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Vertical root distribution and biomass allocation along proglacial chronosequences in Central Switzerland

Konrad Greinwald  , Lea Adina Dieckmann  , Carlotta Schipplick^a, Anne Hartmann^b, Michael Scherer-Lorenzen  , and Tobias Gebauer  

^aGeobotany, Faculty of Biology, University Freiburg, Freiburg, Germany; ^bSection Hydrology, GFZ German Research Centre for Geosciences, Potsdam, Germany

ABSTRACT

Investigating changes in belowground functional plant traits is an important step toward a better understanding of vegetation dynamics during primary succession. However, in alpine glacier forelands, we still lack an accurate assessment of plant rooting patterns. In this study, we established two proglacial chronosequences with contrasting bedrocks to investigate changes in rooting patterns and biomass allocation with terrain age. We extracted soil cores up to 1 m depth and measured root traits every 10 cm of each drilled core. Furthermore, we sampled aboveground biomass determining the contributions of functional groups to total aboveground biomass. We found that root traits associated with the root economics spectrum varied significantly along the chronosequences. Vertical root distribution coefficients revealed that early successional communities had more evenly distributed root systems compared to late successional communities. Biomass allocation showed diverging patterns. We found evidence for both the isometric allocation and optimal partitioning hypotheses. In addition, we observed a significant correlation between rooting parameters and plant community composition, suggesting that the dominance of distinct plant functional groups was one important factor explaining the observed rooting patterns. Our results shed light on the often neglected belowground compartments during plant succession and contribute to a better understanding of hillslope functioning.

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community composition

Introduction

Since the late Pleistocene, glaciers in alpine regions have been shrinking dramatically (e.g., Boxleitner et al. 2019). Former glacier positions can be identified by moraines representing distinct ages of substrate exposure (e.g., Egli, Fitze, and Mirabella 2001; Musso et al. 2019; Maier et al. 2020). This space-for-time substitution, called a chronosequence approach, enables scientists to test hypotheses related to terrain age. In glacier forelands, ecologists have frequently used this approach to study plant succession, which is defined as the turnover of species and communities over time in response to a disturbance (Matthews 1992; Walker and Del Moral 2003; Prach and Walker 2020).

Retreating glaciers expose young soils that have little biological legacy (Matthews 1992; Prach and Walker 2020). Such barren surfaces represent a very inhospitable habitat for early colonizers (Caccianiga et al. 2006).

Although pioneer species must deal with high abiotic stress levels, such as low nutrients, extreme temperatures, and high ultraviolet radiation levels, vegetation cover and aboveground biomass are increasing quickly and a fully vegetated surface is normally observed after a few centuries (Matthews 1992; Walker and Del Moral 2003; Erschbamer and Caccianiga 2017). During succession in alpine glacier forelands, not only do cover and biomass change quickly but species composition does, too (Matthews and Whittaker 1987; Chapin et al. 1994; Raffl and Erschbamer 2004; Raffl et al. 2006; Robbins and Matthews 2009, 2010; Burga et al. 2010). Because of the site-specific differences in abiotic and biotic conditions, successional seres are known to be highly variable (Caccianiga and Andreis 2004; Schumann, Gewolf, and Tackenberg 2016). Nevertheless, patterns of vegetation dynamics show similarities in their main features. Young moraines are often covered by scree plant communities, followed by patches of initial grasslands

CONTACT Konrad Greinwald  konrad.greinwald@biologie.uni-freiburg.de  University of Freiburg, Faculty of Biology, Geobotany, Schänzlestrasse 1, D-79104 Freiburg, Germany.

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and snowbed communities (Andreis, Caccianiga, and Cerabolini 2001). Subsequently, densely covered alpine grasslands are formed. Late successional communities in alpine glacier forelands often show elements of dwarf shrub communities or even subalpine forest stands, depending on elevation (Lüdi 1955, 1958; Burga et al. 2010). For example, because of more stressful conditions, on higher elevations, the speed and success of plant colonization, especially of shrub species, are reduced compared to lower elevations (Schumann, Gewolf, and Tackenberg 2016).

In addition to analyses of species turnover, trait-based approaches have increasingly gained attention for studying succession in recent years (Prach, Pyšek, and Šmilauer 1997; Fukami et al. 2005; Weppler and Stöcklin 2005; Caccianiga et al. 2006; Franzén et al. 2019). Functional traits reflect species' strategies to meet the local requirements during succession, such as environmental conditions or spatiotemporal isolation. Therefore, investigating functional plant traits offers the potential to mechanistically understand the underlying ecological processes of succession (Schleicher, Peppler-Lisbach, and Kleyer 2011; Raever, Violette, and Munoz 2012; Prach and Walker 2020). Until now, most research dealing with functional trait changes along successional gradients has focused on aboveground compartments of plants, whereas our understanding of belowground plant traits has lagged far behind (Holdaway et al. 2011; Erktan, McCormack, and Roumet 2018). However, belowground traits play a key role in understanding ecosystem functioning and for preserving ecosystem services (Bardgett, Mommer, and de Vries 2014; Garnier, Navas, and Grigulis 2016; Erktan, McCormack, and Roumet 2018). For instance, roots are involved in the regulation of plant–soil interactions (van der Putten et al. 2013), crucial for carbon as well as nutrient cycling (Hendricks, Nadelhoffer, and Aber 1993), and essential for maintaining slope stability (Freschet et al. 2017).

Two important belowground traits are specific root length (SRL, length of root per unit mass) and root tissue density (RTD, root mass per root volume). SRL and RTD are commonly used as equivalents to the functional leaf traits specific leaf area and leaf dry matter content (Ryser and Eek 2000; Ryser 2006; Freschet, Swart, and Cornelissen 2015). Because SRL and RTD are correlated with root life span, both are frequently used as key traits for the root economics spectrum (RES) hypothesis (Ryser 1996; Reich 2014; F. Li et al. 2019). Under the RES hypothesis, roots are assumed to follow a gradient in trait syndromes from fast foraging and short life span (acquisitive strategy) to slow foraging and long life span (conservative strategy; Freschet et al. 2010; Reich 2014; Kong et al. 2019). Such plant strategy types can be also identified using aboveground leaf traits (Hodgson et al. 1999; Wright et al. 2004). This approach has

been tested on primary succession in glacier forelands, showing that pioneer communities are dominated by fast-growing species with high nitrogen leaf levels (Caccianiga et al. 2006). During succession, these species are progressively replaced by those with lower growth rates and denser leaves (Caccianiga et al. 2006; Gobbi et al. 2010). Theoretically, these findings should also apply to belowground traits when assuming that SRL and RTD behave like their aboveground equivalents.

Another belowground trait that is poorly investigated is vertical root distribution. This trait provides information about the morphology of rooting systems, which are highly plastic in response to abiotic and biotic changes and determine the efficiency of different rooting functions, such as water and nutrient uptake as well as anchoring in the soil (Bardgett, Mommer, and de Vries 2014). The spatial distribution of roots in the soil varies inter- and intraspecifically and reflects plant strategies of local adaptation to different physical and hydrological soil conditions (P. Hartmann and von Wilpert 2013). Furthermore, root distribution is linked to root architecture, which is known to influence hillslope functioning by creating networks of preferential flow (Ghestem, Sidle, and Stokes 2011; A. Hartmann et al. 2020). The most prominent model dealing with root distribution was proposed by Gale and Grigal (1987). It assumes an asymptotic relationship between root biomass and soil depth, as described by the extinction coefficient β . Gale and Grigal (1987) found that early successional tree species had the potential for a deep exploitation of the soil, which was hypothesized to be an adaptation to the homogenous distribution of nutrients and water in the substrate of early successional habitats. Jackson et al. (1996) and Schenk and Jackson (2002) conducted a global comparison of root distributions across different biomes, revealing that tundra, boreal forest, and temperate grasslands had the shallowest rooting profiles. Furthermore, they showed that plant functional groups differed significantly in their root systems, with grasses having 44 percent of their biomass allocated in the topsoil, followed by trees at 26 percent and shrubs at 22 percent. Consequently, communities consisting of different functional groups are suspected to exhibit distinct rooting patterns. Moreover, in diverse habitats, niche differentiation concerning resource requirements and uptake or species interactions may cause differences in rooting patterns compared to less diverse plant communities (de Kovel, Wilms, and Berendse 2000; Mommer et al. 2010; Poorter et al. 2012). In alpine habitats, however, studies on rooting systems are very scarce (Pohl et al. 2011). To our knowledge, there is no study on vertical root distribution development along chronosequences and, therefore, we lack

a general understanding of how vertical root distribution changes during primary succession.

Measuring vertical root distribution enables researchers to calculate belowground biomass (BGB). If compared to the aboveground equivalent (AGB), this parameter provides information about biomass allocation, which is an important topic in plant ecology (McConaughay and Coleman 1999; Poorter et al. 2012). In general, two contrasting hypotheses are debated (Shipley and Meziane 2002; McCarthy and Enquist 2007; Zeng, Wu, and Zhang 2015). The isometric allocation hypothesis assumes that AGB and BGB are related in an isometric manner; that is, the slope of the log–log relationship between AGB and BGB is not significantly different from one (Enquist and Niklas 2002; Niklas 2006; Yang et al. 2009). This is true for a diverse range of plant species and community types (Müller, Schmid, and Weiner 2000; Yang et al. 2009; Yang and Luo 2011). In contrast, according to the optimal partitioning hypothesis, plants respond to a gradient of environmental conditions (Nie et al. 2016; Dai et al. 2019, 2020) by allocating biomass among various organs to capture nutrients, water, and light (Bloom, Chapin, and Mooney 1985; Freschet et al. 2018). This hypothesis suggests that plants allocate more biomass to photosynthetic tissues under nutrient-rich conditions and invest more biomass into belowground organs in nutrient-poor conditions (McConaughay and Coleman 1999). During primary succession, environmental conditions are expected to change considerably and a strong turnover of plant strategy types is observed (Caccianiga et al. 2006). In these habitats, biomass allocation should reflect the turnover of plant strategies and therefore optimal partitioning is expected during primary succession.

