, M., et al., 2018. Contrasting lake ice responses to winter
- 802 climate indicate future variability and trends on the Alaskan Arctic Coastal Plain.
 803 Environmental Research Letters, 13(12), 125001.
- Bastviken, D., Tranvik, L.J., Downing, J.A., et al., 2011. Freshwater Methane Emissions Offset
 the Continental Carbon Sink. Science, 331(6013), 50-50.
- Brewer, M.C., 1958. The thermal regime of an Arctic lake. Eos, 39(2), 278-284.
- 807 Brosius, L.S., Walter Anthony, K.M., Grosse, G., et al., 2012. Using the deuterium isotope
- 808 composition of permafrost meltwater to constrain thermokarst lake contributions to
- atmospheric CH₄ during the last deglaciation. Journal of Geophysical Research, 117,
- 810 G01022.

811	Cai, C., Leu, A.O., Xie, G.J., et al., 2018. A methanotrophic archaeon couples anaerobic
812	oxidation of methane to Fe(III) reduction. ISME Journal, 12(8), 1929-1939.
813	Canfield, D.E., Kristensen, E., and Thamdrup, B., 2005. The Methane Cycle. Advances in
814	Marine Biology, 48, 383-418.
815	Conant, R.T., Ryan, M.G., Ågren, G.I., et al., 2011. Temperature and soil organic matter
816	decomposition rates- synthesis of current knowledge and a way forward. Global Change
817	Biology, 17, 3392–3404.
818	Conrad, R., 1999. Contribution of hydrogen to methane production and control of hydrogen
819	concentrations in methanogenic soils and sediments. FEMS Microbiology Ecology, 28,
820	193e202.
821	Coolin, M.J.L., and Orsi, W.D., 2015. The transcriptional response of microbial communities in
822	thawing Alaskan permafrost soils. Frontiers in Microbiology,
823	doi.org/10.3389/fmicb.2015.00197
824	Davidson, E.A., Janssens, I.A., 2006. Temperature sensitivity of soil carbon decomposition and
825	feedbacks to climate change. Nature, 440, 165-173.
826	Dean, J.F., Meisel, O.H., Rosco, M.M., et al., 2020. East Siberian Arctic inland waters emit
827	mostly contemporary carbon. Nature Communications, 11, 1627.
828	de Jong, A.E.E., In 't Zandt, M.H., Meisel, O.H., et al., 2018. Increases in temperature and
829	nutrient availability positively affect methane-cycling microorganisms in Arctic
830	thermokarst lake sediments. Environmental Microbiology, 20(12), 4314-4327.
831	Douglas, P.M.J, Moguel, R.G., Walter Anthony, K.M., et al., 2020. Clumped Isotopes Link
832	Older Carbon Substrates With Slower Rates of Methanogenesis in Northern Lakes.
833	Geophysical Research Letters, 47(6), e2019GL086756.

834	Drake, T.W., Wickland, K.P., Spencer, R.G.M., et al., 2015. Ancient low-molecular-weight
835	organic acids in permafrost fuel rapid carbon dioxide production upon thaw. PNAS,
836	112(45), 13946-13951.

- 837 Elder, C.D, Xu, X., Walker, J., et al., 2018. Greenhouse gas emissions from diverse Arctic
- Alaskan lakes are dominated by young carbon. Nature Climate Change, 8(2), 166-171.
- 839 Elder, C.D., Schweiger, M., Lam, B., et al., 2019. Seasonal sources of whole-lake CH₄ and CO₂
- 840 emissions from Interior Alaskan thermokarst lakes. Journal of Geophysical Research:
 841 Biogeosciences, doi.org/10.1029/2018JG004735.
- 842 Engram, M., Walter Anthony, K.M., Scahs, T., et al., 2020. Remote sensing northern lake
- 843 methane ebullition. Nature Climate Change, 10, 511–517.
- 844 Etminan, M., Myhre, G., Highwood, E.J., et al, 2016. Radiative forcing of carbon dioxide,
- 845 methane, and nitrous oxide: A significant revision of the methane radiative forcing.
- 846 Geophysical Research Letters, 43, 12,614–12,623/
- Ettwig, K.F., Butler, M.K., Le Paslier, D., et al., 2010. Nitrite-driven anaerobic methane
 oxidation by oxygenic bacteria. Nature, 464(7288), 543-548.
- 849 Ewing, S.A., O'Donnell, J.A., Aiken, G.R., et al., 2015. Long-term anoxia and release of ancient,
- labile carbon upon thaw of Pleistocene permafrost. Geophysical Research Letters, 42(24),
 10,730-10,738.
- 852 Greene, S., Walter Anthony, K.M., Archer, D., et al., 2014. Modeling the impediment of
- 853 methane ebullition bubbles by seasonal lake ice. Biogeosciences, 11, 6791–6811.
- 854 Günthel, M., Klawonn, I., Woodhouse, J., et al., 2020. Photosynthesis-driven methane
- production in oxic lake water as an important contributor to methane emission.
- Limnology & Oceanography, doi 10.1002/lno.11557