During primary succession, changes in root traits, root distribution, and biomass allocation are widely expected to occur, but both the patterns of these changes and their causes remain largely unexplored. The objective of this study was to shed light on the belowground features of plant succession. For this purpose, we examined the vertical root distribution and biomass allocation in two glacier forelands with distinct bedrocks. We also analyzed the influence of vegetation composition on vertical root distribution and root biomass. Chiefly, we tested the following hypotheses: (a) We expected that functional root traits vary along the chronosequences accordingly to the RES, indicating an acquisitive strategy at young moraines and a conservative strategy at old moraines. (b) We assumed that roots are more evenly distributed in early than in late successional habitats. (c) Due to the strong environmental heterogeneity along the chronosequences, biomass allocation patterns should follow the optimal partitioning hypothesis. (d) Finally, we hypothesized that vertical root

distribution should be correlated with plant community composition.

Material & methods

Study sites

The present study made use of two long-term glacier foreland chronosequences with different bedrocks, each incorporating four moraines that span an age gradient from 30 a (a = years) to 13.5 ka (ka = thousands of years; [Figure 1](#), [Table 1](#); Musso et al. 2019). Both glacier forelands (Stein Glacier, Griess Glacier) are situated in Central Switzerland and are formed over siliceous and calcareous parent material. A summary of soil data along the two chronosequences is provided in Musso et al. (2019). An overview of some important vegetation characteristics is provided in [Table 1](#). It is important to note that because of differences in species composition, the moraines represent different stages along the successional gradients.

- (1) Stein Glacier: This study site is located west of the Susten Pass in the Canton Bern ($47^{\circ}43' N$, $8^{\circ}25' E$). The chronosequence of this site consisted of four moraines with estimated terrain ages of 30 a, 160 a, 3 ka, and 10 ka ([Figure 1](#), [Table 1](#)). The four moraines of this site were dated in previous studies using radiocarbon dating, surface exposure dating as well as historical maps (Schimmelpfennig et al. 2014; Musso et al. 2019). The local bedrock at the Stein Glacier foreland is covered by glacial till on pre-Mesozoic silicate parent material, comprised of metamorphosed meta-granitoids, gneisses, and amphibolites (Schimmelpfennig et al. 2014; Musso et al. 2019). All four moraines are situated above the timberline at elevations between 1,880 and 1,990 m.a.s.l. The soil types of this chronosequence were classified as Hyperskeletal Leptosols (30 a, 160 a) to Skeletal Cambisols (3 ka) and Entic Podzols (10 ka; [Table 1](#); see also Musso et al. 2020).
- (2) Griess Glacier: This chronosequence is situated 25 km away from the Stein Glacier site. It is located near the Klausen Pass, in the glacier foreland of Griess Glacier, Canton Uri ($46^{\circ}50' N$, $8^{\circ}49' E$). The moraines of the Griess Glacier foreland had estimated terrain ages of 80 a, 160 a, 4.9 ka, and 13.5 ka ([Figure 1](#), [Table 1](#)). The determination of terrain ages of the moraines was done by radiocarbon dating of bulk soil and by using historical

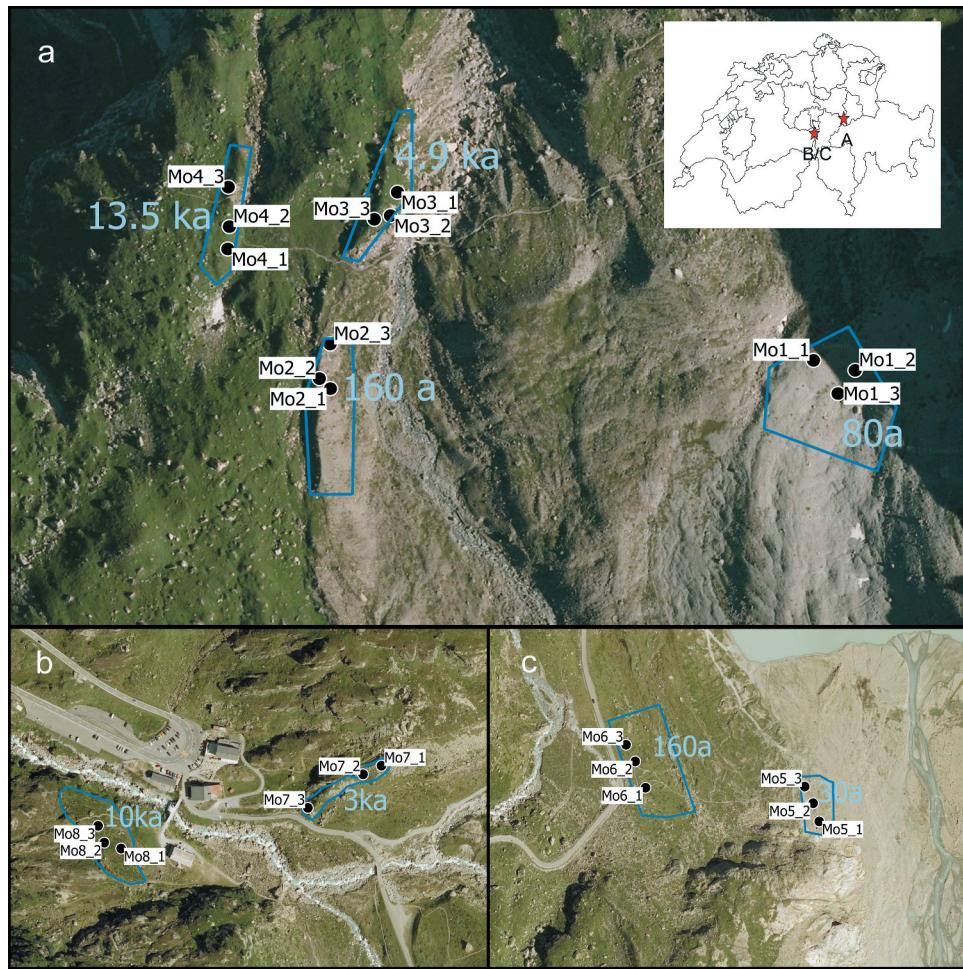


Figure 1. Location of (A) the Griess Glacier foreland ($46^{\circ}50' N$, $8^{\circ}49' E$) and (B), (C) the lower and upper parts of the Stein Glacier foreland ($47^{\circ}43' N$, $8^{\circ}25' E$). The terrain age of the moraines is given in years (a) and thousands of years (ka). Inset in (A) Switzerland with locations of both study sites. Satellite images: Google Maps, 2020, <https://google.de/maps/place/Schweiz/@46.6192509,7.4679619>.

Table 1. Study site characteristics with terrain ages of the moraines, elevation, slope exposition, slope, bedrock, vegetation cover, species richness, and vegetation type of Stein Glacier (Susten Pass) and Griess Glacier (Klausen Pass) forelands.

Moraine	Terrain age (years)	Elevation (m.a.s.l.)	Slope exposition	Slope (%)	Bedrock	Soil type	Vegetation cover (%)	Species richness (n)	Vegetation type
Stein Glacier (Susten Pass, Central Switzerland)									
30 a	30	1,990	ENE	40	Gneiss	Hyperskeletal Leptosol	49	23	Pioneer vegetation
160 a	160	1,990	ENE	31	Gneiss	Hyperskeletal Leptosol	98	43	Initial grassland and <i>Salix</i> shrubberies
3 ka	3,000	1,890	SW	25	Gneiss	Skeletal cambisol	94	33	<i>Carex sempervirens</i> grassland
10 ka	10,000	1,880	N	22	Gneiss	Entic Podzol	100	29	<i>Rhododendron ferrugineum</i> shrubberies and <i>Nardus</i> grassland
Griess Glacier (Klausen Pass, Central Switzerland)									
80 a	80	2,200	NNW	27	Limestone	Hyperskeletal Leptosol	21	31	Pioneer vegetation
160 a	160	2,030	NE	31	Limestone	Hyperskeletal Leptosol	48	32	Scree vegetation
4.9 ka	4,900	2,010	N	37	Limestone	Calcaric Skeletal Cambisol	98	19	<i>Carex ferruginea</i> grassland
13.5 ka	13,500	2,000	NW	44	Limestone	Calcaric Skeletal Cambisol	100	11	<i>Rhododendron hirsutum</i> shrubberies and <i>Carex ferruginea</i> grassland

maps (Musso et al. 2019). In this glacier foreland, the deep-lying gneiss bedrock of the Griess Glacier is covered with limestone scree (Oechslin 1935; Musso

et al. 2019). The elevation of the four moraines of this site is between 2,000 and 2,200 m.a.s.l. The soil types ranged from Hyperskeletal Leptosols (80 a,

160 a) to Calacaric Skeletic cambisols (4.9 ka, 13.5 ka; **Table 1**; see also Musso et al. 2020).

Sampling strategy

On each moraine, we established three plots (each 4 m × 6 m) that were selected based on a structural vegetation complexity measure. Briefly, structural vegetation complexity was defined as an index based on vegetation cover and functional diversity data (Maier et al. 2020). Functional diversity was calculated based on the following eight traits characterizing species along the main axes of plant performance (Garnier, Navas, and Grigulis 2016): specific leaf area, nitrogen content, leaf dry matter content, Raunkiae's life form, seed mass, clonal growth organ, root type, and stem growth form. For the plot selection, we conducted a vegetation mapping differentiating between vegetation units classified according to the characteristic recurrent combination of species. In each unit, we recorded all vascular plant species with proportion cover and calculated the vegetation complexity index. On every moraine, the three plots were placed within the surface units with the lowest, intermediate, and highest vegetation complexity.

On every plot, we conducted vegetation surveys, recording percentage plant cover of every single species by visual estimation. From these data, we calculated total plant cover and the cover of the different functional groups (grasses, forbs, shrubs). Species richness was computed as the total number of species per plot. Furthermore, we selected four 10 cm × 50 cm stripes and harvested AGB, distinguishing between grasses, forbs, and shrubs. The biomass samples were then dried (60°C, 72 hours) and weighed.