857	Guo, M., Zhuang, Q., Tan, Z., et al., 2020. Rising methane emissions from boreal lakes due to
858	increasing ice-free days. Environmental Research Letters, doi 10.1088/1748-
859	9326/ab8254.

- Haroon, M.F., Hu, S., Shi, Y., et al., 2013. Anaerobic oxidation of methane coupled to nitrate
 reduction in a novel archaeal lineage. Nature, 500(7464), 567-70.
- Heslop, JK, Walter Anthony, K.M., Sepulveda-Jauregui, A, et al., 2015. Thermokarst lake
 methanogenesis along a complete talik profile. Biogeosciences 12, 4317–4331.
- 864 Heslop, JK, Walter Anthony, K.M., Zhang, M., 2017. Utilizing pyrolysis GC-MS to characterize
- 865 organic matter quality in relation to methane production in a thermokarst lake sediment
 866 core. Organic Geochemistry 103, 43-50.
- Heslop JK, Walter Anthony KM, Grosse G, et al., 2019a. Century-scale time since permafrost
 thaw affects temperature sensitivity of net methane production in thermokarst-lake and
 talik sediments. Science of The Total Environment, 691, 124-134.
- 870 Heslop JK, Winkel M, Walter Anthony KM, et al., 2019b. Increasing organic carbon biolability
- 871 with depth in yedoma permafrost: ramifications for future climate change. Journal of

872 Geophysical Research: Biogeosciences, 124(7), 2021-2038.

- Huttunen, J.T., Alm, J., Liikanen, A., et al., 2003. Fluxes of methane, carbon dioxide and nitrous
 oxide in boreal lakes and potential anthropogenic effects on the aquatic greenhouse gas
 emissions. Chemosphere, 52, 609–621.
- 876 Ionescu, D., Bizic-Ionescu, M., Khalili, A. et al., 2015. A new tool for long-term studies of
- 877 POM-bacteria interactions: overcoming the century-old Bottle Effect. Scientific Reports,
- 8785, 14706.

879	Jansen, E., Christensen, J.H., Dokken, T., et al., 2020. Past perspectives on the present era of
880	abrupt Arctic climate change. Nature Climate Change https://doi.org/10.1038/s41558-
881	020-0860-7

Jones, B.M., Grosse, G., Arp, C.D., et al., 2011. Modern thermokarst lake dynamics in the

continuous permafrost zone, northern Seward Peninsula, Alaska. Journal of Geophysical
Research: Biogeosciences, 116, G00M03.

Jorgenson, M.T., Yoshikawa, K., Kanevskiy, M., et al., 2008. Permafrost characteristics of
Alaska. In: Kane, D., Hinkel, K. (Eds.), Proceedings of the Ninth International

888 Kessler, M.A., Plug, L.J., and Walter Anthony, K.M., 2012. Simulating the decadal- to

millennial-scale dynamics of morphology and sequestered carbon mobilization of two
thermokarst lakes in NW Alaska. Journal of Geophysical Research: Biogeosciences, 117,

Conference on Permafrost. University of Alaska, Fairbanks, AK, pp. 121-122.

891 G00M06.

- Knittel, K., and Boetius, A., 2009. Anaerobic Oxidation of Methane: Progress with an Unknown
 Process. Annu. Rev. Microbiol., 63, 311–334.
- Knoblauch, C., Beer, C., Liebner, S., et al., 2018. Methane production as key to the greenhouse
 gas budget of thawing permafrost. Nature Climate Change, 8, 309–312.
- Kuhn, MK, Lundin, EJ, Giesler, R, et al., 2018. Emissions from thaw ponds largely offset the
 carbon sink of northern permafrost wetlands. Scientific Reports 8, 9535.
- 898 Langer, M., Westermann, S., Boike, J., et al., 2016. Rapid degradation of permafrost underneath
- 899 waterbodies in tundra landscapes—Toward a representation of thermokarst in land
- 900 surface models. Journal of Geophysical Research: Earth Surface, 121(12), 2446-2470.

901	Lawrence, D.M., Koven, C.D, Swenson, S.C., et al., 2015. Permafrost thaw and resulting soil
902	moisture changes regulate projected high-latitude CO ₂ and CH ₄ emissions.
903	Environmental Research Letters.,10, 094011.
904	Lindgren, P.R., Grosse, G., Walter Anthony, K.M., and Meyer, F.J., 2016. Detection and
905	spatiotemporal analysis of methane ebullition on thermokarst lake ice using high-
906	resolution optical aerial imagery. Biogeosciences, 13, 27-44.
907	Lindgren, P.R., Grosse, G., Meyer, F.J., and Walter Anthony, K.M., 2019. An Object-Based
908	Classification Method to Detect Methane Ebullition Bubbles in Early Winter Lake Ice.
909	Remote Sensing, 11(7), 822.
910	Liu, F., Kou, D., Abbott, B.W., et al., 2019. Disentangling the Effects of Climate, Vegetation,
911	Soil and Related Substrate Properties on the Biodegradability of Permafrost-Derived
912	Dissolved Organic Carbon. Journal of Geophysical Research: Biogeosciences, 124(11),
913	3377-3389.
914	Liu, L., Wilkinson, J., Koca, K., et al., 2016. The role of sediment structure in gas bubble storage
915	and release. Journal of Geophysical Research: Biogeosciences, 121, 1992–2005.
916	Lokshina, L., Vavilin, V., Litti, Y., et al., 2019. Methane Production in a West Siberian
917	Eutrophic Fen Is Much Higher than Carbon Dioxide Production: Incubation of Peat
918	Samples, Stoichiometry, Stable Isotope Dynamics, Modeling. Water Resources, 46,
919	S110–S125.
920	Matthews, E., Johnson, M.S., Genovese, V., et al., 2020. Methane emission from high latitude
921	lakes: methane-centric lake classification and satellite-driven annual cycle of emissions.
922	Scientific Reports, 10, 12465.

923	Martinez-Cruz, K, Sepulveda-Jauregui, A, Walter Anthony, KM., et al., 2015. Geographic and
924	seasonal variation of dissolved methane and aerobic methane oxidation in Alaskan lakes.
925	Biogeosciences 12, 4213–4243.
926	Martinez-Cruz, K., Leewis, MC., Herriott, I.C., et al., 2017. Anaerobic oxidation of methane by
927	aerobic methanotrophs in sub-Arctic lake sediments. Science of the Total Environment,
928	607-608, 23–31.
929	Martinez-Cruz, K, Sepulveda-Jauregui, A, Casper, P., et al., 2018. Ubiquitous and significant
930	anaerobic oxidation of methane in freshwater lake sediments. Water Research, 144.
931	332e340.
932	Metje, M., and Frenzel, P., 2007. Methanogenesis and methanogenic pathways in a peat from
933	subarctic permafrost. Environmental Microbiology, 9(4), 954-64.
934	Nitze, I., Grosse, G., Jones, B.M., et al., 2018. Remote sensing quantifies widespread abundance
935	of permafrost region disturbances across the Arctic and Subarctic. Nature
936	Communications, 9, 5423.
937	Oh, Y., Zhuang, Q., Liu, L. et al., 2020. Reduced net methane emissions due to microbial
938	methane oxidation in a warmer Arctic. Nature Climate. Change 10, 317–321.
939	Olefeldt, D., Goswami, Grosse, G., et al., 2016. Circumpolar distribution and carbon storage of
940	thermokarst landscapes. Nature Communications, 7, 13043.
941	Osudar, R., Liebner, S., Alawi, M., et al., 2016. Methane turnover and methanotrophic
942	communities in arctic aquatic ecosystems of the Lena Delta, Northeast Siberia. FEMS
943	Microbiology Ecology, 92(8), fiw116.