In addition, we extracted at least four soil cores per plot using a stainless steel soil column cylinder (diameter = 5 cm) that was drilled into the soil with a heavy, electrically powered percussion hammer (Makita HM 1800, Ratingen, Germany). The extracted cores were analyzed to a maximum core length of 1 m to maintain consistency. Each core was separated into 10-cm samples (0–10, 10–20, 20–30, 40–50, 60–70, 80–90, 90–100 cm). In total, we analyzed 394 samples at the Stein Glacier site and 432 samples at the Griess Glacier site. Sampling was done in July and August 2018 (Stein Glacier) and 2019 (Griess Glacier).

Processing and analyzing the roots

The roots were cleaned of soil and sorted into three diameter classes: <1 mm (i.e., fine roots), 1–2 mm (i.e., fibrous roots), and >2 mm (i.e., coarse roots). The samples were scanned in water with a flatbed scanner (Epson Perfection V700 Photo, Seiko Epson Corporation, Nagano, Japan, resolution 800 dpi). We used the software WinRHIZO

Table 2. List of traits investigated in the present study.

Root trait	Unit	Description
FRMD	g cm ⁻³	Dry weight of fine roots (diameter <1 mm) per unit soil volume
FRLD	cm cm ⁻³	Root length of fine roots (diameter <1 mm) per unit soil volume
SRL	cm g ⁻¹	Root length per unit dry weight of fine roots
RTD	g cm ⁻³	Dry weight per root volume of fine roots
β		Measure for shallow and deep rooting, respectively: $Y = 1 - \beta^d$
BGB	kg m ⁻²	Belowground biomass
AGB	kg m ⁻²	Aboveground biomass
R:S ratio	%	Root-to-shoot ratio

Reg 2013e (Régent Instruments Inc. 2013) to measure the root length and root volume from the scans. Thereafter, each sample was dried (60°C, 72 hours) and weighed. To reduce workload, where more than ten scans would be necessary to capture all fine roots, material was subsampled.

For each sample, fine root mass density (FRMD), fine root length density (FRLD), SRL, and RTD were calculated (**Table 2**). FRMD and FRLD were obtained by dividing the root mass and root length, respectively, of the fine roots by the soil volume, which was determined by subtracting the volume of stones (mesh size of sieve >2 mm) from the volume of the corresponding soil cylinder. SRL was calculated as the length of fine roots per fine root mass of the sample and RTD as root mass per root volume of fine roots. Furthermore, we calculated the total root biomass by adding up the root dry weight of every soil depth increment and relating them to the area of 1 m² (**Table 2**).

Statistical analyses

As a first step, we calculated the per plot averages of all samples. To test hypothesis (a), we built linear mixed effect models. We used a three-way interaction of site (Griess Glacier, Stein Glacier), soil depth, and terrain age as fixed effects to investigate the changes of root traits (FRMD, FRLD, SRL, RTD) as a response of terrain age and soil depth. As such, soil depth was fitted as a factor variable. The affiliation of the plots to the moraines was included as a random term. To meet model assumptions, response variables were log-transformed. Linear mixed models were created using the R package lme4 (Bates et al. 2007). The model outputs are shown in Appendix S1.

To test hypothesis (b), we calculated the vertical root distribution coefficient according to the model of Gale and Grigal (1987):

$$Y = 1 - \beta^d,$$

where Y is the cumulative root fraction, d is the soil depth (cm), and β is the fitted extinction coefficient. The model suggests a nonlinear asymptotic relationship between root

measures and soil depth. We used the fine root measures of FRMD and FRLD to calculate the cumulative root fraction (Y) for any 10-cm step from the surface to 1 m soil depth and derived the extinction coefficient (β). The cumulative root fraction represents the percentage of FRMD/FRLD from the soil surface to the depth considered relative to total FRMD/FRLD of the soil profile. The extinction coefficient β is a measure of vertical fine root distribution and may be interpreted as the allocation pattern of FRMD/FRLD. The values of β range from 0 to 1, where 1 indicates that the whole root biomass or root length, respectively, is located in deep soil and 0 indicates that the whole root biomass or root length, respectively, is concentrated at the surface.

To test the isometric allocation hypothesis (c), we performed simple linear models fitting log AGB as a function of log BGB. To investigate the change of AGB, BGB, and root-to-shoot (R:S) ratio along the chronosequences, we computed one-way analyses of variance and subsequent post hoc comparisons using the least significant difference (LSD) test. Significance levels were corrected according to the Bonferroni adjustment for multiple testing (Scheiner 1993).

The relationships between aboveground vegetation characteristics and vertical root distribution (hypothesis (d)) were analyzed using Pearson correlation coefficients. All analyses were performed with R version 3.5.1 (R Core Team 2018).

Results

Functional root traits along the chronosequences

The vertical distribution of the investigated functional root traits (FRMD, FRLD, SRL, RTD) showed large variations across all moraines and soil depths (Figure 2). At the Stein Glacier, FRMD ranged from 0 to 2.06 g cm⁻³, FRLD from 0.03 to 12.51 cm cm⁻³, SRL from 3.3 to 157.3 m g⁻¹, and RTD from 0.01 to 0.67 g cm⁻³. At the Griess Glacier, FRMD ranged from 0 to 3.35 g cm⁻³, FRLD from 0.02 to 24.68 cm cm⁻³, SRL from 6.5 to 200.0 cm g⁻¹, and RTD from 0.24 to 0.65 g cm⁻³ (Figure 2). Across all moraines, FRMD and FRLD decreased exponentially with soil depth (Figures 2a–2d). In the top 20 cm of the soil profiles, FRMD and FRLD were considerably higher on old moraines (2.21 g cm⁻³, 16.48 cm cm⁻³) than on young moraines (0.14 g cm⁻³, 1.58 cm cm⁻³). This finding was also reflected by positive effect sizes, indicating the rate of change with terrain age at certain soil depths (Figures 3a–3d). In the topsoil, SRL decreased, and deeper in the soil it increased

with terrain age (Figures 2e and 2f, Figures 3e and 3f). An opposite pattern was observed for RTD, showing positive effect sizes in upper soil layers and negative effect sizes deeper in the soil (Figures 3g and 3h).

β coefficients along the chronosequences

The vertical root distribution coefficient β ranged from 0.818 to 0.974 for FRMD and from 0.838 to 0.963 for FRLD (Table 3). At the Griess Glacier site, the coefficients changed significantly, whereas at the Stein Glacier site no significant changes were found. The coefficients showed a similar pattern along both chronosequences (Figure 4, Table 3). We found the highest β values at the 30 a, 80 a moraines and the lowest values at the 3 ka and 4.9 ka moraines (Figure 4, Table 3). On the oldest moraines of both sites, more than 90 percent of the root biomass was allocated in the uppermost 30 cm of soil, versus approximately 68 percent on the youngest moraines (Table 3).

Biomass allocation patterns along the chronosequences

AGB showed diverging patterns in the two glacier forelands. At the Stein Glacier site, AGB increased significantly along the chronosequence (Figure 5a), ranging from 0.16 to 1.29 kg m⁻² (Table 4). At the Griess Glacier foreland, it remained constant at around 0.40 kg m⁻² (Figure 5b, Table 4). Along both chronosequences, shrubs accounted for the largest proportions of AGB (Table 4). At the Stein Glacier foreland, the proportion of shrubs to total AGB increased with terrain age (except for the 3 ka), whereas grass and forb proportions decreased. At the Griess Glacier foreland, the proportion of grasses to total AGB increased with terrain age, forb proportion remained constant, and shrub proportion decreased (Table 4).

BGB increased along both chronosequences and showed significant changes across the moraines (Figures 5c–5d, Table 4). At the Stein Glacier site, BGB showed minimal values at the 30 a moraine (0.16 kg m⁻²), reaching maximal values at the 10 ka moraine (2.29 kg m⁻²; Table 4). Similarly, at the Griess Glacier site, BGB increased constantly from the youngest moraine (0.09 kg m⁻²) to the oldest moraine (2.03 kg m⁻²; Table 4). On young moraines (terrain age \leq 160 a) of both sites, fine roots (<1 mm) accounted for the largest proportion of BGB. On old moraines (terrain age \geq 3 ka), contributions to BGB depended on vegetation composition. Where grasses dominated, fine roots constituted the main percentage of BGB. In contrast, coarse root