- Oswald, K., Milucka, J., Brand, A., et al., 2016. Aerobic gammaproteobacterial methanotrophs
 mitigate methane emissions from oxic and anoxic lake waters. Limnology and
 Oceanography, 61(S1), S101-S118.
- 947 Parsekian, A.D., Grosse, G., Walbrecker, J.O., et al., 2013. Detecting unfrozen sediments below
- 948 thermokarst lakes with surface nuclear magnetic resonance. Geophysical Research
 949 Letters, 40, 535-540.
- Post, E., Alley, R.B., Christensen, T.R., et al, 2019. The polar regions in a 2°C warmer world,
 Science Advances, 5(12), eaaw9883.
- 952 Reeburgh, W.R., 2007. Oceanic Methane Biogeochemistry. Chem. Rev., 107, 486–513.
- Repo, E., Huttunen, J.T., Naumov, A.V., et al., 2007. Release of CO₂ and CH₄ from small
- wetland lakes in western Siberia. Tellus B: Chemical and Physical Meteorology, 59(5),
 788-796.
- Scandella, B.P., Varadharajan, C., Hemond, H.F., et al., 2011. A conduit dilation model of
 methane venting from lake sediments, Geophysical Research Letters, 38, L06408.
- 958 Schädel, C., Schuur, E., Bracho, R., et al., 2014. Circumpolar assessment of permafrost C quality
- 959 and its vulnerability over time using long-term incubation data. Global Change Biology,
 960 20, 641–652.
- Schirrmeister, L., Meyer, H., Andreev, A., et al., 2016. Late Quaternary paleoenvironmental
 records from the Chatanika River valley near Fairbanks (Alaska). Quaternary Science
 Reviews, 147, 259-278.
- 964 Schneider von Deimling, T., Grosse, G., Strauss, J., et al., 2015. Observation-based modelling of
- 965 permafrost carbon fluxes with accounting for deep carbon deposits and thermokarst

966 activity, Biogeosciences, 12, 3469–3488.

967	Schuur, E., McGuire, A., Schädel, C., 201:	5. Climate change and the permafrost carbon feedback.
968	Nature 520, 171–9.	

969 Selvam, B.P., Lapierre, J.-F., Guillemette, F., et al., 2017. Degradation potentials of dissolved

970 organic carbon (DOC) from thawed permafrost peat. Scientific Reports 7, 45811.

- 971 Sepulveda-Jauregui, A., Walter Anthony, K.M., Martinez-Cruz, K., et al., 2015. Methane and
- 972 carbon dioxide emissions from 40 lakes along a north south latitudinal transect in Alaska.
 973 Biogeosciences 12, 3197–3223.
- 974 Sepulveda-Jauregui, A., Hoyos-Santillan, J., Martinez-Cruz, K., et al., 2018. Eutrophication
- 975 exacerbates the impact of climate warming on lake methane emission. Science of The
 976 Total Environment, 636, 411-419.
- Smith, L.C., Sheng, Y., and MacDonald, G.M., 2007. A First Pan-Arctic Assessment of the
 Influence of Glaciation, Permafrost, Topography and Peatlands on Northern Hemisphere

279 Lake Distribution, Permafrost and Periglacial. Processes, 18, 201–208.

- 980 Spangenberg, I., Overduin, P.P., Damm, E., et al., 2020. Methane Pathways in Winter Ice of
- 981 Thermokarst Lakes, Lagoons and Coastal Waters in North Siberia. The Cryosphere
 982 Discuss., in review.
- Tan, Z., Zhuang, Q., and Walter Anthony, K.M., 2015. Modeling methane emissions from arctic
 lakes: model development and site-level study, J. Adv. Model. Earth Sy. 07,
- 985 doi:10.1002/2014MS000344.
- 786 Tan, Z., Zhuang, Q., Shurpali, N.J., et al., 2017. Modeling CO₂ emissions from Arctic lakes:
- 987 Model development and site-level study, J. Adv. Model. Earth Sy. 09, 2190-2213.
- 988 Tanski, G., Wagner, D., Knoblauch, C., et al., 2019. Rapid CO₂ Release From Eroding
- 989 Permafrost in Seawater. Geophysical Research Letters 46, 11244–11252.