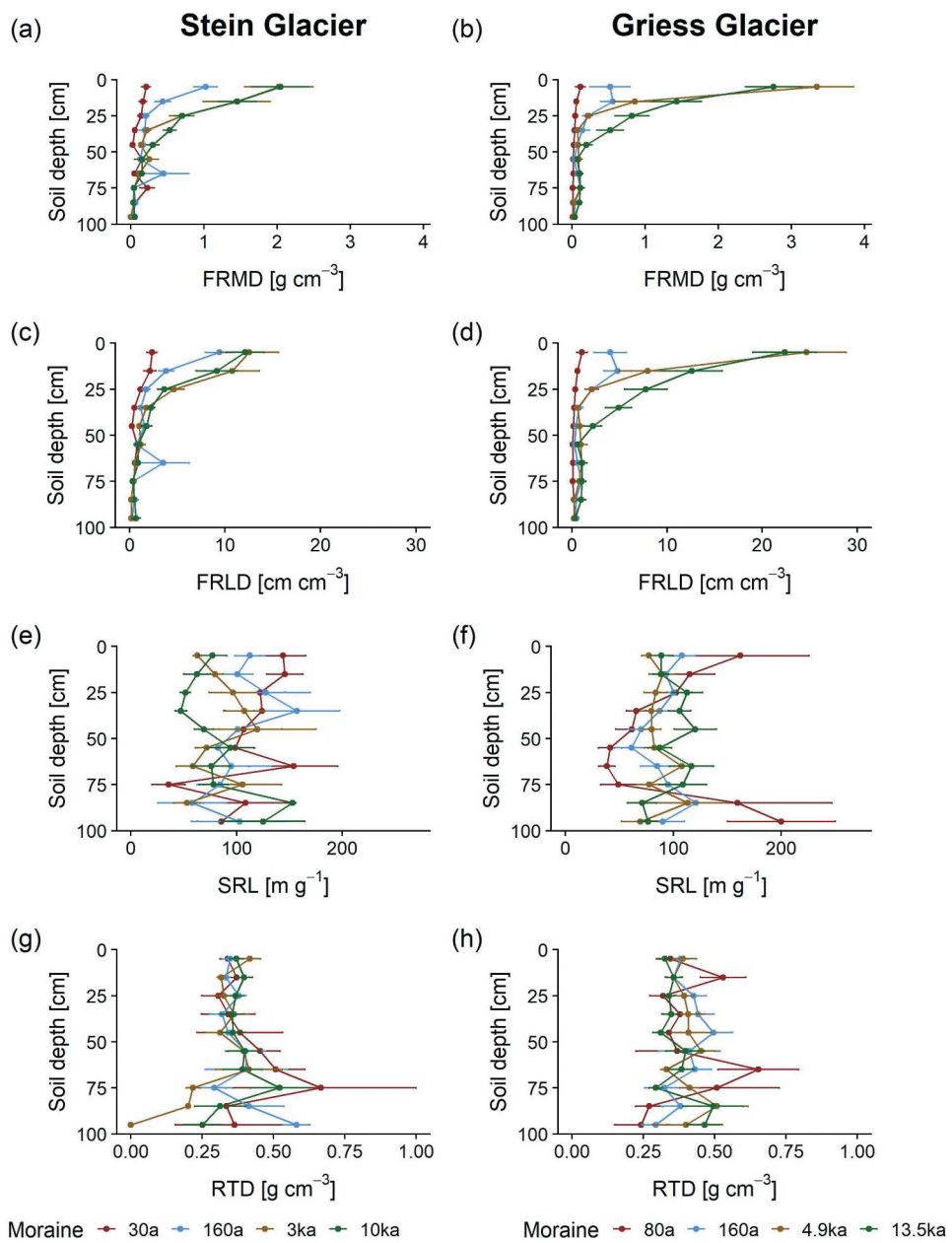


Figure 2. Vertical distributions of mean FRMD, FRLD, SRL, and RTD per soil depth increment across the moraines of (a), (c), (e), (g) Stein Glacier and (b), (d), (f), (h) Griess Glacier forelands. Error bars show the standard error of the mean. The terrain age of the moraines is given in years (a) and thousands of years (ka).

fraction (>2 mm) was highest where shrubs dominated (Table 4).

The R:S ratio showed a diverging pattern between the two study sites (Figures 5e–5f, Table 4). At the Stein Glacier site, the R:S ratio did not show significant changes between the moraines (range: 0.96–2.32; Figure 5e, Table 4), whereas at the Griess Glacier site, the R:S ratio increased significantly along the chronosequence (Figure 5f), ranging from 0.22 at the youngest moraine to 5.29 at the oldest moraine (Table 4).

At the Stein Glacier site, we found a significant relationship between log AGB and log BGB, with a slope of

approximately one ($R^2 = 0.7$, $p < .05$, Figure 5g). At the Griess Glacier site, no such trend was observed (Figure 5h).

Correlations between rooting parameters and vegetation composition

The Pearson correlation matrix showed different correlations between β coefficients, BGB, and vegetation measures (Table 5). The AGB of grass species correlated negatively with β coefficients and positively with BGB ($p < .05$). We also found a significant positive correlation

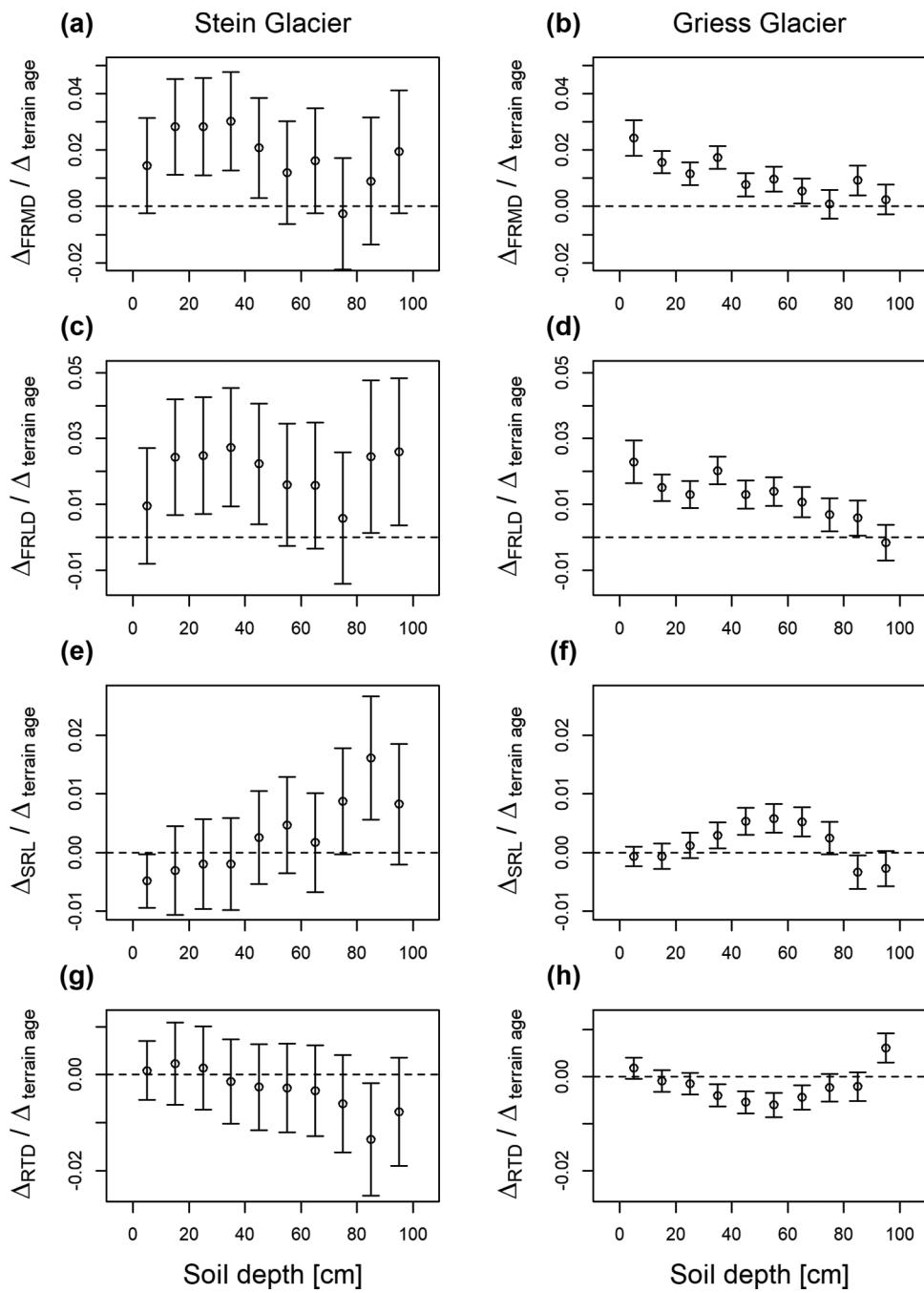


Figure 3. Effect sizes of FRMD, FRLD, SRL, and RTD of (a), (c), (e), (g) Stein Glacier and (b), (d), (f), (h) Griess Glacier forelands. Effect sizes indicate the rate of change of the root traits per 100 years of terrain age. The response variables are on a log scale.

($p < .05$) between BGB and AGB as well as between BGB and shrub cover (Table 5).

Discussion

Functional root traits along the chronosequences

The large differences in root densities across the moraines are related to a set of parameters linked to terrain age. Because AGB and BGB are typically positively

correlated, the observed increase in FRMD and FRLD should be a function of vegetation dynamics; for example, increasing plant cover and biomass along the chronosequences (Y. Li, Luo, and Lu 2008; Yang et al. 2009; Dai et al. 2019, 2020). Furthermore, various soil properties are known to influence root growth (Unger and Kaspar 1994; Ho et al. 2005). For example, an increase in bulk density leads to an increase in soil resistance, which can impede root penetration. Musso et al. (2019) and A. Hartmann et al. (2020) reported a strong gradient

Table 3. Mean vertical root distribution coefficient, β , and root fractions in the soil top 30 cm ($RF_{Top30cm}$) across the moraines of the study sites.

Site	Moraine	β_{FRMD}	$RF_{Top30cm}$	β_{FRLD}	$RF_{Top30cm}$
Stein Glacier	30 a	0.974 ^a	70.15 ^a	0.961 ^a	54.61 ^a
Stein Glacier	160 a	0.942 ^a	80.63 ^a	0.947 ^a	83.22 ^a
Stein Glacier	3 ka	0.921 ^a	92.08 ^a	0.919 ^a	91.50 ^a
Stein Glacier	10 ka	0.936 ^a	87.38 ^a	0.933 ^a	86.09 ^a
Griess Glacier	80 a	0.961 ^a	67.46 ^b	0.963 ^a	69.71 ^b
Griess Glacier	160 a	0.956 ^a	72.53 ^b	0.958 ^a	73.68 ^{ab}
Griess Glacier	4.9 ka	0.818 ^b	99.50 ^a	0.838 ^b	99.76 ^a
Griess Glacier	13.5 ka	0.922 ^a	88.31 ^{ab}	0.931 ^a	91.34 ^{ab}

Notes. The model of Gale and Grigal (1987) was used to calculate β coefficients from FRMD and FRLD. A post hoc LSD test was conducted to compare root distribution measures among the moraines. The significance level was adjusted according to the Bonferroni method. Superscript letters indicate significant differences. The terrain age of the moraines is given in years (a) and thousands of years (ka).

in bulk density along the chronosequences of Stein Glacier and Griess Glacier. According to Musso et al. (2019), bulk density showed values close to 2 g cm^{-3} on the youngest moraines, which is considered to be limiting for root growth (Unger and Kaspar 1994). With

ongoing succession, bulk density decreased due to an accumulation of root biomass and organic matter (Musso et al. 2019, 2020). Such soils are richer in fine material and hence provide better conditions for roots to penetrate the substrate. Thus, we conclude that the observed root density patterns are an expression of both the increasing colonization of plants and the changing physical soil conditions.

Depending on soil depth, the results for RTD and SRL presented an ambiguous picture. In the topsoil of the youngest moraines (30 a, 80 a), the comparatively high values of SRL and the tendency for low RTD values reflect a fast water and nutrient uptake strategy of early colonizers (Caccianiga et al. 2006; Erschbamer and Caccianiga 2017). In contrast, species of late successional communities seem to invest more biomass into strengthening root tissues. This trait pattern supports the RES hypothesis postulating a gradient in trait syndromes from an acquisitive to a conservative strategy (Ryser 1996; Reich 2014; Kong et al. 2019; F. Li et al.

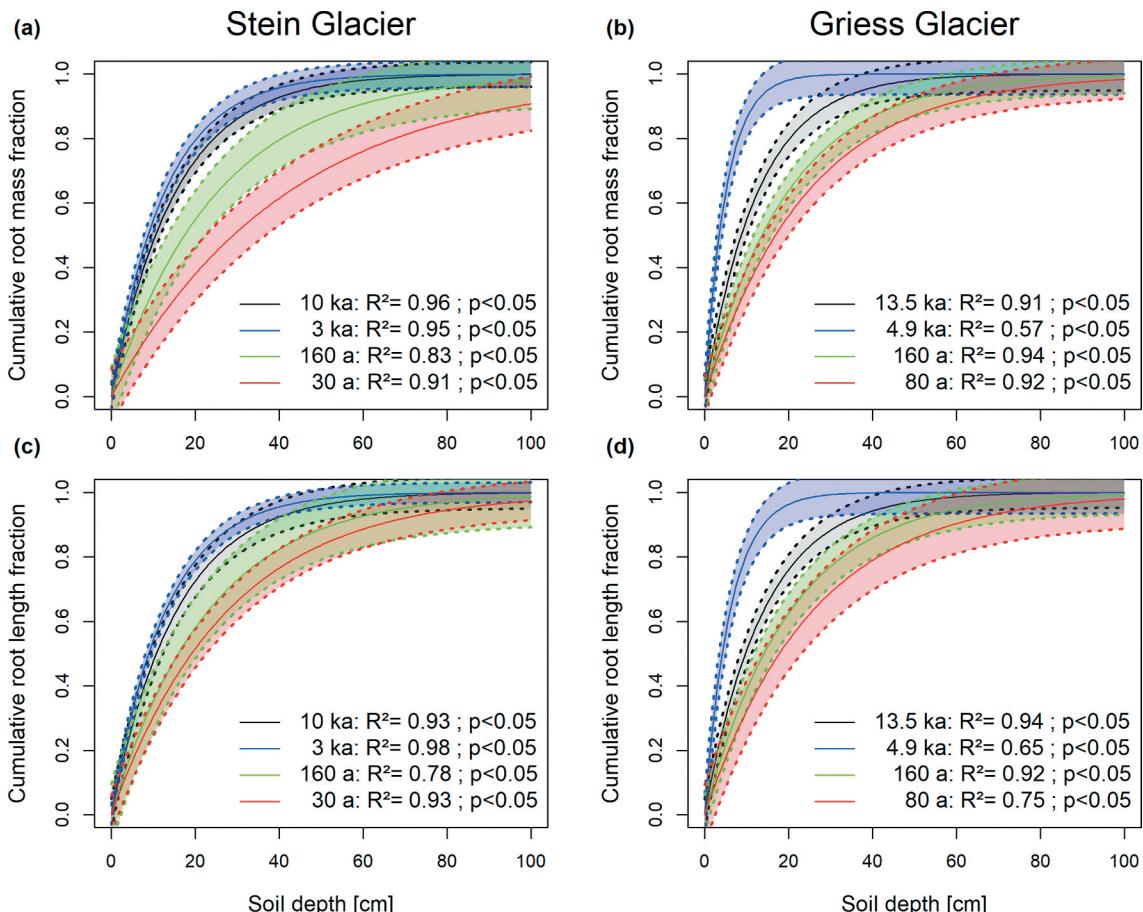


Figure 4. Vertical distribution of (a), (b) root mass and (c), (d) root length across the moraines (see Table 1) of the (a), (c) Stein Glacier and (b), (d) Griess Glacier forelands. The vertical root distribution was fitted using the model proposed by Gale and Grigal (1987). The resulting β coefficients are shown in Table 4. The cumulative root fraction represents the percentage of roots from the soil surface to the depth considered relative to the total roots of the soil profile. Models are plotted showing the 95 percent confidence intervals.

**Table 4.** Mean BGB, AGB, and R:S ratio across the moraines of the study sites.

Site	Moraine	BGB (kg m^{-2}) (<1 mm/1–2 mm/>2 mm)	AGB (kg m^{-2}) (grasses/forbs/shrubs)	R:S ratio
Stein Glacier	30 a	0.16 (82/11/16)	0.16 (14/23/63)	0.96
Stein Glacier	160 a	0.99 (48/14/38)	0.84 (6/11/84)	1.18
Stein Glacier	3 ka	0.96 (82/10/8)	0.41 (28/51/20)	2.32
Stein Glacier	10 ka	2.29 (40/10/49)	1.95 (5/5/91)	1.17
Griess Glacier	80 a	0.09 (55/21/24)	0.44 (2/19/80)	0.22
Griess Glacier	160 a	0.37 (82/7/31)	0.34 (14/35/51)	1.07
Griess Glacier	4.9 ka	1.21 (75/9/15)	0.48 (37/17/46)	2.50
Griess Glacier	13.5 ka	2.03 (71/6/23)	0.38 (37/22/62)	5.29

Notes. Percentage contributions of root diameter classes (<1 mm, 1–2 mm, >2 mm) and functional groups (grasses, forbs, shrubs) are given in parentheses. The terrain age of the moraines is given in years (a) and thousands of years (ka).

2019). Similarly, from a belowground perspective, our findings confirm the shift from ruderal to stress-tolerant plant strategies as being derived from aboveground traits along glacier foreland succession by Caccianiga et al. (2006). In deeper soil horizons, we found an opposite SRL/RTD pattern compared to the upper layers. Here, SRL increased and RTD decreased with terrain age, which could be attributed to physical soil conditions. Typically, bulk density increases with soil depth and on young moraines, with stony soils, it is expected to reach very high levels. In these habitats, roots of early colonizers might have to face a trade-off between resource acquisition and mechanical resistance (Freschet et al. 2020).

β coefficients along the chronosequences

The vertical root distribution coefficients β showed a pattern that was in accordance with our second hypothesis. We found evidence that early successional habitats in glacier forelands have more evenly distributed root profiles in contrast to late successional communities. Our findings suggest that vegetation of young moraines is adapted to sites with limiting resources because of the potential to exploit larger volumes of soil (Gale and Grigal 1987). We further assume that late successional vegetation develops shallow root systems to exploit the resources that are concentrated in the upper soil layers as a result of biocycling and soil development (Gale and Grigal 1987; Gao et al. 2014; Ma et al. 2020). The vertical root distribution coefficients β of this study covered a wide range of values compared to a great variety of biomes, from tundra ecosystems to sclerophyllous shrublands (Jackson et al. 1996). At the second oldest moraines (3 ka, 4.9 ka) of both glacier forelands, we found exceptionally low β coefficients. Here, a lot of tufted species occurred (e.g., *Carex sempervirens*, *Festuca* sp.), forming a dense net of roots in the upper soil layers. β values of the oldest moraines (10 ka, 13.5 ka) were similar to those of Yang et al. (2009) from grasslands of the Tibetan plateau. The vegetation

of these moraines was characterized by alpine grassland species and dwarf shrubs (e.g., *Rhododendron* sp., *Vaccinium* sp.), which are known to have deep-growing roots (Kutschera and Lichtenegger 2002). These differences in vegetation composition might be a further explanation for the emerging root distribution patterns.

Biomass allocation patterns along the chronosequences

The development of biomass allocation patterns along the Griess Glacier chronosequence met our expectation that biomass allocation should follow the optimal partitioning hypothesis (Kang et al. 2013; Nie et al. 2016; Dai et al. 2019, 2020). In contrast, at the Stein Glacier site, we found support for the isometric allocation hypothesis, which has been confirmed by Yang et al. (2009), Yang and Luo (2011), and Peng and Yang (2016). The differences in the allocation patterns between the two study sites are because AGB scaled differently with terrain age: AGB remained constant at the Griess Glacier and increased at the Stein Glacier. At the latter site, the vegetation was composed of more woody species, producing higher amounts of AGB (de Kovel, Wilms, and Berendse 2000), which in turn resulted in a balanced R:S ratio. The comparatively high occurrence of species belonging to subalpine mesic dwarf shrub heathlands on siliceous bedrock has been attributed to a higher subsurface water availability compared to locations with calcareous bedrock (Michalet et al. 2002).

Percentage contributions of root diameter classes to total BGB were related to the community composition at certain moraines as reflected by the AGB fractions of grasses, forbs, and shrubs. The different proportions of coarse roots present in the soil are expected to have implications for hillslope functioning. Roots with a large diameter are known to create networks of preferential flow via root channels, thus affecting subsurface flow (Mitchell, Ellsworth, and Meek 1995; Ghoshem, Sidle, and Stokes 2011). To illustrate that, at the Stein