- Telling, J., Boyd, E.S., Bone, N., et al., 2015. Rock comminution as a source of hydrogen for
 subglacial ecosystems. Nature Geoscience, 8, 851-855.
- Thalasso, F., Sepulveda-Jauregui, A., Gandois, L., et al., 2020. Sub-oxycline methane oxidation
 can fully uptake CH₄ produced in sediments: case study of a lake in Siberia. Scientific
 Reports, 10, 3423.
- Tveit, A.T., Urich, T., Frenzel, P., and Svenning, M.M., 2015. Metabolic and trophic interactions
 modulate methane production by Arctic peat microbiota in response to warming. PNAS,
 112(19), E2507-E2516.
- Turetsky, MR, Abbott, BW, Jones, MC, Anthony, KW, 2019. Permafrost collapse is accelerating
 carbon release. Nature 569, 32-34.
- Turetsky, M.R., Abbott, B.W., Jones, M.C. et al., 2020. Carbon release through abrupt
 permafrost thaw. Nature Geoscience, 13, 138–143.
- 1002 Varadharajan, C., 2009. Magnitude and spatio-temporal variability of methane emissions from a
 1003 eutrophic freshwater lake. Ph.D. dissertation, Mass. Inst. of Technol., Cambridge.
- 1004 Voigt, C., Marushchak, M.E., Abbott, B.W., et al., 2020. Nitrous oxide emissions from
- 1005 permafrost-affected soils. Nature Reviews Earth and Environment, 1, 420–434.
- 1006 Vonk, J. E., Tank, S. E., Mann, P. J., et al., 2015. Biodegradability of dissolved organic carbon in
 1007 permafrost soils and aquatic systems: a meta-analysis, Biogeosciences, 12, 6915–6930.
- Walter, Zimov, Chanton, et al., 2006. Methane bubbling from Siberian thaw lakes as a positive
 feedback to climate warming. Nature 443, 71–75.
- 1010 Walter, K.M., Chanton, J.P., Chapin III, F.S., et al., 2008. Methane production and bubble
- 1011 emissions from arctic lakes: Isotopic implications for source pathways and ages. Journal
- 1012 of Geophysical Research: Biogeosciences, 113, G00A08.

1013	Walter Anthony, K.M., Vas, D.A., Brosius, L., et al., 2010. Estimating methane emissions from
1014	northern lakes using ice- bubble surveys. Limnol. Oceanogr. Methods, 8, 592-609.
1015	Walter Anthony, K.M., Anthony, P., Grosse, G., and Chanton, J., 2012. Geologic methane seeps
1016	along boundaries of arctic permafrost thaw and melting glaciers. Nature Geoscience, 5,
1017	419–426.
1018	Walter Anthony, KM, Zimov, SA, Grosse, G, et al., 2014. A shift of thermokarst lakes from
1019	carbon sources to sinks during the Holocene epoch. Nature 511, 452–456.
1020	Walter Anthony, K.M., Daanen, R., Anthony, P., et al., 2016. Methane emissions proportional to
1021	permafrost carbon thawed in Arctic lakes since the 1950s. Nature Geoscience, 9, 679-
1022	681.
1023	Walter Anthony, K., Schneider von Deimling, T., Nitze, I. et al., 2018. 21st-century modeled
1024	permafrost carbon emissions accelerated by abrupt thaw beneath lakes. Nature
1025	Communications, 9, 3262.
1026	Walter Anthony, K.M., Lindgren, P., Hanke, P., et al., submitted. Decadal-scale hotspot CH4
1027	emission following abrupt permafrost thaw. Environmental Research Letters.
1028	Wei, S., Cui, H., Zhu, Y., et al., 2018. Shifts of methanogenic communities in response to
1029	permafrost thaw results in rising methane emissions and soil property changes.
1030	Extremophiles, 22, 447–459.
1031	Weyhenmeyer, G., Kosten, S., Wallin, M. et al., 2015. Significant fraction of CO ₂ emissions
1032	from boreal lakes derived from hydrologic inorganic carbon inputs. Nature Geoscience,
1033	8, 933–936.
1034	Wik, M., Varner, R., Anthony, K. et al., 2016. Climate-sensitive northern lakes and ponds are

1035 critical components of methane release. Nature Geoscience, 9, 99–105.

- 1036 Williams, P. J. and Smith, M. W., 1989. The Frozen Earth: Fundamentals of Geocryology,
- 1037 Cambridge University Press, Cambridge, UK.
- 1038 Winkel, M., Sepulveda-Jauregui, A., Martinez-Cruz, K., et al., 2019. First evidence for cold-
- 1039 adapted anaerobic oxidation of methane in deep sediments of thermokarst lakes.
- 1040 Environmental Research Communications, 1, 2.
- 1041 Zimov, S.A., Voropaev, Y.V., Semiletov, I.P., et al., 1997. North Siberian Lakes: A Methane
- 1042 Source Fueled by Pleistocene Carbon. Science, 277(5327), 800-802.