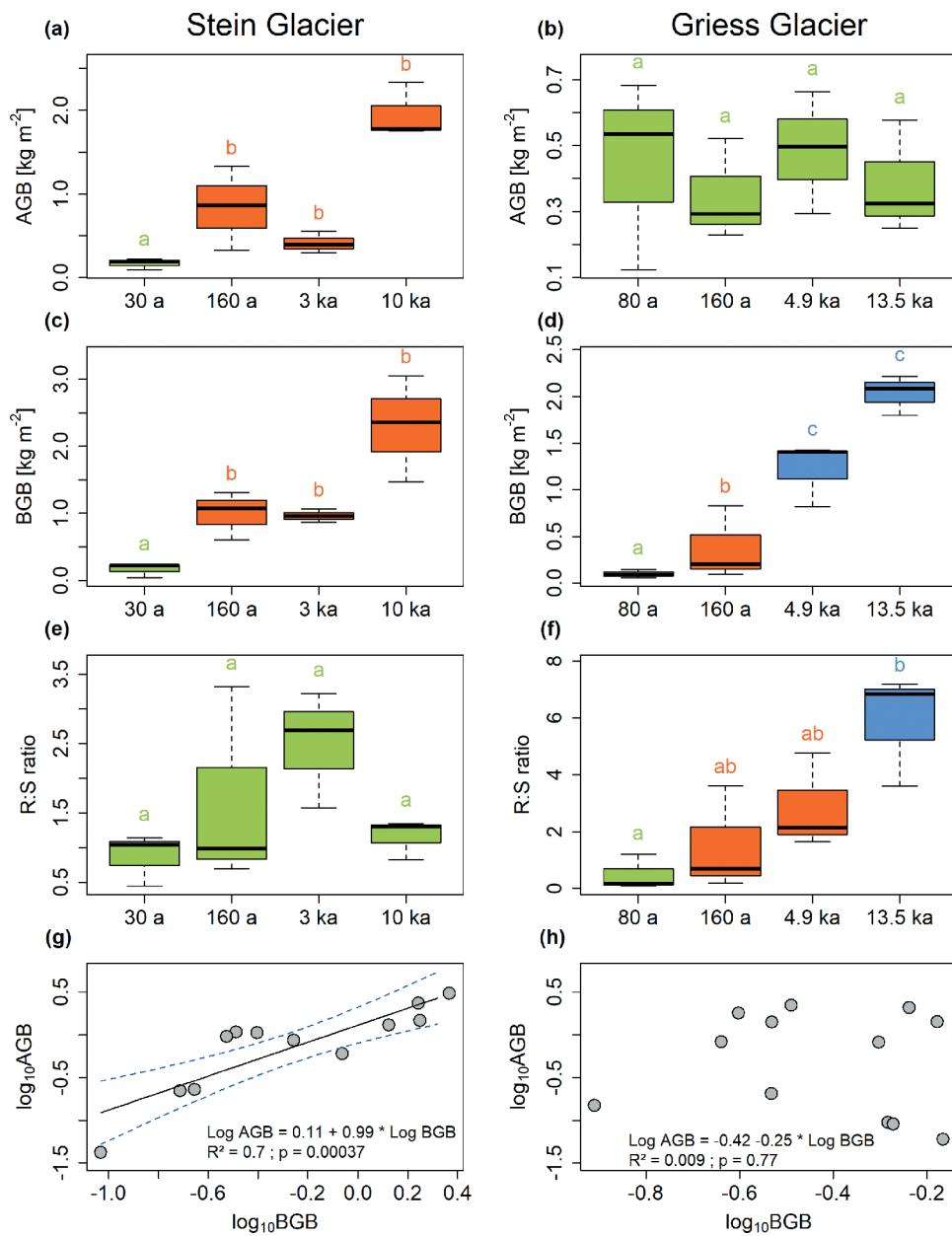


Figure 5. AGB, BGB, and R:S ratio across different moraines and the relationships between AGB and BGB of (a), (c), (e), (g) Stein Glacier and (b), (d), (f), (h) Griess Glacier forelands. A post hoc LSD test was conducted to compare biomass measures among the moraines. The significance level was adjusted according to the Bonferroni method. Superscript letters indicate significant differences.

Table 5. Pearson correlation coefficients of vertical root distribution coefficients and vegetation composition.

	β_{RD}	β_{RLD}	BGB
Species richness	-0.3	-0.35	0.23
AGB	-0.01	0.02	0.61
Grass AGB	-0.65	-0.59	0.54
Forb AGB	0.02	0.04	0.05
Shrub AGB	0.06	0.08	0.55
Vegetation cover	-0.33	-0.33	0.38
Grass cover	-0.29	-0.35	0.32
Forb cover	-0.25	-0.25	0.13
Shrub cover	-0.16	-0.13	0.57

Note. Bold numbers indicate significant correlations ($p < .05$).

Glacier site, A. Hartmann et al. (2020) found different flow types at the different moraines, ranging from matrix flow at younger terrain ages to macropore flow via root channels at the oldest moraine.

Correlations between rooting parameters and vegetation composition

Concerning the relationship between vegetation characteristics and rooting parameters, we found evidence that the occurrence of distinct functional groups drove ver-

tical root distributions, therefore supporting our fourth hypothesis. Grass species seemed especially influential to the rooting patterns due to their generally high root biomass allocation in the uppermost soil horizons (Jackson et al. 1996). Thus, we conclude that the functional group composition of the plant communities had a major influence on the development of root distribution and BGB. However, because we did not measure the distributions of available nutrients and water in the profiles, there is still some uncertainty about how abiotic factors shape the rooting patterns of such ecosystems.

Conclusions

This study is the first accurate assessment of rooting patterns including vertical root distribution along proglacial chronosequences. The presented data set is of interest for a broader understanding of functional root traits in alpine communities and provides comprehensive information on the hidden half of succession-related vegetation dynamics. We illustrated remarkable differences in rooting patterns of alpine plant communities growing along two alpine chronosequences. We found a strong variation in root traits along the successional gradients, which reflected a turnover of plant strategy types. Furthermore, we showed that plant community composition was correlated with the investigated root parameters. Our findings contribute to a deeper understanding of plant succession in glacier forelands and may also have implications for hillslope functioning in these areas. However, because we did not measure nutrient availability and water status, there remains some uncertainty as to whether plant composition or environmental conditions caused the site-specific patterns. Both nutrient and water resources in the soils are hypothesized to have a large impact on root characteristics and, therefore, we encourage the measurement of these parameters and their interrelations with root patterns in follow-up research.

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ORCID

- Konrad Greinwald  <http://orcid.org/0000-0003-1103-2529>
Lea Adina Dieckmann  <http://orcid.org/0000-0002-4844-8379>
Michael Scherer-Lorenzen  <http://orcid.org/0000-0001-9566-590X>
Tobias Gebauer  <http://orcid.org/0000-0003-2751-4071>

References

- Andreis, C., M. Caccianiga, and B. Cerabolini. 2001. Vegetation and environmental factors during primary succession on glacier forelands: Some outlines from the Italian Alps. *Plant Biosystems - an International Journal Dealing with All Aspects of Plant Biology* 135 (3):295–310. doi:[10.1080/11263500112331350930](https://doi.org/10.1080/11263500112331350930).
- Bardgett, R. D., L. Mommer, and F. T. de Vries. 2014. Going underground: Root traits as drivers of ecosystem processes. *Trends in Ecology & Evolution* 29 (12):692–99. doi:[10.1016/j.tree.2014.10.006](https://doi.org/10.1016/j.tree.2014.10.006).
- Bates, D., D. Sarkar, M. D. Bates, and L. Matrix. 2007. The lme4 package. *R Package Version* 2 (1):74.
- Bloom, A. J., F. S. Chapin, and H. A. Mooney. 1985. Resource limitation in plants—an economic analogy. *Annual Review of Ecology and Systematics* 16 (1):363–92. doi:[10.1146/annurev.es.16.110185.002051](https://doi.org/10.1146/annurev.es.16.110185.002051).
- Boxleitner, M., S. Ivy-Ochs, M. Egli, D. Brandova, M. Christl, and M. Maisch. 2019. Lateglacial and early Holocene glacier stages - New dating evidence from the Meiental in central Switzerland. *Geomorphology* 340:15–31. doi:[10.1016/j.geomorph.2019.04.004](https://doi.org/10.1016/j.geomorph.2019.04.004).
- Burga, C. A., B. Krüsi, M. Egli, M. Wernli, S. Elsener, M. Ziefle, T. Fischer, and C. Mavris. 2010. Plant succession and soil development on the foreland of the Morteratsch glacier (Pontresina, Switzerland): Straight forward or chaotic? *Flora* 205:561–76. doi:[10.1016/j.flora.2009.10.001](https://doi.org/10.1016/j.flora.2009.10.001).
- Caccianiga, M., A. Luzzaro, S. Pierce, R. M. Ceriani, and B. Cerabolini. 2006. The functional basis of a primary succession resolved by CSR classification. *Oikos* 112:10–20. doi:[10.1111/j.0030-1299.2006.14107.x](https://doi.org/10.1111/j.0030-1299.2006.14107.x).
- Caccianiga, M., and C. Andreis. 2004. Pioneer herbaceous vegetation on glacier forelands in the Italian Alps. *Phytocoenologia* 34:55–89. doi:[10.1127/0340-269X/2004/0034-0055](https://doi.org/10.1127/0340-269X/2004/0034-0055).
- Chapin, F. S., L. R. Walker, C. L. Fastie, and L. C. Sharman. 1994. Mechanisms of primary succession following deglaciation at Glacier Bay, Alaska. *Ecological Monographs* 64:149–75. doi:[10.2307/2937039](https://doi.org/10.2307/2937039).
- Dai, L., X. Guo, X. Ke, Y. Lan, F. Zhang, Y. Li, L. Lin, Q. Li, G. Cao, B. Fan, et al. 2020. Biomass allocation and productivity-richness relationship across four grassland types at the Qinghai Plateau. *Ecology and Evolution* 10 (1):506–16. doi:[10.1002/ece3.5920](https://doi.org/10.1002/ece3.5920).

- Dai, L., X. Ke, X. Guo, Y. Du, F. Zhang, Y. Li, Q. Li, L. Lin, C. Peng, K. Shu, et al. 2019. Responses of biomass allocation across two vegetation types to climate fluctuations in the northern Qinghai-Tibet Plateau. *Ecology and Evolution* 9 (10):6105–15. doi:10.1002/ece3.5194.
- de Kovel, C. F. G., Y. J. O. Wilms, and F. Berendse. 2000. Carbon and nitrogen in soil and vegetation at sites differing in successional age. *Plant Ecology* 149 (1):43–50. doi:10.1023/A:1009898622773.
- Egli, M., P. Fitze, and A. Mirabella. 2001. Weathering and evolution of soils formed on granitic, glacial deposits: Results from chronosequences of Swiss alpine environments. *Catena* 45 (1):19–47. doi:10.1016/S0341-8162(01)00138-2.
- Enquist, B. J., and K. J. Niklas. 2002. Global allocation rules for patterns of biomass partitioning in seed plants. *Science* 295 (5559):1517. doi:10.1126/science.1066360.
- Erktan, A., M. L. McCormack, and C. Roumet. 2018. Frontiers in root ecology: Recent advances and future challenges. *Plant and Soil* 424 (1):1–9. doi:10.1007/s11104-018-3618-5.
- Erschbamer, B., and M. S. Caccianiga. 2017. Glacier forelands: Lessons of plant population and community development. In *Progress in Botany*, ed. F. M. Cánovas, U. Lüttge, and R. Matyssek, vol. 78, 259–84. Cham: Springer International Publishing.
- Franzén, M., P. Dieker, J. Schrader, and A. Helm. 2019. Rapid plant colonization of the forelands of a vanishing glacier is strongly associated with species traits. *Arctic, Antarctic, and Alpine Research* 51 (1):366–78. doi:10.1080/15230430.2019.1646574.
- Freschet, G. T., C. Roumet, L. Comas, M. Weemstra, A. Bengough, B. Rewald, R. Bardgett, G. B. Deyn, D. Johnson, J. Klimešová, M. Lukac, M. McCormack, I. Meier, L. Pagès, H. Poorter, I. Prieto, N. Wurzburger, M. Zadworny, A. Bagniewska-Zadworna, and F. Schnabel. 2020. Root traits as drivers of plant and ecosystem functioning: Current understanding, pitfalls and future research needs. *New Phytologist* accepted. doi:10.1111/nph.17072.
- Freschet, G. T., C. Violle, M. Y. Bourget, M. Scherer-Lorenzen, and F. Fort. 2018. Allocation, morphology, physiology, architecture: The multiple facets of plant above- and below-ground responses to resource stress. *New Phytologist* 219 (4):1338–52. doi:10.1111/nph.15225.
- Freschet, G. T., E. M. Swart, and J. H. C. Cornelissen. 2015. Integrated plant phenotypic responses to contrasting above- and below-ground resources: Key roles of specific leaf area and root mass fraction. *New Phytologist* 206 (4):1247–60. doi:10.1111/nph.13352.
- Freschet, G. T., J. H. C. Cornelissen, R. S. P. Van Logtestijn, and R. Aerts. 2010. Evidence of the ‘plant economics spectrum’ in a subarctic flora. *The Journal of Ecology* 98 (2):362–73. doi:10.1111/j.1365-2745.2009.01615.x.
- Freschet, G. T., O. J. Valverde-Barrantes, C. M. Tucker, J. M. Craine, M. L. McCormack, C. Violle, F. Fort, C. B. Blackwood, K. R. Urban-Mead, C. M. Iversen, A. Bonis, L. H. Comas, J. H. C. Cornelissen, M. Dong, D. Guo, S. E. Hobbie, R. J. Holdaway, S. W. Kembel, N. Makita, V. G. Onipchenko, C. Picon-Cochard, P. B. Reich, E. G. de la Riva, S. W. Smith, N. A. Soudzilovskaia, M. G. Tjoelker, D. A. Wardle, and C. Roumet. 2017. Climate, soil and plant functional types as drivers of global fine-root trait variation. *Journal of Ecology* 105 (5):1182–96. doi:10.1111/1365-2745.12769.
- Fukami, T., T. Martijn Bezemer, S. R. Mortimer, and W. H. van der Putten. 2005. Species divergence and trait convergence in experimental plant community assembly. *Ecology Letters* 8 (12):1283–90. doi:10.1111/j.1461-0248.2005.00829.x.
- Gale, M., and D. F. Grigal. 1987. Vertical root distributions of northern tree species in relation to successional status. *Canadian Journal of Forest Research* 17 (8):829–34. doi:10.1139/x87-131.
- Gao, Y., N. He, G. Yu, W. Chen, and Q. Wang. 2014. Long-term effects of different land use types on C, N, and P stoichiometry and storage in subtropical ecosystems: A case study in China. *Ecological Engineering* 67:171–81. doi:10.1016/j.ecoleng.2014.03.013.
- Garnier, E., M.-L. Navas, and K. Grigulis. 2016. *Plant functional diversity. Organism traits, community structure, and ecosystem properties*. 1st ed. Oxford, UK: Oxford University Press.
- Ghestem, M., R. C. Sidle, and A. Stokes. 2011. The influence of plant root systems on subsurface flow: Implications for slope stability. *BioScience* 61 (11):869–79. doi:10.1525/bio.2011.61.11.6.
- Gobbi, M., M. Caccianiga, B. Cerabolini, F. de Bernardi, A. Luzzaro, and S. Pierce. 2010. Plant adaptive responses during primary succession are associated with functional adaptations in ground beetles on deglaciated terrain. *Community Ecology* 11 (2):223–31. doi:10.1556/ComEc.11.2010.2.11.
- Hartmann, A., E. Semenova, M. Weiler, and T. Blume. 2020. Field observations of soil hydrological flow path evolution over 10 millennia. *Hydrology and Earth System Sciences* 24 (6):3271–88. doi:10.5194/hess-24-3271-2020.
- Hartmann, P., and K. von Wilpert. 2013. Fine-root distributions of Central European forest soils and their interaction with site and soil properties. *Canadian Journal of Forest Research* 44 (1):71–81. doi:10.1139/cjfr-2013-0357.
- Hendricks, J. J., K. J. Nadelhoffer, and J. D. Aber. 1993. Assessing the role of fine roots in carbon and nutrient cycling. *Trends in Ecology & Evolution* 8 (5):174–78. doi:10.1016/0169-5347(93)90143-D.
- Ho, M. D., J. C. Rosas, K. M. Brown, and J. P. Lynch. 2005. Root architectural tradeoffs for water and phosphorus acquisition. *Functional Plant Biology* 32 (8):737–48. doi:10.1071/FP05043.
- Hodgson, J. G., P. J. Wilson, R. Hunt, J. P. Grime, and K. Thompson. 1999. Allocating C-S-R plant functional types: A soft approach to a hard problem. *Oikos* 85 (2):282–94. doi:10.2307/3546494.
- Holdaway, R. J., S. J. Richardson, I. A. Dickie, D. A. Peltzer, and D. A. Coomes. 2011. Species-and community-level patterns in fine root traits along a 120 000-year soil chronosequence in temperate rain forest. *Journal of Ecology* 99 (4):954–63. doi:10.1111/j.1365-2745.2011.01821.x.
- Jackson, R. B., J. Canadell, J. R. Ehleringer, H. A. Mooney, O. E. Sala, and E. D. Schulze. 1996. A global analysis of root distributions for terrestrial biomes. *Oecologia* 108 (3):389–411. doi:10.1007/BF00333714.
- Kang, M., C. Dai, W. Ji, Y. Jiang, Z. Yuan, and H. Y. H. Chen. 2013. Biomass and its allocation in relation to temperature, precipitation, and soil nutrients in inner Mongolia Grasslands, China. *PLoS One* 8 (7):e69561. doi:10.1371/journal.pone.0069561.
- Kong, D., J. Wang, H. Wu, O. J. Valverde-Barrantes, R. Wang, H. Zeng, et al. 2019. Nonlinearity of root trait relationships

- and the root economics spectrum. *Nature Communications* 10 (1):2203. doi:10.1038/s41467-019-10245-6.
- Kutschera, L., and E. Lichtenegger. 2002. *Wurzelatlas mitteleuropäischer Waldbäume und Sträucher*. Leopold Stocker Verlag Graz.
- Li, F., H. Hu, M. L. McCormack, F. De Feng, X. Liu, and W. Bao. 2019. Community-level economics spectrum of fine-roots driven by nutrient limitations in subalpine forests. *Journal of Ecology* 107 (3):1238–49. doi:10.1111/1365-2745.13125.
- Li, Y., T. Luo, and Q. Lu. 2008. Plant height as a simple predictor of the root to shoot ratio: Evidence from alpine grasslands on the Tibetan Plateau. *Journal of Vegetation Science* 19 (2):245–52. doi:10.3170/2007-8-18365.
- Lüdi, W. 1955. Die Vegetationsentwicklung seit dem Rückzug der Gletscher in den mittleren Alpen und ihrem nördlichen Vorland. *Ber. Geobot. Forsch. Inst. Rübel Zürich* 36–68.
- Lüdi, W. 1958. Beobachtungen über die Besiedlung von Gletschervorfeldern in den Schweizeralpen. *Flora* 146:386–407. doi:10.1016/S0367-1615(17)32526-0.
- Ma, R., F. Hu, J. Liu, C. Wang, Z. Wang, G. Liu, and S. Zhao. 2020. Shifts in soil nutrient concentrations and C: N: P stoichiometry during long-term natural vegetation restoration. *PeerJ* 8:e8382. doi:10.7717/peerj.8382.
- Maier, F., I. van Meerveld, K. Greinwald, T. Gebauer, F. Lustenberger, A. Hartmann, and A. Musso. 2020. Effects of soil and vegetation development on surface hydrological properties of moraines in the Swiss Alps. *Catena* 187:1–17. doi:10.1016/j.catena.2019.104353.
- Matthews, J. A. 1992. *The ecology of recently-deglaciated terrain. A geoecological approach to glacier forelands*. Cambridge, UK: Cambridge University Press (Cambridge studies in ecology).
- Matthews, J. A., and R. J. Whittaker. 1987. Vegetation succession on the Storbreen glacier foreland, Jotunheimen, Norway: A review. *Arctic and Alpine Research* 19:385–95. doi:10.2307/1551403.
- McCarthy, M. C., and B. J. Enquist. 2007. Consistency between an allometric approach and optimal partitioning theory in global patterns of plant biomass allocation. *Functional Ecology* 21 (4):713–20.
- McConaughay, K. D. M., and J. S. Coleman. 1999. Biomass allocation in plants: Ontogeny or optimality? A test along three resource gradients. *Ecology* 80 (8):2581–93.
- Michalet, R., C. Gandoy, D. Joud, J.-P. Pagès, and P. Choler. 2002. Plant community composition and biomass on calcareous and siliceous substrates in the northern French Alps: Comparative effects of soil chemistry and water status. *Arctic, Antarctic, and Alpine Research* 34 (1):102–13. doi:10.1080/15230430.2002.12003474.
- Mitchell, A. R., T. R. Ellsworth, and B. D. Meek. 1995. Effect of root systems on preferential flow in swelling soil. *Communications in Soil Science and Plant Analysis* 26 (15–16):2655–66. doi:10.1080/00103629509369475.
- Mommer, L., J. van Ruijven, H. de Caluwe, A. E. Smit-Tiekstra, C. A. M. Wagemaier, N. Joop Ouborg, G. M. Bögemann, G. M. Van Der Weerden, F. Berendse, and H. Kroon. 2010. Unveiling below-ground species abundance in a biodiversity experiment: A test of vertical niche differentiation among grassland species. *Journal of Ecology* 98 (5):1117–27. doi:10.1111/j.1365-2745.2010.01702.x.
- Müller, I., B. Schmid, and J. Weiner. 2000. The effect of nutrient availability on biomass allocation patterns in 27 species of herbaceous plants. *Perspectives in Plant Ecology, Evolution and Systematics* 3 (2):115–27. doi:10.1078/1433-8319-00007.
- Musso, A., K. Lamorski, C. Ślawiński, C. Geitner, A. Hunt, K. Greinwald, and M. Egli. 2019. Evolution of soil pores and their characteristics in a siliceous and calcareous proglacial area. *Catena* 182:1–16. doi:10.1016/j.catena.2019.104154.
- Musso, A., M. Ketterer, K. Greinwald, C. Geitner, and M. Egli. 2020. Rapid decrease of soil erosion rates with soil formation and vegetation development in periglacial areas. *Earth Surface Processes and Landforms* (45):2824–39. doi:10.1002/esp.4932.
- Nie, X., Y. Yang, L. Yang, and G. Zhou. 2016. Above- and belowground biomass allocation in shrub biomes across the Northeast Tibetan Plateau. *PLoS One* 11 (4). doi: 10.1371/journal.pone.0154251.
- Niklas, K. J. 2006. A phyletic perspective on the allometry of plant biomass-partitioning patterns and functionally equivalent organ-categories. *New Phytologist* 171 (1):27–40. doi:10.1111/j.1469-8137.2006.01760.x.
- Oechslin, M. 1935. Beitrag zur Kenntnis der pflanzlichen Besiedelung der durch Gletscher freigegebenen Grundmoränenböden. *Ber Naturforsch Ges Uri* 4:27–48.
- Peng, Y., and Y. Yang. 2016. Allometric biomass partitioning under nitrogen enrichment: Evidence from manipulative experiments around the world. *Scientific Reports* 6 (1):28918. doi:10.1038/srep28918.
- Pohl, M., R. Stroude, A. Buttler, and C. Rixen. 2011. Functional traits and root morphology of alpine plants. *Annals of Botany* 108 (3):537–45. doi:10.1093/aob/mcr169.
- Poorter, H., K. J. Niklas, P. B. Reich, J. Oleksyn, P. Poot, and L. Mommer. 2012. Biomass allocation to leaves, stems and roots: Meta-analyses of interspecific variation and environmental control. *New Phytologist* 193 (1):30–50. doi:10.1111/j.1469-8137.2011.03952.x.
- Prach, K., and L. R. Walker. 2020. *Comparative plant succession among terrestrial biomes of the World*. Cambridge: Cambridge University Press (Ecology, Biodiversity and Conservation).
- Prach, K., P. Pyšek, and P. Šmilauer. 1997. Changes in species traits during succession: A search for pattern. *Oikos* 79 (1):201–05. doi:10.2307/3546109.
- R Core Team. 2018. R: A language and environment for statistical computing. R foundation for statistical computing. Vienna, Austria. Accessed April 2, 2019. <https://www.r-project.org/>.
- Raevel, V., C. Violette, and F. Munoz. 2012. Mechanisms of ecological succession: Insights from plant functional strategies. *Oikos* 121:1761–70. doi:10.1111/j.1600-0706.2012.20261.x.
- Raffl, C., and B. Erschbamer. 2004. Comparative vegetation analyses of two transects crossing a characteristic glacier valley in the Central Alps. *Phytocoenologia* 34:225–40. doi:10.1127/0340-269X/2004/0034-0225.
- Raffl, C., M. Mallau, R. Mayer, and B. Erschbamer. 2006. Vegetation succession pattern and diversity changes in a Glacier Valley, Central Alps, Austria. *Arctic, Antarctic, and Alpine Research* 38:421–28.
- Régent Instruments Inc. 2013. *WinRHIZO 2013 Reg, user manual*. Quebec City, QC: Régent Instruments Inc.
- Reich, P. B. 2014. The world-wide ‘fast–slow’ plant economics spectrum: A traits manifesto. *Journal of Ecology* 102 (2):275–301. doi:10.1111/1365-2745.12211.

- Robbins, J. A., and J. A. Matthews. 2009. Pioneer vegetation on glacier forelands in southern Norway: Emerging communities? *Journal of Vegetation Science* 20:889–902. doi:[10.1111/j.1654-1103.2009.01090.x](https://doi.org/10.1111/j.1654-1103.2009.01090.x).
- Robbins, J. A., and J. A. Matthews. 2010. Regional variation in successional trajectories and rates of vegetation change on glacier forelands in south-central Norway. *Arctic, Antarctic, and Alpine Research* 42 (3):351–61. doi:[10.1657/1938-4246-42.3.351](https://doi.org/10.1657/1938-4246-42.3.351).
- Ryser, P. 1996. The importance of tissue density for growth and life span of leaves and roots: A comparison of five ecologically contrasting grasses. *Functional Ecology* 10 (6):717–23. doi:[10.2307/2390506](https://doi.org/10.2307/2390506).
- Ryser, P. 2006. The mysterious root length. *Plant and Soil* 286 (1):1–6. doi:[10.1007/s11104-006-9096-1](https://doi.org/10.1007/s11104-006-9096-1).
- Ryser, P., and L. Eek. 2000. Consequences of phenotypic plasticity vs. interspecific differences in leaf and root traits for acquisition of aboveground and belowground resources. *American Journal of Botany* 87 (3):402–11. doi:[10.2307/2656636](https://doi.org/10.2307/2656636).
- Scheiner, S. M. 1993. MANOVA: Multiple response variables and multispecies interactions. *Design and Analysis of Ecological Experiments* 94:112.
- Schenk, H. J., and R. B. Jackson. 2002. The global biogeography of roots. *Ecological Monographs* 72 (3):311–28. doi:[10.1890/0012-9615\(2002\)072\[0311:TGBOR\]2.0.CO;2](https://doi.org/10.1890/0012-9615(2002)072[0311:TGBOR]2.0.CO;2).
- Schimmelpfennig, I., J. M. Schaefer, N. Akçar, T. Koffman, S. Ivy-Ochs, R. Schwartz, R. Finkel, S. Zimmerman, and C. Schlüchter. 2014. A chronology of holocene and little ice age glacier culminations of the Steingletscher, Central Alps, Switzerland, based on high-sensitivity beryllium-10 moraine dating. *Earth and Planetary Science Letters* 393:220–30. doi:[10.1016/j.epsl.2014.02.046](https://doi.org/10.1016/j.epsl.2014.02.046).
- Schleicher, A., C. Peppler-Lisbach, and M. Kleyer. 2011. Functional traits during succession: Is plant community assembly trait-driven? *Preslia* 83 (3):347–70.
- Schumann, K., S. Gewolf, and O. Tackenberg. 2016. Factors affecting primary succession of glacier foreland vegetation in the European Alps. *Alpine Botany* 126 (2):105–17. doi:[10.1007/s00035-016-0166-6](https://doi.org/10.1007/s00035-016-0166-6).
- Shipley, B., and D. Meziane. 2002. The balanced-growth hypothesis and the allometry of leaf and root biomass allocation. *Functional Ecology* 16 (3):326–31. doi:[10.1046/j.1365-2435.2002.00626.x](https://doi.org/10.1046/j.1365-2435.2002.00626.x).
- Unger, P. W., and T. C. Kaspar. 1994. Soil compaction and root growth: A review. *Agronomy Journal* 86 (5):759–66. doi:[10.2134/agronj1994.00021962008600050004x](https://doi.org/10.2134/agronj1994.00021962008600050004x).
- van der Putten, W. H., R. D. Bardgett, J. D. Bever, T. M. Bezemer, B. B. Casper, T. Fukami, P. Kardol, J. N. Klironomos, A. Kulmatiski, J. A. Schweitzer, et al. 2013. Plant–soil feedbacks: The past, the present and future challenges. *Journal of Ecology* 101 (2):265–76. doi:[10.1111/1365-2745.12054](https://doi.org/10.1111/1365-2745.12054).
- Walker, L. R., and R. Del Moral. 2003. *Primary succession and ecosystem rehabilitation*. Cambridge: Cambridge University Press.
- Weppeler, T., and J. Stöcklin. 2005. Variation of sexual and clonal reproduction in the alpine *Geum reptans* in contrasting altitudes and successional stages. *Basic and Applied Ecology* 6 (4):305–16. doi:[10.1016/j.baee.2005.03.002](https://doi.org/10.1016/j.baee.2005.03.002).
- Wright, I. J., P. B. Reich, M. Westoby, D. D. Ackerly, Z. Baruch, F. Bongers, J. Cavender-Bares, T. Chapin, J. H. C. Cornelissen, M. Diemer, et al. 2004. The worldwide leaf economics spectrum. *Nature* 428:821. doi:[10.1038/nature02403](https://doi.org/10.1038/nature02403).
- Yang, Y., J. Fang, C. Ji, and W. Han. 2009. Above- and below-ground biomass allocation in Tibetan grasslands. *Journal of Vegetation Science* 20 (1):177–84. doi:[10.1111/j.1654-1103.2009.05566.x](https://doi.org/10.1111/j.1654-1103.2009.05566.x).
- Yang, Y., and Y. Luo. 2011. Isometric biomass partitioning pattern in forest ecosystems: Evidence from temporal observations during stand development. *The Journal of Ecology* 99 (2):431–37. doi:[10.1111/j.1365-2745.2010.01774.x](https://doi.org/10.1111/j.1365-2745.2010.01774.x).
- Zeng, C., J. Wu, and X. Zhang. 2015. Effects of grazing on above- vs. below-ground biomass allocation of alpine grasslands on the Northern Tibetan Plateau. *PLoS One* 10 (8):e0135173. doi:[10.1371/journal.pone.0135173](https://doi.org/10.1371/journal.pone.0135173).